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**The accuracy and reliability of traditional surface flow type mapping: Is it time for a new method of characterising physical river habitat?**

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## The accuracy and reliability of traditional surface flow type mapping: Is it time for a new method of characterising physical river habitat?

Woodget, A.S., Visser, F., Maddock, I.P. and Carbonneau, P.E.

### Abstract

Surface flow types (SFT) are advocated as ecologically relevant hydraulic units, often mapped visually from the bankside to characterise rapidly the physical habitat of rivers. SFT mapping is simple, non-invasive and cost-efficient. However, it is also qualitative, subjective and plagued by difficulties in recording accurately the spatial extent of SFT units. Quantitative validation of the underlying physical habitat parameters is often lacking, and does not consistently differentiate between SFTs. Here, we investigate explicitly the accuracy, reliability and statistical separability of traditionally mapped SFTs as indicators of physical habitat, using independent, hydraulic and topographic data collected during three surveys of a c. 50m reach of the River Arrow, Warwickshire, England. We also explore the potential of a novel remote sensing approach, comprising a small unmanned aerial system (sUAS) and Structure-from-Motion photogrammetry (SfM), as an alternative method of physical habitat characterisation. Our key findings indicate that SFT mapping accuracy is highly variable, with overall mapping accuracy not exceeding 74%. Results from analysis of similarity (ANOSIM) tests found that strong differences did not exist between all SFT pairs. This leads us to question the suitability of SFTs for characterising physical habitat for river science and management applications. In contrast, the sUAS-SfM approach provided high resolution, spatially continuous, spatially explicit, quantitative measurements of water depth and point cloud roughness at the microscale (spatial scales  $\leq 1\text{m}$ ). Such data are acquired rapidly, inexpensively, and provide new opportunities for examining the heterogeneity of physical habitat over a range of spatial and temporal scales. Whilst continued refinement of the sUAS-SfM approach is required, we propose that this method offers an opportunity to move away from broad, mesoscale classifications of physical habitat (spatial scales 10-100m), and towards continuous, quantitative measurements of the continuum of hydraulic and geomorphic conditions which actually exists at the microscale.

### Introduction

Surface flow types (SFTs) are visible water surface patterns within river systems often used as a proxy for biotopes or in-stream hydraulic habitat units (Padmore, 1998; Newson and Newson, 2000). The spatial distribution of SFTs is thought to be determined by local variations in morpho-hydraulic conditions, including topography, substrate size, water depth and flow velocity (Padmore, 1998; Wadeson and Rowntree, 1998; Newson and Newson, 2000; Dyer and Thoms, 2006; Reid and Thoms, 2009). Various classification schemes have been proposed to define SFTs or their equivalent biotopes (e.g. Bisson et al., 1982; Wadeson, 1994; Montgomery and Buffington, 1997; Raven et al., 1997; Newson and Newson, 2000) and their visual identification has been advocated as an effective means of characterising the physical river habitat template at the mesoscale (Reid and Thoms, 2008; Harvey and Clifford, 2009; Zavadil et al., 2012).

The UK's River Habitat Survey (RHS) (Environment Agency, 2003) is used to characterise river habitat over 500m reaches, including the classification of nine different SFTs (Table 1) for assessing physical habitat diversity. RHS results feed into river condition assessments required by the European Union's Water Framework Directive (European Commission, 2000) and allow repeat monitoring of habitat conditions. The view of SFTs as hydraulically meaningful and consistent units, and the strong influence of hydraulic conditions on instream biota (Reid and Thoms, 2008; Hill et al., 2013), means that there has been a growing

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2  
3 popularity within the field of ecohydraulics for conducting rapid habitat assessments using  
4 SFTs (Newson and Newson, 2000; Dyer and Thoms, 2006; Reid and Thoms 2008). As well  
5 as characterising hydraulic diversity, SFTs are also thought to be indicative of geomorphic  
6 diversity (Zavadil et al., 2012).  
7

8  
9 Traditionally, SFT mapping is carried out by visual assessment from the bankside. The RHS  
10 records the dominant SFT across one metre wide cross sections at ten spot check locations  
11 every 50m along a 500m reach (Environment Agency, 2003). Other approaches have  
12 recorded the relative proportions of SFTs within 50m cells (Dyer and Thoms, 2006),  
13 estimated the spatial extent of SFTs (Hill et al., 2013), and recorded SFTs at specific  
14 distances along channel cross sections (Wadeson and Rowntree, 1998). Such approaches are  
15 rapid, non-invasive and advocated to provide a valid alternative to the labour intensive, time  
16 consuming point measurements of bed elevations, water depths and velocities for  
17 characterising hydraulic and geomorphic diversity (Reid and Thoms, 2008). However, whilst  
18 cost effective and easy to conduct, these visual SFT mapping approaches are qualitative and  
19 are usually unaccompanied by quantitative validation. They are subjective, affected by user  
20 bias and plagued by difficulties in determining the precise spatial extent of different SFT  
21 units due to the indistinct nature of SFT boundaries and the difficulty of producing accurate  
22 maps from an oblique, bankside viewpoint (Reid and Thoms, 2009; Milan et al., 2010).  
23 Efforts to characterise the hydraulic and geomorphic nature of SFTs (using measures such as  
24 flow velocity, Froude number or cross-sectional geometry) find that SFTs are not always  
25 statistically separable, and that significant overlaps in conditions between SFTs are often  
26 present (Wadeson, 1994; Newson et al., 1998; Wadeson and Rowntree, 1998; Clifford et al.,  
27 2006; Moir and Pasternack, 2008; Reid and Thoms, 2008; Gosselin et al., 2012; Zavadil et  
28 al., 2012). As a result, can we be certain of the accuracy and reliability of traditional SFT  
29 mapping as a universal approach for characterising physical habitat within rivers?  
30  
31

32  
33 Remote sensing surveys offer an alternative means of SFT mapping (Newson and Newson,  
34 2000), which offer the opportunity for permanent, digital, spatially continuous and spatially  
35 explicit datasets, against which change over time can be assessed and other types of spatial  
36 analysis performed. Historically, the expense of acquiring remotely sensed data seems to  
37 have prohibited its widespread use for SFT mapping. However, some research has  
38 demonstrated the use of multispectral imagery (Hardy et al., 1994; Wright et al., 2000; Reid  
39 and Thoms, 2009), hyperspectral imagery (Marcus, 2002; Marcus et al., 2003) and measures  
40 of roughness derived from terrestrial laser scanning (TLS) point clouds for mapping SFTs  
41 over a range of scales (Large and Heritage, 2007; Milan et al., 2010). Most of these studies  
42 report good agreement between SFTs mapped from the remote sensing and using the  
43 traditional bankside approach, and in some cases it is argued that the remote mapping might  
44 actually provide a more accurate and precise representation of the spatial arrangement of  
45 SFTs (Hardy et al., 1994; Reid and Thoms, 2009; Marcus, 2002). Unfortunately, quantitative  
46 and independent validation data are not usually available to assess such findings and  
47 therefore we remain largely uncertain of the accuracy and reliability of traditional bankside  
48 SFT mapping or that based on remotely sensed data.  
49  
50

51  
52 Our aims within this paper are twofold:

- 53 (1) To use a range of independent, spatially explicit, microscale quantitative  
54 measurements to investigate the accuracy and reliability of traditionally mapped SFTs  
55 as indicators of physical habitat conditions.  
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3 (2) To explore the potential of a novel remote sensing technique for providing an  
4 alternative approach to SFT mapping for characterising physical habitat.  
5

6 To achieve our first aim, we acquire quantitative validation data; (i) in the field, using  
7 established methods and; (ii) from remotely sensed imagery acquired using a small unmanned  
8 aerial system (sUAS), otherwise known as a drone, RPAS (remotely piloted aircraft system)  
9 or UAV (unmanned aerial vehicle). In recent years sUAS, combined with Structure-from-  
10 Motion photogrammetry (SfM), have provided a new remote sensing tool capable of rapid  
11 and flexible acquisition of hyperspatial resolution imagery (<10cm) and the production of  
12 high accuracy, quantitative, topographic data (e.g. Harwin and Lucieer, 2012). This sUAS-  
13 SfM approach has seen a huge surge in popularity for a wide range of environmental science  
14 and monitoring applications, especially for smaller sites, which would otherwise be  
15 impossible or too costly to survey using established remote sensing approaches.  
16 Quantification of some river habitat parameters has already been demonstrated (e.g. Reid and  
17 Thoms, 2009; Fonstad et al., 2013; Tamminga et al., 2015; Woodget et al., 2015), and we  
18 anticipate that the sUAS-SfM toolkit now offers new possibilities for high resolution,  
19 spatially explicit and quantitative assessments of the accuracy and reliability of traditional  
20 SFT mapping. For our second aim, we take this assessment a step further to explore whether  
21 the sUAS-SfM approach itself offers an alternative means of characterising physical habitat.  
22 Our objectives are:  
23

- 24  
25 1. To compare quantitatively the traditional SFT mapping against;  
26 a. An independent SFT ground truth dataset collected in the field, and;  
27 b. SFT mapping conducted from hyperspatial resolution sUAS-SfM imagery  
28 (which itself is then compared to the ground truth data).  
29  
30 2. To test statistically whether SFTs mapped traditionally from the bankside can be  
31 differentiated reliably based on quantitative data derived from;  
32 a. Water depth and flow velocity data acquired in the field, and;  
33 b. Hyperspatial resolution sUAS-SfM imagery and topography data.  
34  
35 3. To quantify physical habitat parameters using sUAS-SfM derived topography data.  
36

### 37 Site Location

38 We conducted this research on a 50m reach of the River Arrow, a small (5-12m wide),  
39 lowland river (catchment area 93.72 km<sup>2</sup>) located near Studley, Warwickshire, UK (Figure  
40 1). The site provides a diverse range of SFTs at the mesoscale, has a wadeable channel  
41 (<0.7m deep), is safe for sUAS flying and was easily accessible for repeat surveying. We  
42 collected three repeat surveys of the site to allow comparison of results obtained under  
43 differing conditions.  
44

45 The reach of interest forms a meandering, pool-riffle system, with steep banks and active  
46 erosion on the outer meander bends. Point bars are present, with gravels and cobbles also  
47 accumulating around the margins of vegetated islands. There are some submerged  
48 macrophytes, and in places the channel is obscured by overhanging vegetation. The channel  
49 bed is composed predominantly of cobbles with some patches of gravel, sands and silt  
50 substrates. Flow level is highly responsive to rainfall events. National River Flow Archive  
51 data for the time of our surveys reports a discharge range between 0.94 (August) and 1.72  
52 cumecs (May) from the nearest gauging station, which is located further downstream at  
53 Broom (catchment area 319 km<sup>2</sup>) (CEH, 2015). These discharge values equate to flow  
54 exceedance percentages of Q85 to Q48 respectively, based on long term (1957-2013) flow  
55 data acquired at this gauging station (CEH, 2015).  
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## Methods

### *Traditional SFT Mapping*

During each of our three field campaigns (May, June and August 2013), we mapped the spatial extents of SFTs from the bankside using the RHS classification (Table 1) and hardcopy base maps at a scale of 1:150 (May) or 1:200 (June/August). We subsequently scanned and georeferenced our annotated maps, then digitised the mapped SFT units using ArcGIS v. 10 (ESRI Inc.).

### *Independent Ground Truthing*

During each campaign, we also conducted an independent or 'ground truth' SFT survey in the field by assessing SFTs at 60-100 point locations, the positions of which were recorded using a Trimble R8 differential global positioning system (dGPS) or Leica Builder 500 total station (TS) by a team of two people. We viewed the SFTs from different angles, rather than solely from a bankside or in-channel location, and at each point measured water depth ( $\pm 1\text{cm}$ ) and mean column velocity (m/s) at 0.6 depth using a Valeport EM801 electromagnetic flow meter.

### *sUAS-SfM Survey and Image Processing*

The sUAS imagery we use here is presented in Woodget et al., (2015). A summary of the sUAS-SfM method is provided here, and readers are referred to the aforementioned paper for greater detail. Prior to our sUAS surveys, we characterised the camera geometry to establish the flying height necessary to obtain c. 1cm resolution imagery using the Panasonic Lumix DMC-LX3 (10.1MP) camera attached to our Draganflyer X6 rotary-winged sUAS (Figure 2). We also distributed c. 20 ground control points (GCPs) throughout the site prior to flying and surveyed their positions using the dGPS or TS. We collected imagery from an altitude of 25-30m and with a high level of overlap (typically 60-80%) to ensure successful SfM processing.

Following the field campaigns, we assessed sUAS image quality to remove those affected by blur. We processed the remaining images using a SfM workflow within Agisoft's PhotoScan Pro software (v. 0.9.1.1714). Outputs included a hyperspatial resolution orthophoto mosaic, a digital elevation model (DEM) and a dense point cloud. The latter comprises a cloud of points arranged in 3D space, each of which has XYZ co-ordinates describing its location within a 3D model of the scene. Point clouds created by the SfM process are similar to those produced by laser scanning and permit different ways of examining topographic surfaces (e.g. Rychov et al., 2012).

We used the sUAS-SfM orthophoto mosaics to map SFTs as polygons within ArcGIS at a scale of 1:50. We conducted this manually (rather than using image classification procedures) based on the visual properties of the water surface. We derived quantitative data from the sUAS-SfM process in the form of refraction corrected (RC) water depth and point cloud roughness, both as raster datasets. Woodget et al., (2015) provides further detail concerning water depth estimation and the required refraction correction procedure, as initially proposed by Westaway et al., (2000). We used the freeware package CloudCompare (Girardeau-Montaut, 2014) to compute the roughness (i.e. fine scale variation in elevation) of the point cloud. In theory, greater roughness in the point cloud may result from greater water surface roughness, greater water depths and larger submerged grain sizes. Therefore, variations in point cloud roughness may be indicative of variations in physical habitat conditions. Roughness is defined as the distance between each point and the least squares best fitting plane computed on its nearest neighbours within a sphere of a given size. We chose a sphere

radius of 0.2m based on a priori knowledge that the typical size of SFT features at this site does not exceed 0.4m. Per-point roughness data were rasterised at 5cm resolution.

### Analysis

*Objective 1.a)* To compare quantitatively the spatial positioning of the traditional SFT mapping against the ground truth data, we extracted the bankside SFT classifications from each dataset at the location of each ground truth point and compiled the data into a confusion matrix. These matrices are often employed for assessing the accuracy of automated or manual image classifications (Lillesand and Kiefer, 2000) by comparing their performance against datasets of true or known values. We computed the overall accuracy, user's accuracy and kappa co-efficient from the confusion matrix to assess the performance of the traditional SFT mapping. Definitions of these accuracy measures are provided in Table 2.

*Objective 1.b)* We also compared quantitatively the traditional SFT mapping against the SFT mapping conducted on the high resolution sUAS-SfM orthophoto. SFT classes were extracted from both mapping datasets at 0.5m spaced grid points in ArcGIS. Here, we seek not to use the sUAS-SfM SFT mapping as ground truth, but instead we compiled the resulting data to compare levels of agreement between these two mapping approaches. For completeness, we also compared quantitatively the spatial positioning of the sUAS-SfM SFT mapping against the ground truth data, in the same way as described for the traditional SFT mapping in objective 1a.

*Objective 2)* To test whether the traditionally mapped SFTs were statistically separable from each other, analysis of similarity (ANOSIM) tests were performed. We conducted two ANOSIM tests for each survey, using different combinations of spatially explicit, quantitative input data that provide an indication of topographic and hydraulic diversity;

- Test A: Water depth and flow velocity data acquired in the field.
- Test B: Hyperspatial resolution image and topography data in the form of RC water depth and point cloud roughness, obtained readily from the sUAS-SfM data.

We assessed SFT separability for each survey and scenario using PRIMER 6 (PRIMER-E Ltd, v. 6.1.13) and  $\alpha = 0.05$ . Data were checked for normality and standardised by subtracting the mean and dividing by the standard deviation. A resemblance matrix was generated for each test, which details the differences between all data points in Euclidean distances. The SFT classifications for each point were added and the resemblance matrix used to conduct a one-way ANOSIM test. ANOSIM tests are based on the corresponding rank similarities between data points in the resemblance matrix, and not on actual Euclidean distances (Clarke and Warwick, 2001). The outputs included a *global R value* (comparative measure of the degree of difference between all SFTs), *global significance* or *p value*, and a series of *R statistic* and *p values* for each pair-wise comparison of SFTs.

*Objective 3)* To characterise the physical habitat conditions for each survey using the sUAS-SfM approach, we produced maps, distribution profiles and descriptive statistics of RC water depth and point cloud roughness. Whilst other factors also contribute to the physical habitat, such as grain size and flow velocity, we did not attempt to quantify these parameters in this instance.

## Results and Analyses

### *Objective 1.a) Comparison with ground truth*

A quantitative comparison of the traditionally mapped SFTs with the independent ground truth data is provided in Table 3 (columns headed TB). The overall results are variable, as

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2  
3 indicated by the range in overall accuracy and kappa coefficient, with overall accuracy not  
4 rising above 74% for any of the three surveys. The accuracy of mapping individual SFTs is  
5 highly variable across different SFTs and between different surveys. This is particularly  
6 notable within the August survey where user's accuracies range from 0% to 100%. Areas of  
7 smooth flow and unbroken standing waves, which typically occupy the larger proportions of  
8 the channel, give user's accuracies consistently greater than 64% and 58% respectively,  
9 although again there is variability between surveys. Very poor user's accuracies (<30%) are  
10 observed for areas of upwelling during the May and August surveys. This is possibly because  
11 areas of upwelling occupy a very small proportion of the channel during these surveys. As a  
12 result, they may be more susceptible to spatial misalignments or misclassifications relating to  
13 transition zones between SFTs. Overall, comparison against independent ground truth data  
14 suggests that we have not been able to map the positions and types of surface flow with a  
15 consistently high accuracy using the traditional approach of visual assessment from the  
16 bankside.  
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#### 19 *Objective 1.b) Comparison with sUAS-SfM SFT mapping*

20 Figure 3 provides examples of SFT mapping conducted in the traditional way from the  
21 bankside (top), and from the sUAS-SfM imagery (bottom). A quantitative comparison of  
22 these two approaches is provided for all surveys in Figure 4, where percentage agreement  
23 indicates the percentage of points mapped as a given SFT by the traditional survey, which are  
24 also mapped as this SFT by the sUAS-SfM survey. These data indicate that the highest levels  
25 of agreement and greatest levels of between-survey consistency are found for the smooth SFT  
26 (76-87%). The range in agreement across SFTs within a single survey is lowest for the June  
27 dataset (54-80%). Otherwise however, levels of agreement across different SFTs and between  
28 different surveys are highly variable. For example, percentage agreements range 0-65% for  
29 no perceptible flow and 18-100% for upwelling across the three surveys. Within the August  
30 survey, percentage agreement for individual SFTs ranges from 0% (no perceptible flow) to  
31 100% (upwelling). These results may be due, in part, to the low number of comparison points  
32 falling within some of the smaller SFTs (e.g. upwelling, Figure 3). Overall, the results  
33 suggest that the traditional mapping is not able to match reliably the position and type of  
34 SFTs as mapped from hyperspatial resolution imagery across the different surveys. However,  
35 a quantitative assessment of the sUAS-SfM SFT mapping itself against the ground truth data  
36 finds that it too is incapable of producing consistently high accuracies, across the three  
37 surveys and across the different SFTs (Table 3).  
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#### 42 *Objective 2) Statistical separability*

43 The global R values and pair-wise comparisons for all surveys and scenarios are presented in  
44 Table 4. Whilst global R values are higher for Test A, they are notably low across both tests  
45 (<0.32). This indicates that despite the use of objective, spatially explicit and quantitative  
46 data to define SFTs, the differences between these SFTs are not strong for any of the three  
47 surveys or under either of the two ANOSIM test scenarios. Results for all surveys and  
48 scenarios are found to be statistically significant at  $\alpha = 0.05$ .  
49  
50

51 With regards to the pair-wise comparisons, the most consistent separation is typically  
52 observed for 'smooth-upwelling' and 'unbroken standing waves-upwelling' for both Tests A  
53 and B. Test A also produces fairly consistent high R values for SFT pairs 'unbroken standing  
54 waves-no perceptible flow' and 'ripples-upwelling', which are always statistically significant.  
55 However, low or inconsistent R values are observed in most other pair-wise comparisons.  
56 Overall, this suggests that differentiation of all SFTs cannot be conducted reliably with the  
57 use of quantitative data from either of the two test scenarios.  
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### *Objective 3) Quantifying physical habitat using sUAS-SfM data*

Example high resolution maps from the May survey and distribution profiles for all surveys of RC depth and point cloud roughness are provided in Figure 5. Descriptive statistics are shown in the inset table of Figure 6. We observe a greater mean water depth, range of depths and standard deviation of depth during the June survey than in May or August. In contrast, roughness statistics are broadly similar across all three surveys, with the August survey showing slightly higher mean, minimum, maximum and range values. Percentiles of the RC depth distribution were compared with percentiles obtained from the roughness distribution (Figure 6). Excellent linear relationships are evident ( $R^2 > 0.97$ ,  $n = 101$ ,  $p < 0.01$ ), with the slope of the trend line for the June survey being clearly different from the other two surveys due to the presence of greater water depths overall, and typically lower roughness values than observed for the equivalent depths of the May and August surveys.

### **Discussion**

Our work has demonstrated that accurate, reliable, objective identification of SFTs is not possible, even with the use of quantitative hydraulic data or hyperspatial image and topography data. This raises important questions on the suitability of SFTs as a physical quantity in process-based science and evidence-based river management. In theory, SFT mapping provides a framework for classifying river habitat, in which the detail of the microscale is summarised, with a view to enabling larger scale habitat assessments for river science and management (Newson and Newson, 2000). The successful application of this approach is thus reliant on the assumption that SFT mapping provides an accurate and consistent overview of the detailed hydraulic and geomorphic conditions (i.e. elements of the physical habitat), or at least represents the diversity of such conditions within a mesoscale length of channel. Our study has assessed a traditional SFT mapping approach against independent validation data and found that in practice, the overall classification accuracy of SFT mapping does not exceed 74%. Furthermore and crucially, SFT classification mapping accuracy is highly inconsistent between SFTs and surveys, and is not improved by mapping SFTs onto the high resolution sUAS-SfM imagery. Similar results are found by other quantitative studies (e.g. Marcus, 2002; Marcus et al., 2003; Gosselin et al., 2012).

The statistical separability of SFT pairs using quantitative variables has also been tested within this paper, and elsewhere. Some results provide quantitative evidence of the separability of specific SFT pairs using hydraulic or geomorphic data (e.g. Reid and Thoms, 2008; Newson and Newson, 2000; Zavadil et al., 2012), yet no single study has been able to demonstrate statistical separability of *all* SFT pairs, leading many to suggest that the amalgamation of certain units is necessary (Wadson and Rowntree, 1998; Reid and Thoms, 2008; Zavadil et al., 2012). Additionally, wider results lack consistency between studies conducted on different river systems and under differing flow conditions (Gosselin et al., 2012), lending further support to the notion that traditional SFT mapping is not reliable as a universal method for characterising hydraulic and geomorphic conditions. This may in part relate to visual misclassifications of SFTs due to similar surface topography signatures and varying scene illumination conditions, such as the commonly recognised difficulty in distinguishing between ripples and unbroken standing waves (Padmore, 1998; Milan et al., 2010). However, we anticipate that the observed lack of reliability also results from within-SFT hydraulic and geomorphic heterogeneity and genuine overlaps in conditions between SFTs, which are reflected within our microscale, quantitative validation data. This has been observed elsewhere in the form of transition zones and as small patches of hydraulic conditions (sometimes expressed as different surface patterns) typically associated with one

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3 SFT being found within another (e.g. Wadson, 1994; Wadson and Rowntree, 1998; Marcus  
4 2002; Marcus et al., 2003; Legleiter and Goodchild, 2005; Milan et al., 2010; Wallis et al.,  
5 2010). For example, the work of Wallis et al., (2010), on an adjacent reach of the same River  
6 Arrow we study here, found that transition zones between hydraulic patches occupied a  
7 significant proportion of the channel extent (33-38%) over a range of discharges (Q13 to  
8 Q70). Using remote sensing approaches, the works of Marcus (2002) and Marcus et al.,  
9 (2003) note the presence of small, pixel-scale patches on classifications of instream habitat  
10 produced from 1m resolution hyperspectral imagery. They argue that this heterogeneity  
11 represents real differences in physical habitat, which are often returned as false  
12 misclassifications due to the coarser scale validation units mapped in the field. Harvey and  
13 Clifford (2009) explored the microscale hydrodynamics of physical biotopes and found that  
14 hydraulic heterogeneity not only existed within biotopes but that the magnitude of this  
15 heterogeneity varied between biotopes. They also highlighted the importance of this  
16 microscale hydraulic diversity (such as refugia and other marginal features) as having a more  
17 direct influence on the survival of in-stream biota than hydraulic variability at the mesoscale.  
18 Such findings underline the importance of scale in assessments of physical river habitat, and  
19 further reinforce the idea that traditional SFT mapping at the mesoscale greatly simplifies the  
20 continuum of microscale hydraulic and geomorphic conditions that are actually present  
21 (Clifford et al., 2006). Whilst this is a known characteristic of classification schemes in  
22 general, the wider findings of this paper lead us to question the on-going use of SFTs for  
23 characterising physical river habitat.  
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28 We propose that the time has come to move away from broad, mesoscale classification  
29 schemes for characterising the hydraulic and geomorphic elements of physical habitat, and  
30 towards spatially continuous, spatially explicit, quantitative measurements at the microscale.  
31 Initially, the SFT classification scheme was developed from a need for rapid, inexpensive  
32 surveys, which did not require the need for specialised knowledge. At that time, “...*the*  
33 *‘luxury’ of a full hydraulic description of habitat at the microscale...*” was not available, and  
34 “...*must await technological progress*” (Newson and Newson, 2000, p. 202). We argue that  
35 significant technological advances in sUAS platforms and SfM processing techniques now  
36 provide an alternative approach for assessing physical habitat, which is still rapid,  
37 inexpensive and becoming increasingly accessible to the non-specialist. Furthermore, sUAS-  
38 SfM approaches are capable of providing the ‘luxury’ of high resolution, quantitative,  
39 spatially continuous and explicit data at the microscale. We have demonstrated the generation  
40 of orthophoto and topography data (water depth and point cloud roughness) using this  
41 approach. Such data provide new opportunities for examining the detailed heterogeneity of  
42 hydraulic and geomorphic conditions, including small and marginal features, which might  
43 otherwise be overlooked by broad classification systems. Next, we can examine the  
44 ecological importance of this heterogeneity, its spatial significance and dynamics over time,  
45 as well as summarise the detail of this information over a range of spatial scales. Why would  
46 we continue to infer the characteristics of mesoscale physical habitat units (using traditional  
47 SFT mapping), when we are now capable of measuring them quantitatively at the microscale?  
48 Of course, the answer to this question will be determined by the specific requirements of the  
49 intended application, as well as the availability of resources, time and funds. In our  
50 experience, the rapid mobilisation and image acquisition using the Draganflyer X6 means  
51 that c. 500m lengths of channel can easily be covered within a day’s fieldwork with a team of  
52 two people. Newer fixed-wing sUAS are capable of even faster acquisitions, resulting from  
53 longer battery life and autopilot functions, so that up scaling of this approach to larger  
54 reaches is feasible. In comparison to other remote sensing approaches, these are larger areas  
55 and much quicker acquisitions than currently possible using equivalent high resolution  
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3 methods like TLS (e.g. Large and Heritage, 2007; Milan et al., 2010), and significant time  
4 and cost savings over the commissioning of multi- or hyper-spectral aerial surveys (e.g.  
5 Marcus, 2002; Marcus et al., 2003), albeit for smaller spatial areas. In terms of the initial  
6 purchase, sUAS prices have dropped dramatically in recent years so that a platform broadly  
7 equivalent to the Draganflyer X6, which was acquired for approximately £29,500 in 2010,  
8 can now be purchased for around £1,150 (GBP). In terms of the processing software,  
9 Agisoft's PhotoScan Pro can be purchased for \$549/\$3499 (USD) for educational or  
10 commercial stand-alone licences respectively.  
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13 Whilst we advocate a sUAS-SfM approach for quantifying physical habitat conditions, we  
14 note that further testing over a range of different fluvial settings and flow conditions is  
15 needed. We also note some important caveats concerning this method. For example, accurate  
16 water depth estimation requires clear water, shallow depths, minimal white water and  
17 sufficient knowledge of the SfM process to avoid introducing systematic errors. Readers are  
18 referred to Woodget et al., (2015) for a more detailed evaluation of the water depth  
19 quantification method. In terms of point cloud roughness, whilst we do not observe a direct  
20 correlation with hydraulic conditions, our detailed local knowledge of the River Arrow site  
21 leads us to suggest that a combination of water surface topography, water depth and  
22 submerged grain size are responsible for the observed roughness signature. As such, the maps  
23 and distribution profiles of roughness (Figures 5 and 6) provide useful indicators of physical  
24 habitat heterogeneity, and future work should explore the ecological significance of these  
25 patterns. However, we suspect some factors that do not have a direct influence on physical  
26 habitat also influence point cloud roughness, namely sUAS image quality and water clarity.  
27 Image quality can be degraded by blurring, caused by poor scene illumination or a lack of  
28 platform or sensor stability (de Haas et al., 2014). Blur introduces noise to the point cloud.  
29 This results in patches of erroneously high roughness, which may be variable across the  
30 scene. Greater water turbidity and depth are likely to reduce water clarity, causing material on  
31 the channel bed to become obscured and thus reduce the texture within the sUAS imagery.  
32 The SfM process is heavily reliant on texture for accurate point matching. Where texture is  
33 lacking, point matching is less successful and either fails completely, or produces a noisy,  
34 rough point cloud. Therefore, we anticipate that the sUAS-SfM approach is not appropriate  
35 for much deeper and more turbid waters than shown here. Woodget et al., (2015) find success  
36 at depths up to 0.7m, although we have not yet quantified at precisely what depth and  
37 turbidity level the method would begin to fail. Fortunately, ongoing research and  
38 development means that a reduction of the impact of image quality and water clarity on point  
39 cloud roughness may be possible. For example, improvements in sUAS camera gimbals and  
40 automated methods for blur removal are already in progress (e.g. Sieberth et al., 2015), and  
41 whilst water clarity issues are widely recognised as limiting factors in fluvial remote sensing  
42 studies in general (Gilvear et al., 1995), they might be ameliorated to some extent in future  
43 with the use of higher bit depth imagery. We also note that there remain some important  
44 physical habitat parameters that we have not quantified using the sUAS-SfM approach within  
45 this paper, including cover, flow velocity and grain size. Given the rapid developments  
46 occurring in this field however (e.g. Rivas Casado et al., 2015; Tauro et al., 2015; Woodget  
47 et al., 2016), we anticipate that it is only a matter of time before the quantification of these  
48 parameters is also possible. As such, we believe that the sUAS-SfM approach has great  
49 potential, and with further rigorous and quantitative testing, could become the tool of choice  
50 for routine and reliable assessments of physical river habitat in future.  
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## 56 **Conclusions**

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3 Within this paper, we have investigated the accuracy and reliability of traditional SFT  
4 mapping using a range of independent, quantitative data. We have also explored the potential  
5 of a novel remote sensing approach as an alternative means of characterising physical habitat  
6 within a small, shallow, lowland river in the UK. The overall accuracy of SFT mapping was  
7 found to be variable across three repeat surveys of the same reach, and did not exceed 74%  
8 on any occasion. The accuracy of mapping specific SFT units was also found to be highly  
9 variable between SFTs and different surveys. Analysis of similarity tests provided evidence  
10 that the SFTs considered within this study did not have a strong identity and thus could not be  
11 reliably differentiated from each other using quantitative hydraulic and geomorphic data. This  
12 led us to question the use of traditional SFT mapping as a universal approach for  
13 characterising physical habitat accurately and reliably, for both science and management  
14 applications. In contrast, a novel remote sensing approach using an unmanned aerial system  
15 and structure-from-motion photogrammetric processing was able to provide high resolution,  
16 spatially continuous, quantitative data to describe physical habitat at the microscale. Our  
17 results suggest that with continued testing and development, this approach holds great  
18 potential for becoming a valuable tool for quantification and monitoring of physical river  
19 habitat.  
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### 22 **Acknowledgements**

23 This work was conducted as part of a University of Worcester funded PhD studentship. The  
24 fieldwork assistance of James Atkins, Robbie Austrums, Jenni Lodwick, William Woodget  
25 and Martin Wilkes was much appreciated. Advice on using PRIMER from Tory Milner is  
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27 co-operation and two reviewers for helpful comments on an earlier draft.  
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For Peer Review



## Tables

Table 1. Descriptions of SFTs and associated biotopes  
(Newson and Newson, 2000; Environment Agency, 2003).

Surface Flow Type	Description	Associated biotope(s)
Free fall (FF)	Where vertically-falling water clearly separates from the 'back-wall' of a distinct vertical rock face.	Waterfall
Chute (CH)	Low, curving flow with substantial water contact 'hugging' the substrate.	Spill or Cascade
Broken standing waves (BSW)	Water appears to be trying to flow upstream. A white water tumbling wave must be present.	Cascade, Rapid or Riffle
Unbroken standing waves (USW)	'Babbling' water with a disturbed 'dragon-back' surface, which has upstream facing wavelets that have not broken.	Riffle
Chaotic flow (CF)	A chaotic mixture of several faster flow types (free fall, chute, broken and unbroken standing waves) in no organised pattern.	
Rippled (R)	Water surface with distinct, symmetrical, small ripples that are generally only a centimetre or so high and moving downstream.	Run
Upwelling (UP)	Upwellings are found where strong upward flow movements disturb the surface, creating an appearance of bubbling or boiling water.	Boil
Smooth (S)	Laminar flow where movement does not produce a disturbed surface.	Glide
No perceptible flow (NPF)	In ponded reaches, where it may be difficult to perceive any surface water movement.	Pool or Marginal Deadwater

Table 2. Metrics used to assess the accuracy of SFT mapping (after Liu and Mason, 2009).

Accuracy measure	Definition
Overall accuracy $\alpha = \left(\frac{1}{N}\right) \sum_{i=1}^m C_{ii}$	Percentage representing total number of correct classifications (indication of the accuracy of the SFT mapping as a whole).  N = total number of ground truth observations C <sub>ii</sub> = total number of correctly classified observations
User's accuracy $\beta = \frac{C_{ii}}{Nr_i}$	Percentage of correctly classified observations within a single mapped SFT category as a proportion of the total number of observations in that category (i.e. the commission errors).  C <sub>ii</sub> = total number of correctly classified observations in any given SFT category Nr <sub>i</sub> = total number of observations in a particular SFT category (as mapped from bankside/from orthophoto)
Kappa co-efficient (or Cohen's k) $\kappa = \frac{(N \cdot \sum_{i=1}^m C_{ii}) - (\sum_{i=1}^m (Nr_i \cdot Nc_j))}{N^2 - ((\sum_{i=1}^m (Nr_i \cdot Nc_j))}$	Statistical measure of the difference between (a) the observed agreement and (b) chance agreement. A maximum kappa value of 1 would suggest that the SFT mapping is 100% better than one resulting purely by chance and a value of 0 would suggest it is no better.

Table 3. Summary of accuracy measures of the traditional bankside SFT mapping (TB) and SFT mapping conducted on the sUAS-SfM imagery (sUAS), by comparison with independent ground truth data. Percentage of ground truth validation points classed as each SFT is given in columns headed GT. Note: this table does not consider BSW and therefore for the May survey the percentage of ground truth validation points does not sum up to 100%.

Date of Survey		May			June			August		
Method		TB	sUAS	GT	TB	sUAS	GT	TB	sUAS	GT
Overall Accuracy (%)		66	75	<i>N/a</i>	74	56	<i>N/a</i>	57	48	<i>N/a</i>
Kappa Coefficient		0.54	0.66	<i>N/a</i>	0.59	0.40	<i>N/a</i>	0.30	0.27	<i>N/a</i>
User's accuracies (%) / Percentage of ground truth validation points	Smooth	71.0	85.7	28%	66.7	44.0	23%	64.7	48.5	31%
	USW	79.4	83.3	34%	85.7	76.5	39%	58.3	73.3	34%
	Rippled	50.0	58.3	25%	61.1	33.3	23%	37.5	0.0	21%
	NPF	66.7	80.0	9%	71.4	100.0	9%	100.0	<i>N/a</i> *	7%
	Upwelling	28.6	75.0	3%	100.0	100.0	6%	0.0	33.3	7%

Table 4. ANOSIM test results. Please refer to Table 1 for SFT abbreviations. \*Denotes results which are statistically significant at the  $\alpha = 0.05$  level. Pair-wise comparisons comprising BSW are not provided for Test B for the May 2013 data because the SfM process failed to produce a reliable refraction-corrected water depth value in the area of BSW.

Furthermore, this SFT was not recorded by the ground truth survey or the sUAS mapping during the June or August campaigns. Pairs shown in bold text are referred to in the main text as most separable.

ANOSIM Results		Test A			Test B		
		May	June	Aug	May	June	Aug
Global R		0.279*	0.318*	0.197*	0.183*	0.221*	0.184*
Pair-wise comparisons	S, USW	0.249*	0.271*	0.093*	0.19*	0.099	0.145*
	S, R	0.114*	0.204*	-0.045	0.101*	0.220*	0.016
	S, NPF	0.428*	0.066	0.591*	0.261*	0.288	0.249
	<b>S, UP</b>	<b>0.790*</b>	<b>0.728*</b>	<b>0.701*</b>	<b>0.444*</b>	<b>0.962*</b>	<b>0.287</b>
	USW, R	0.051	0.198*	-0.004	0.039	0.196*	0.212*
	USW, NPF	0.661*	0.516*	0.712*	0.395*	-0.189	0.627*
	<b>USW, UP</b>	<b>0.776*</b>	<b>0.746*</b>	<b>0.687*</b>	<b>0.730*</b>	<b>0.565*</b>	<b>0.373*</b>
	R, NPF	0.406*	0.335*	0.676*	0.191*	-0.083	0.017
	R, UP	0.564*	0.472*	0.864*	0.365*	0.504*	0.054
	NPF, UP	0.174	0.259	0.556	-0.196	0.944*	-0.185
	NPF, BSW	0.982	n/a	n/a	n/a	n/a	n/a
	R, BSW	0.932*	n/a	n/a	n/a	n/a	n/a
	USW, BSW	0.881*	n/a	n/a	n/a	n/a	n/a
S, BSW	0.999*	n/a	n/a	n/a	n/a	n/a	
BSW, UP	0.556	n/a	n/a	n/a	n/a	n/a	

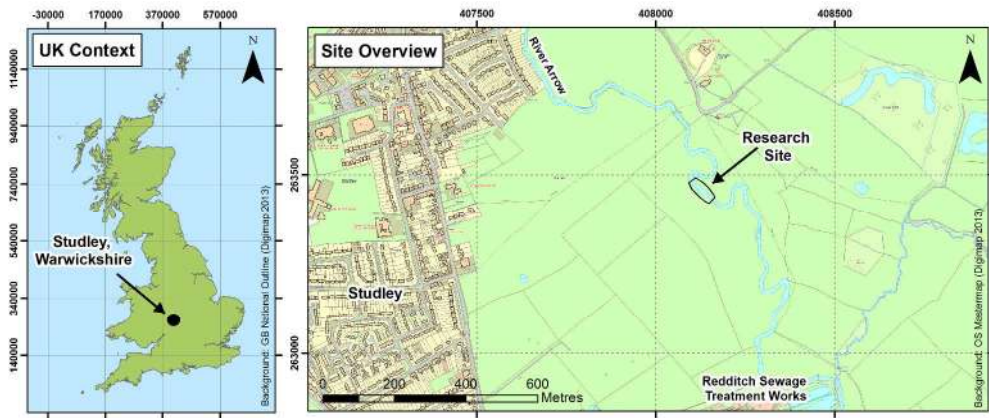


Figure 1. Location of the River Arrow field site.  
279x117mm (300 x 300 DPI)

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Figure 2. The Draganflyer X6 UAS.  
259x183mm (72 x 72 DPI)

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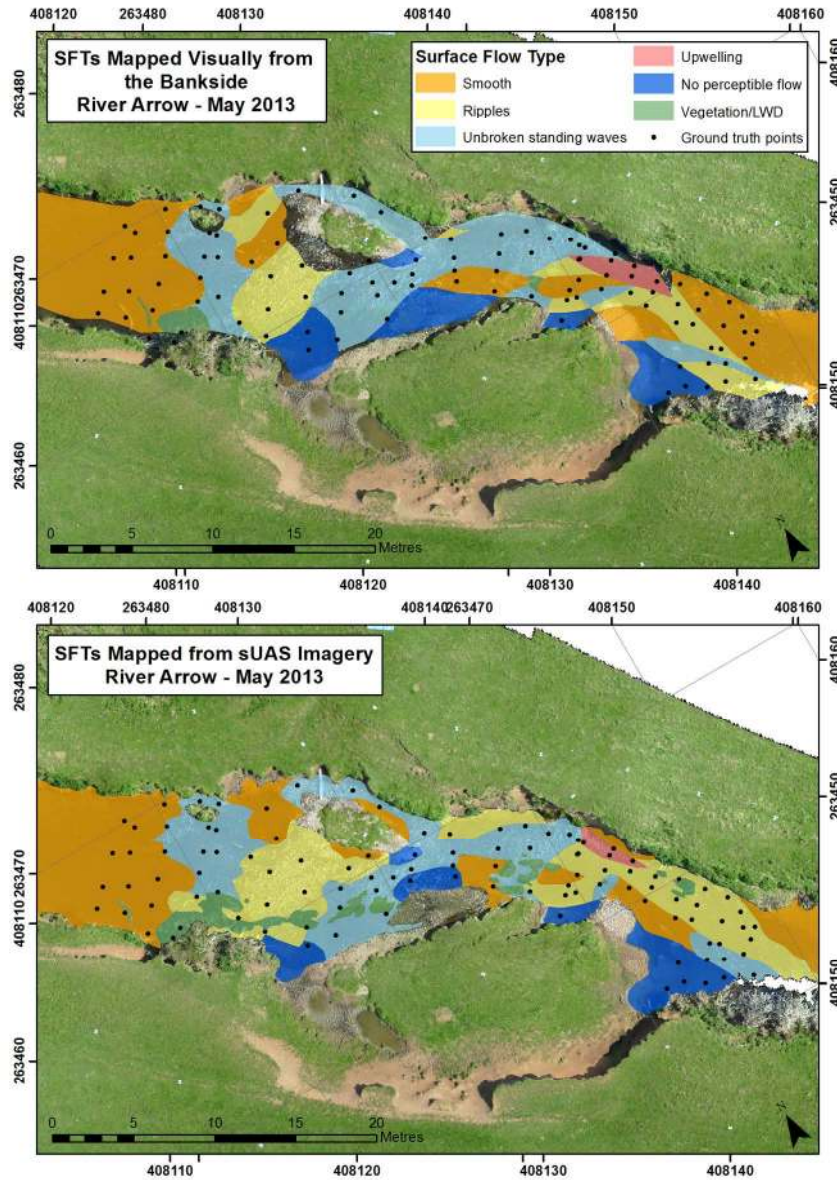


Figure 3. SFT mapping conducted visually from the bankside (top) and from the sUAS-SfM orthophoto (bottom) for the River Arrow May 2013 survey. LWD denotes areas of large woody debris. 210x297mm (200 x 200 DPI)

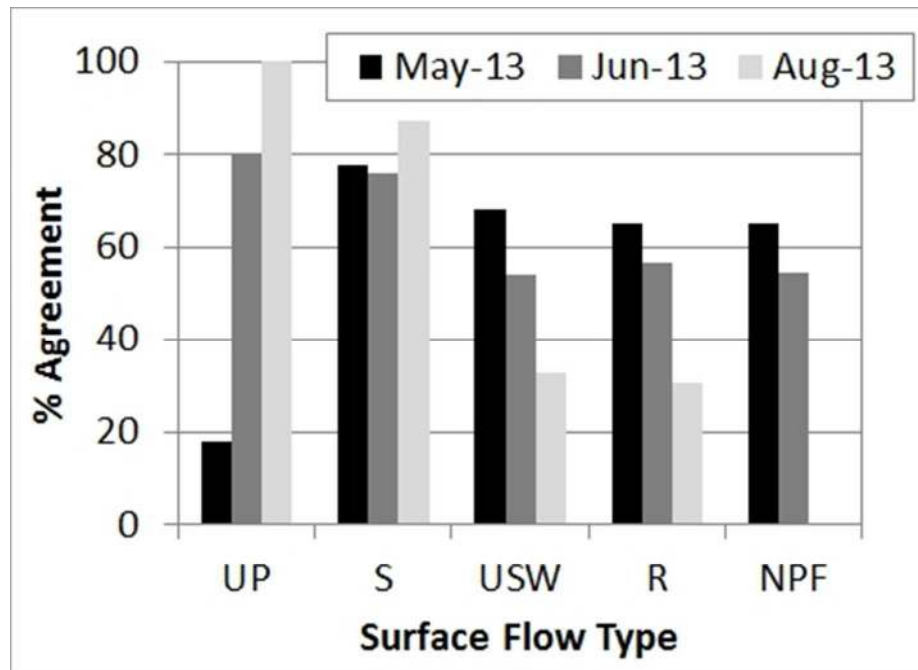


Figure 4. Level of agreement (%) between bankside and UAS-SfM mapped SFTs by survey and SFT.



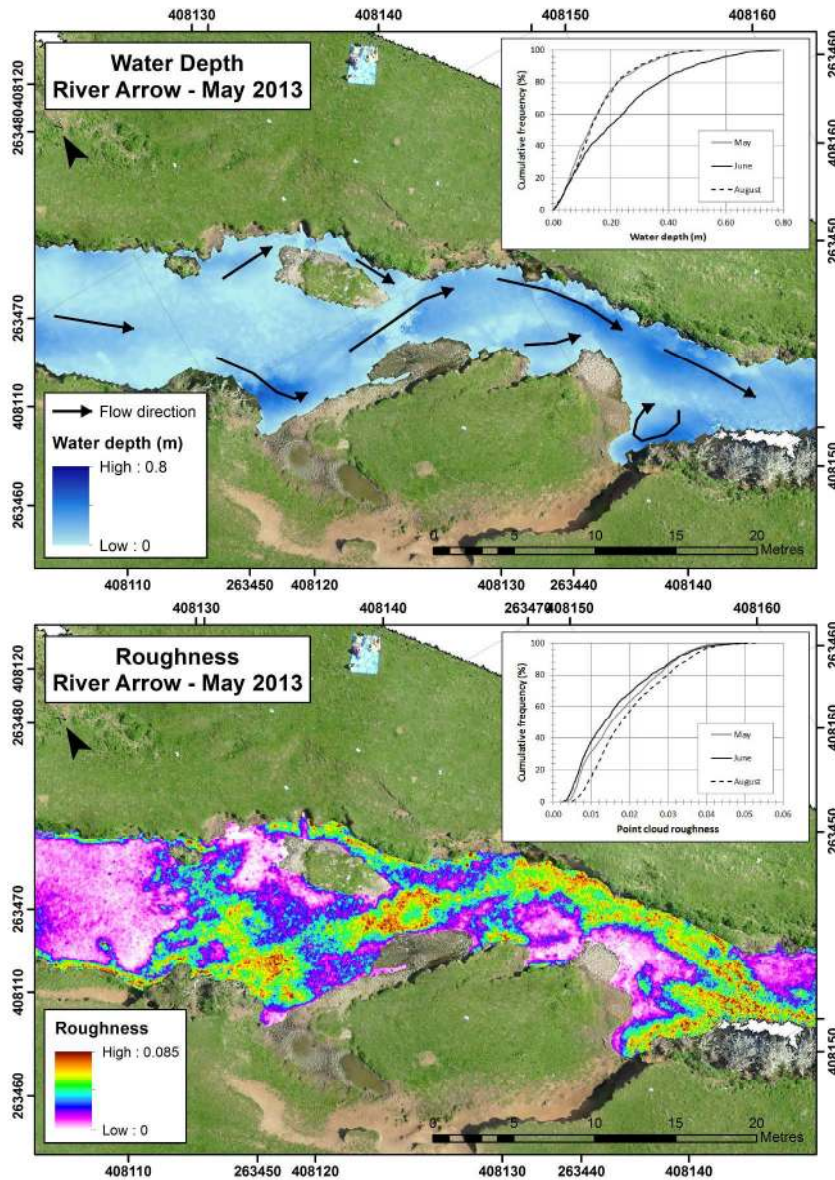


Figure 5. Spatially continuous water depth (top) and point cloud roughness data (bottom), derived from the sUAS-SfM process for the River Arrow May 2013 survey. Inset graphs: Cumulative frequency distributions of water depth and roughness for May, June and August surveys.  
297x420mm (300 x 300 DPI)

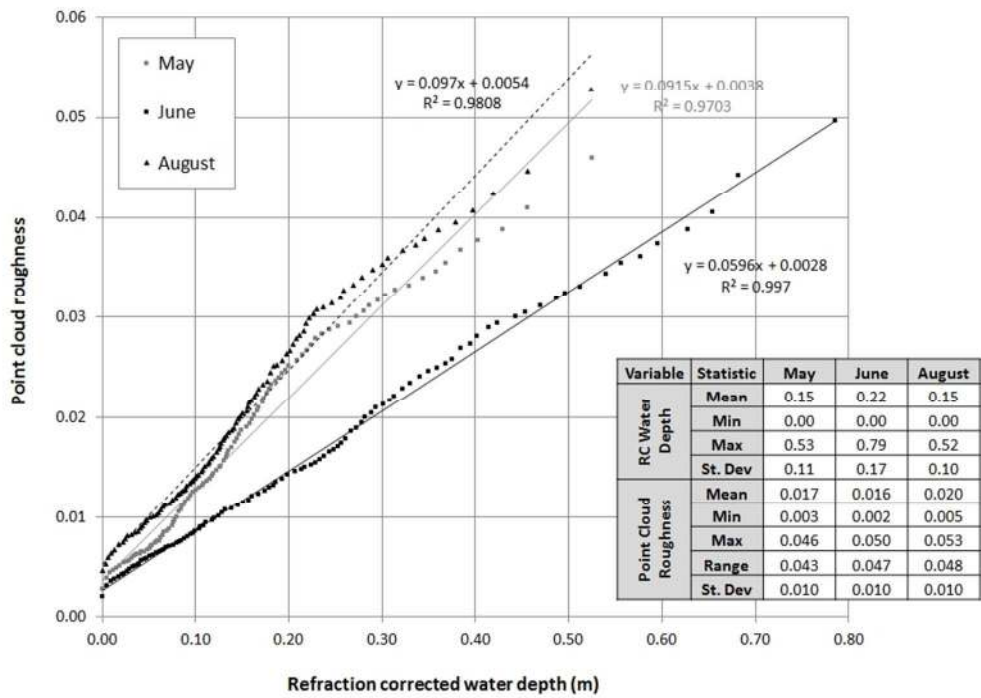


Figure 6. Regression of point cloud roughness percentiles against RC water depth percentiles by survey date. Inset table: Descriptive statistics by survey date.