

The Arthropod Community, Especially Crustacea, as a Bioindicator in Algeciras Bay (Southern Spain) Based on a Spatial Distribution

J.E. Sánchez-Moyano and J.C. García-Gómez

Laboratorio de Biología Marina
Dpto. Fisiología y Biología Animal
Facultad de Biología
Universidad de Sevilla, Apdo. 1095
41080 Sevilla, Spain

ABSTRACT



SÁNCHEZ-MOYANO, J.E. and GARCÍA-GÓMEZ, J.C., 1998. The arthropod community, especially crustacea, as a bioindicator in Algeciras Bay (southern Spain) based on spatial distribution. *Journal of Coastal Research*, 14(3), 1119–1133. Royal Palm Beach (Florida), ISSN 0749-0208.

Spatial variation of the arthropod communities of the alga *Halopteris scoparia* in relation to the influence of environmental factors in the Algeciras Bay (Southern Spain) was investigated by means of multivariate analyses. We have been able to determine a clear separation in composition of communities between external and internal areas of the bay, with hydrodynamism and algal morphology being the most determinant factors. In the external zone a greater number of species appear, whose quantitative dominances are more even than among the species of the internal zone, where some clearly dominate over the others, e.g., the amphipods *Corophium acutum* and *Jassa marmorata*. Crustaceans show in a clear way the differences among the localities and are useful in characterization studies of environmental quality of coastal waters because of the abundance, high species richness and the wide environmental spectrum in the epiphytic communities.

ADDITIONAL INDEX WORDS: Epifauna, environmental factors, hydrodynamism, sedimentation, habitat complexity, canonical correspondence analysis, *Halopteris scoparia*, Southern Spain.

INTRODUCTION

The arthropods, and more specifically the crustaceans, have been often used in macrophytobenthic studies (BIBILONI, 1981; JIMENO, 1993; CASTELLÓ *et al.*, 1987; GARCÍA-RASO, 1988; COSTELLO and MYERS, 1987) to show various relationships of predation and competition (NELSON, 1979a and b; COEN *et al.*, 1981; GUNNILL, 1984; EDGAR, 1990a and b; HOLMLUND *et al.*, 1990; POORE, 1994) or to establish the environmental patterns that control the communities (BELLAN-SANTINI, 1964, 1966; FENWICK, 1976; STONER, 1983; AOKI, 1988; PROCACCINI and SCIPIONE, 1992; RUSSO, 1989; AOKI and KIKUCHI, 1990; ARRONTES and ANADÓN, 1990; CONRADI, 1995; among other). In this sense, the crustacean communities have been considered as one of the most sensitive to the changes produced by environmental variables (DOMMASNES, 1968, 1969; DESROSIERS *et al.*, 1982, 1986 and 1990; MOORE, 1986; GRAHAME and HANNA, 1989).

Our objective has been to analyze variation in the composition of arthropod communities (mainly crustaceans) associated with the alga *Halopteris scoparia* (L.) Sauvageau (Phaeophyta, Sphacelariales) in relationship to the various environmental conditions found in the Bay of Algeciras (Southern Spain). The selection of this algal species was to eliminate the variations in structure of the community provoked by configuration of the alga, assuming, of course, that

the intraspecific variations provoked as well by the different environmental conditions will be less than among various substrata (SÁNCHEZ-MOYANO, 1996). According to HACKER and STENECK (1990), it is a branching alga within a small space. This means, on the one hand, that the abundance of individuals will be high because of the great quantity of available microhabitats, in addition to constituting a good sediment and periphyton trap (especially diatoms which constitute the diet of most of the epifauna—Orth and Van Montfrans, 1984), and additionally, it restricts the size of the organisms. It is one of the most abundant algal species and most widely distributed in Algeciras Bay. This bay, located in the most eastern zone of the Strait of Gibraltar, is composed of heterogeneous environments imposed by the natural configuration of the coast as well as by intense port and industrial activity. This makes diverse environmental conditions in spite of its relatively small extent (approximately 30 km long with a width of 8 km at its extreme), something which must influence the composition of the associated arthropod communities.

MATERIALS AND METHODS

A total of 13 sampling stations were chosen and grouped in five areas distributed along the coast of the Bay of Algeciras in order to encompass the broadest range of environmental conditions (Figure 1). Island las Palomas (IP) is a photofile rock zone and located in the extreme west of the bay.

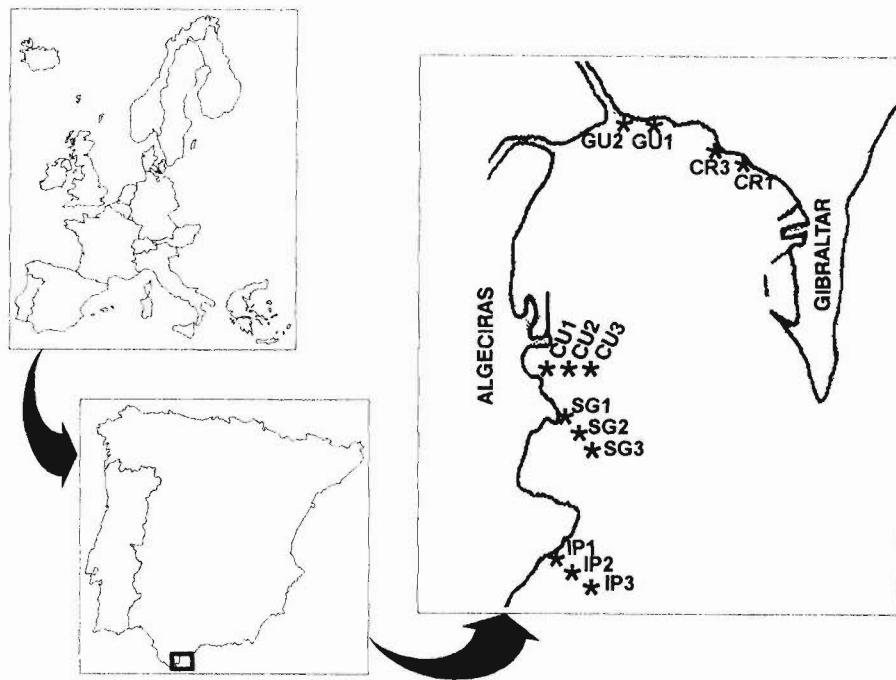


Figure 1. Location of the sampling stations in Algeciras Bay.

Punta de San García (SG), with similar characteristics, though a sciafile environment in more shallow waters. In both areas, the stations were established in a transect 200 meters long and 5, 8 and 10 meters deep. Inlet of the Cucareo (CU), nearby area to the Port of Algeciras, is located on a wide platform that varies between 3 and 5 meters deep. The stations were established on a transect of 200 meters. All stations were numbered beginning from the most shallow. Los Rocadillos (Guadarranque) (GU) is located in the internal zone of the bay, near the mouth of the Guadarranque River on a natural rock band along the coast, between 3 and 5 meters deep. Unable to establish transects in depth, the two stations were located according to their distance from the outlet: GU1, the farthest; and GU2. In the internal zone, Crinavis (CR) is a artificial substratum belonging to a shipyard in disuse. The stations were located along this breakwater, to a depth of 5 meters.

Four replicate samples were taken at each station during five samplings throughout one year (September 92; December 92; March 93; June 93; and September 93). Each sample consists of an alga specimen, pocketed *in situ* and extracted from the bottom by SCUBA diving.

The samples were put through a sieve of 0.5 mm of light mesh and were then separated, classified and quantified into the different species. The abundance data were in terms of number of individuals per 100 dry weight grams of alga.

Once the fauna were separated, a series of statistics was calculated for each alga: wet and dry weight, maximum height, diameter and volume. The volume was estimated by water displacement of the alga in a manometer. Assuming that *H. scoparia* adopts a geometric form in the environment

like a paraboloid, the theoretical volume was calculated from this. Deducting the real volume from the theoretical one we obtained the interstitial volume, which represents the living space. The relationship between the theoretical volume and the real one gives an idea of the level of compactness of the alga (Index of Compactness), in such a way that how much more is approximated to one it will be more compact. This compactness can measure the habitat complexity.

In order to establish which parameters influence the composition of the arthropods communities some concrete modules were sounded along the arch of the Bay (Figure 2). The samplings were taken monthly from November of 1992 until November of 1993.

The variables measured were: maximum and minimal temperature, hydrodynamism, sedimentation rate, % organic matter of the sedimentation, solid and % of organic matter in suspension.

For hydrodynamism (or water movement), the method of "plaster dissolution" described by MUUS (1968) and modified by GAMBI *et al* (1989) was used, measured as "equivalent speed of the water" (V) (BAILEY-BROCK, 1979).

The sedimentation rate was measured by the placement of sediment traps (six bottles of a liter capacity in our case). Part of the sedimentation is used to calculate the % of organic matter through calcination at 500° C. The data are expressed as gr/m²/month.

The solids and organic matter in suspension were measured according to the method of STRICKLAND and PARSON (1969).

The possible spatial variations of these variables were tested through one-way ANOVA, after verifying normality (Kol-

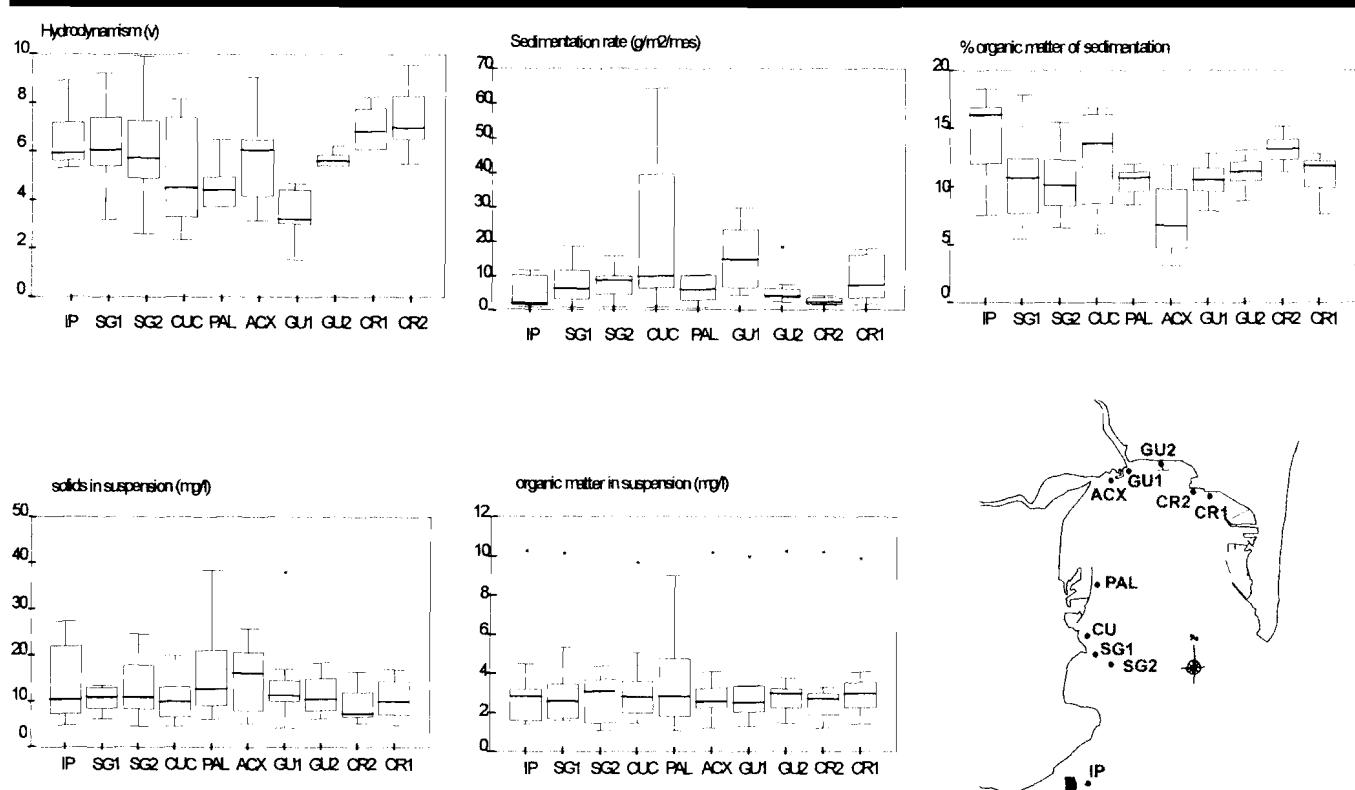


Figure 2. Variation of the environmental parameters in the stations. The thick line corresponds to the median; the rectangles contain 50% of the values, between 1st and 3rd quartile; the thin lines connect the extreme values, unless located at a distance superior to 3 times the height of the rectangle than they are indicated by an asterisk.

mogorov-Smirnov test) and variances homogeneity (Barlett test). The homogeneous groups were separated by the Tukey test.

The affinities between stations were established through a cluster analysis using the UPGMA method (unweighted pair-group method using arithmetic averages) (SNEATH and SOKAL, 1973) and based on the index of similarity of Bray-Curtis (BRAY and CURTIS, 1957). The data of abundance were transformed by the fourth root ($^4\sqrt{x}$). In order to confirm the results of the cluster, an MDS analysis (non-metric multidimensional scaling) was used. To test the veracity of ordination, the stress coefficient of Kruskal was employed (KRUSKAL and WISH, 1978).

To determine the responsible species for the groupings, the analysis of percentages of similarity or SIMPER (PRIMER package) was used. It calculates the contribution of each species either to the dissimilarity between different groups (discriminating species) or for similarity within a group (typical species). The differences in community composition were tested through the non-parametric ANOSIM test from the PRIMER package (CLARKE and GREEN, 1988).

To determine if environmental variables influence the composition of the community, we applied two techniques: BIO-ENV analysis and canonical correspondence analysis (CCA). The BIO-ENV analysis (CLARKE and AINSWORTH, 1993) consists of the comparison through the harmonic rank correla-

tion coefficient of Spearman of the rank similarity matrix (based on the index of Bray-Curtis) accomplished with the data of abundance of species and the rank similarity matrix obtained through euclidean distances with the abiotic variables, and thus establishing that environmental variables favor the maximum correlation between the two.

The canonical correspondence analysis (CCA) is a direct gradient technique where it is obtained by graphic representation where the stations and/or species are shown by points and the environmental factors by arrows. The station points and species together represent the dominant model in the composition of the community in each zone so much as can be explained by the environmental variables, while the species and the arrows reflect the distribution of the species throughout environmental gradients (TER BRAAK, 1986, 1990). In order to avoid the distortion provoked in the analysis by the rare species, these were downweighted. To verify the statistical significance of the analysis, the Monte Carlo test for the first axis was applied.

RESULTS

Environmental Variables

The general trend of abiotic variables throughout the bay is shown in Figure 2 and the significant differences among sites in Table 1. The most external zones of the bay present

Table 1. Results of the one-way ANOVA for the environmental variables. df: degrees of freedom; (MS) Squared of the average; (F) statistic; *: $P < 0.001$; **: $P < 0.05$; ns: not significant. The homogeneous group according to the Tukey test ($P > 0.05$) are indicated with a continuous line.

| | df | MS | F | Homogeneous Group |
|---------------------|----|--------|---------|--|
| Hydrodynamism | 9 | 0.307 | 29.66* | CR1 CR2 IP SG1 SG2 ACX CU PA_ GU2 GU1 |
| Sedimentation | 9 | 80.211 | 171.97* | IP SG1 SG2 PA_ CR1 CR2 GU2 GU1 CU ACX |
| Org. mat. sediment. | 9 | 23.607 | 2.54** | ACX_ IP SG1 SG2 CU PA_ CR1 CR2 GU1 GU2 |
| Solids suspension | 9 | 45.185 | 1.09ns | |
| Org. mat. suspens. | 9 | 0.016 | 0.08ns | |

a greater hydrodynamism (water movement) with a clear trend to reduce toward the interior, a rise in the area of Crinavis (CR1 and CR2) again, due to the existing tidal currents in the bay. These tidal currents are conditioned by the circulation of Atlantic and Mediterranean water masses through Strait of Gibraltar (in direction to East or West according to the flow or ebb state of the tide (GARCÍA, 1986). Thus, in flow state (high tide), a current is originated surrounding the whole interior of the bay, entering by the area of the Island Las Palomas and leaving by Punta Europa (Gibraltar). While, in ebb state, the superficial current of the Strait is divided into two branches, one enters by Gibraltar and other by Island Las Palomas, so that two currents are originated with parallel circulation to both margins of the bay and that are meeting in the interior (area of Crinavis), originating a new current of direction South by the central zone joining with the general current of the Strait. The sedimentation rate has a slightly inverse behavior, with a greater rate in the most internal zones, especially those located in the influence areas of the outlets of the rivers Palmes (ACX) and Guadarranque (GU1). In the case of the Inlet of Cucareo (CUC) a lot of movement on the bottom is joined to the sedimentation rate due to its shallow depth (5-3 m) and to the influence of the surge provoked by the strong prevailing east winds in the zone. The solids in suspension behave similarly (although not statistically significant), with a trend to increase toward the interior and with a notable influence of the presence of the rivers (ACX and GU1) and of the proximity

to urban sewage; for example, in the external dike of the Port of Algeciras (PAL).

With respect to the morphological characteristics of the algae (Table 2), a trend toward a greater interstitial volume (direct measurement of the useful vital space, considering the surface of the fronds as well as the existing spaces between them) was observed, as well as compactness in the external zone and a gradual decrease toward the optimum growth was shown by small-sized samples, with branches close together, presumably caused by great turbidity provoked by the river.

Faunal Analysis

A total of 107 taxa were identified at 13 stations: 6 Pantopoda; 1 Insecta; and 100 Crustacea (61 Amphipoda, 5 Cumacea, 12 Decapoda, 13 Isopoda, 1 Leptostraca, 1 Mysidacea and 6 Tanaidacea) (Table 3).

Multivariate Analysis

Through the classification analysis of the stations according to the annual average abundance of the different species, we obtain a clear separation between the stations located at the external extreme of the Bay (Island Las Palomas, San García and Cucareo) and those situated close to the internal zone (Guadarranque and Crinavis) (Figure 3). The same occurs among the stations within each group, especially in the external stations. In these, there is a separation between the subgroup of Island Las Palomas and San García (both areas

Table 2. Average values of algal parameters. (CI) compactness; (INTV) interstitial volume (ml).

| | SEP92 | | DEC | | MAR | | JUN | | SEP93 | |
|-----|--------|--------|--------|--------|--------|----------|-------|--------|-------|--------|
| | CI | INTV | CI | INTV | CI | INTV | CI | INTV | CI | INTV |
| IP1 | 119.91 | 680.50 | 72.69 | 567.03 | 51.28 | 608.38 | 54.10 | 655.26 | 34.75 | 431.25 |
| IP2 | 104.47 | 835.81 | 60.95 | 750.60 | 66.53 | 852.77 | 59.23 | 778.56 | 50.20 | 516.68 |
| IP3 | 64.53 | 510.60 | 62.63 | 883.28 | 92.86 | 914.09 | 67.64 | 902.93 | 48.46 | 358.77 |
| SG1 | 29.61 | 502.43 | 105.56 | 799.29 | 100.72 | 677.91 | - | - | 29.78 | 275.43 |
| SG2 | 43.54 | 444.30 | 91.08 | 660.90 | 97.95 | 724.93 | - | - | - | - |
| SG3 | 39.16 | 333.12 | 76.75 | 682.30 | - | - | - | - | - | - |
| CU1 | 45.14 | 614.32 | 49.18 | 614.65 | 87.24 | 1,029.06 | 64.46 | 809.58 | 38.35 | 593.90 |
| CU2 | 53.00 | 484.84 | 63.34 | 519.85 | 72.59 | 737.50 | 60.19 | 824.81 | 35.95 | 615.66 |
| CU3 | 51.26 | 618.51 | 83.60 | 766.15 | 60.06 | 805.53 | 33.26 | 430.62 | 23.81 | 455.42 |
| GU1 | 33.17 | 282.17 | 47.12 | 374.38 | 37.89 | 443.76 | 13.94 | 148.95 | 25.10 | 281.02 |
| GU2 | 32.68 | 149.80 | 81.31 | 489.92 | 19.42 | 237.37 | 46.34 | 431.05 | - | - |
| CR1 | 33.11 | 385.41 | 39.09 | 479.23 | 22.92 | 429.23 | 33.85 | 363.09 | 17.17 | 175.88 |
| CR3 | 54.27 | 392.12 | 57.36 | 490.58 | 43.54 | 403.40 | 4.16 | 43.64 | 17.97 | 168.50 |

Table 3. Annual average abundance number of individuals/100 g dry alga weight) of the different species.

| Species | IP1 | IP2 | IP3 | SG1 | SG2 | SG3 | CU1 | CU2 | CU3 | GU1 | GU2 | CR1 | CR3 |
|---|--------|--------|----------|--------|----------|----------|----------|----------|----------|----------|----------|----------|----------|
| Pantopoda | | | | | | | | | | | | | |
| <i>Achelia</i> sp (ACL) | 98.68 | 29.71 | 190.99 | 221.13 | 286.59 | 388.07 | 58.70 | 228.66 | 157.60 | 550.66 | 419.14 | 325.85 | 221.44 |
| <i>Callipallenae tibieri</i> (CLL) | 56.95 | 109.91 | 107.94 | 219.54 | 145.95 | 281.25 | 143.13 | 48.89 | 40.98 | 138.83 | 29.26 | 156.70 | 54.00 |
| <i>Endeis charvadacea</i> (END) | 0.00 | 2.73 | 11.16 | 1.31 | 9.70 | 5.06 | 8.33 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Nymphon gracile</i> (NYM) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.37 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Anoplodactylus angulatus</i> (ANA) | 8.33 | 9.52 | 11.08 | 16.45 | 68.42 | 40.25 | 12.65 | 25.52 | 28.21 | 20.26 | 23.09 | 15.31 | 6.67 |
| <i>Anoplodactylus virgescens</i> (ANV) | 2.33 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.22 | 1.44 | 11.21 | 2.50 | 0.00 | 0.00 | 3.97 |
| Amphipoda | | | | | | | | | | | | | |
| <i>Iphimedia minuta</i> (IPM) | 0.00 | 10.79 | 31.68 | 18.63 | 123.06 | 85.97 | 26.69 | 15.01 | 30.78 | 2.23 | 0.00 | 45.73 | 138.06 |
| <i>Ampelisca cf. undulata</i> (AMP) | 6.02 | 0.00 | 1.66 | 0.00 | 10.21 | 0.00 | 6.06 | 16.78 | 6.49 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Amphilochus neapolitanus</i> (APN) | 268.05 | 248.86 | 220.44 | 369.13 | 446.10 | 214.56 | 334.16 | 211.89 | 255.76 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Amphilochus pucatarius</i> (APP) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 22.64 | 0.00 | 2.08 | 0.00 |
| <i>Peltocoxa gibbosa</i> (PEL) | 8.78 | 2.24 | 12.55 | 1.31 | 14.26 | 43.08 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 6.34 | 0.00 |
| <i>Amphioche ramondi</i> (AMR) | 607.32 | 301.67 | 201.84 | 336.47 | 392.85 | 54.51 | 238.14 | 1.115.28 | 601.64 | 0.00 | 0.00 | 39.06 | 47.47 |
| <i>Aora spinicornis</i> (AOR) | 146.16 | 26.13 | 60.38 | 69.14 | 217.24 | 238.81 | 94.92 | 378.46 | 74.89 | 345.37 | 499.56 | 526.05 | 1.328.92 |
| <i>Lembos vigueri</i> (LEV) | 22.60 | 2.65 | 0.00 | 102.07 | 0.00 | 73.55 | 0.00 | 10.04 | 0.00 | 6.89 | 0.00 | 0.00 | 0.00 |
| <i>Lembos websteri</i> (LEW) | 31.28 | 47.08 | 119.40 | 37.46 | 226.43 | 35.96 | 13.22 | 19.01 | 4.06 | 56.62 | 55.37 | 180.90 | 261.29 |
| <i>Leptocherius guttatus</i> (LEG) | 1.82 | 2.85 | 3.29 | 0.00 | 0.00 | 5.06 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Leptocheirus pectinatus</i> (LEP) | 15.66 | 0.00 | 6.58 | 4.46 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.45 |
| <i>Microdeutopus armatus</i> (MIA) | 0.00 | 0.00 | 0.00 | 0.00 | 34.78 | 0.00 | 1.62 | 0.00 | 0.00 | 46.80 | 20.45 | 0.00 | 0.00 |
| <i>Microdeutopus chelifer</i> (MIC) | 0.00 | 0.00 | 0.00 | 11.66 | 0.00 | 0.00 | 45.17 | 105.77 | 78.69 | 228.07 | 118.13 | 34.72 | 64.10 |
| <i>Microdeutopus stationis</i> (MIS) | 7.18 | 12.91 | 0.00 | 55.76 | 0.00 | 2.65 | 13.42 | 26.64 | 15.24 | 5.25 | 40.93 | 14.60 | 24.72 |
| <i>Microdeutopus versicoloratus</i> (MIV) | 0.00 | 4.15 | 0.00 | 0.00 | 8.83 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 13.79 |
| Aoridae (AOD) | 544.54 | 395.51 | 1.037.47 | 854.50 | 1.180.67 | 296.68 | 1.840.58 | 1.081.98 | 967.20 | 1.636.86 | 1.981.44 | 2.353.82 | |
| <i>Apherusa bispinosa</i> (APB) | 571.08 | 432.46 | 554.58 | 211.94 | 442.67 | 1.050.02 | 434.53 | 500.89 | 234.27 | 7.52 | 0.00 | 62.12 | 104.84 |
| <i>Apherusa</i> sp. (APH) | 834.42 | 57.30 | 8.85 | 22.37 | 31.10 | 47.90 | 2.514.47 | 2.557.99 | 1.495.42 | 3.24 | 0.00 | 0.00 | 0.00 |
| <i>Caprella acanthifera acanthifera</i> (CAA) | 11.88 | 3.07 | 0.00 | 0.00 | 0.00 | 0.00 | 86.61 | 49.48 | 16.14 | 0.00 | 0.00 | 5.86 | 6.67 |
| <i>C. acanthifera discrepans</i> (CAD) | 390.92 | 520.71 | 1.201.99 | 446.60 | 345.84 | 122.49 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.15 | 0.00 |
| <i>C. acanthifera typica</i> (CAT) | 26.46 | 3.05 | 0.00 | 34.39 | 75.55 | 0.00 | 55.49 | 35.39 | 13.16 | 0.00 | 0.00 | 163.13 | 23.40 |
| <i>Caprella grandinana</i> (CAG) | 55.56 | 11.32 | 0.00 | 119.01 | 0.00 | 0.00 | 10.68 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Caprella liparicensis</i> (CAL) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 76.51 | 0.00 |
| <i>Caprella mutis</i> (CAM) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 39.95 | 167.47 | 60.55 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Caprella penantis</i> (CAP) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 11.98 | 2.22 |
| <i>Colomastix pusilla</i> (COL) | 0.00 | 0.00 | 0.00 | 3.74 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 6.72 | 0.00 | 0.00 |
| <i>Corophium acutum</i> (COA) | 444.93 | 799.54 | 1.188.27 | 384.06 | 141.72 | 62.88 | 2.310.63 | 4.573.93 | 2.271.24 | 3.498.93 | 3.387.38 | 5.087.33 | 4.773.03 |
| <i>Corophium sextonae</i> (COS) | 0.00 | 0.00 | 0.00 | 16.00 | 2.85 | 13.31 | 0.00 | 0.00 | 0.00 | 3.77 | 24.02 | 141.71 | 150.18 |
| <i>Erichthonius brasiliensis</i> (ERI) | 11.03 | 0.00 | 0.00 | 0.00 | 5.10 | 10.59 | 0.00 | 1.87 | 0.00 | 1.331.99 | 2.558.90 | 2.25 | 358.53 |
| <i>Siphonocetes</i> sp. (SIP) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 165.43 | 1.37 | 9.72 | 0.00 | 10.35 | 0.00 | 20.00 |
| <i>Cressa cristata</i> (CRC) | 0.00 | 0.00 | 8.70 | 9.75 | 2.85 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Cressa mediterranea</i> (CRM) | 0.00 | 8.47 | 55.53 | 0.00 | 0.00 | 13.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Dexamine spiniventris</i> (DXS) | 57.06 | 41.89 | 47.37 | 65.23 | 26.15 | 153.86 | 48.12 | 32.73 | 12.19 | 0.00 | 0.00 | 7.19 | 5.19 |
| <i>Dexamine spinipes</i> (DXP) | 227.50 | 179.17 | 165.04 | 298.67 | 364.15 | 211.92 | 476.62 | 234.88 | 234.57 | 110.05 | 18.70 | 108.56 | 53.58 |
| <i>Guerneea coerulea</i> (GUE) | 373.99 | 336.94 | 245.07 | 25.26 | 11.27 | 10.12 | 12.41 | 2.88 | 15.12 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Tritaea gibbosa</i> (TRI) | 0.00 | 0.00 | 0.00 | 5.54 | 0.00 | 0.00 | 0.00 | 0.00 | 5.15 | 0.00 | 0.00 | 5.61 | 29.33 |

Table 3. *Continued.*

| Species | IP1 | IP2 | IP3 | SG1 | SG2 | SG3 | CU1 | CU2 | CU3 | GU1 | GU2 | CR1 | CR3 |
|--|--------|--------|----------|----------|--------|----------|--------|----------|----------|----------|----------|----------|----------|
| <i>Eusimidae</i> sp. (EUD) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Eusimoides dellavallei</i> (EUS) | 2.14 | 7.42 | 0.00 | 38.35 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 6.21 | 0.00 |
| <i>Cheiropodus sanderae</i> (CHE) | 0.00 | 10.78 | 4.95 | 0.00 | 0.00 | 54.76 | 1.76 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Elaenius</i> sp. (ELA) | 157.15 | 32.98 | 7.68 | 39.32 | 27.32 | 77.45 | 97.58 | 61.74 | 38.54 | 88.64 | 156.68 | 1.04 | 0.00 |
| <i>Gammarella fucicola</i> (GAM) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.11 | 1.37 | 1.83 | 28.37 | 10.35 | 4.03 | 0.00 |
| <i>Gammareopsis maculata</i> (GMM) | 89.38 | 62.36 | 142.91 | 1.195.32 | 778.62 | 2.184.60 | 34.41 | 267.04 | 237.18 | 1.487.34 | 1.512.24 | 547.41 | 519.32 |
| <i>Gammareopsis polynata</i> (GMP) | 320.75 | 492.53 | 1.152.35 | 1.404.81 | 891.61 | 1.684.53 | 519.14 | 3.947.78 | 1.035.56 | 53.52 | 63.84 | 2.800.17 | 6.017.42 |
| <i>Megamphopus cornutus</i> (MEG) | 0.00 | 7.59 | 4.95 | 0.00 | 363.07 | 8.01 | 0.00 | 0.00 | 1.05 | 0.00 | 0.00 | 1.04 | 5.15 |
| <i>Microprotopus</i> (MTP) | 206.92 | 310.53 | 308.31 | 362.96 | 381.41 | 390.21 | 254.62 | 211.07 | 199.36 | 2.65 | 0.00 | 3.76 | 0.00 |
| <i>Photis longipes</i> (PHO) | 21.04 | 3.68 | 56.86 | 24.65 | 48.03 | 198.81 | 0.00 | 2.02 | 0.87 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Isacidae</i> sp. (ISA) | 0.00 | 0.00 | 0.00 | 30.13 | 0.00 | 15.65 | 2.70 | 3.78 | 3.15 | 0.00 | 1.04 | 0.00 | 0.00 |
| <i>Ischyrocerus inexpectatus</i> (ISC) | 1.82 | 5.10 | 3.06 | 18.17 | 39.72 | 3.75 | 0.00 | 0.00 | 5.82 | 65.39 | 25.31 | 14.18 | 14.18 |
| <i>Jassa marmorata</i> (JAS) | 9.38 | 0.00 | 0.00 | 40.77 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 7.896.63 | 2.266.99 | 991.57 | 991.57 |
| <i>Leucathoe spinicarpa</i> (LEU) | 1.44 | 1.34 | 3.31 | 24.59 | 17.73 | 32.57 | 47.39 | 34.29 | 44.14 | 2.23 | 13.44 | 4.32 | 19.69 |
| <i>Lysianassa costa</i> cf. (LYS) | 13.06 | 85.84 | 62.94 | 30.80 | 33.07 | 33.86 | 3.30 | 11.16 | 5.62 | 0.00 | 0.00 | 4.46 | 20.69 |
| <i>Orchonene</i> sp. (ORC) | 8.48 | 12.21 | 21.60 | 7.47 | 6.17 | 23.45 | 24.15 | 16.62 | 62.47 | 7.60 | 47.38 | 32.49 | 83.37 |
| <i>Synchelidium longidigitatum</i> (SYN) | 2.53 | 2.16 | 1.66 | 0.00 | 0.00 | 0.00 | 6.30 | 3.90 | 2.99 | 27.24 | 2.59 | 13.47 | 18.91 |
| <i>Portimambus typicus</i> (PAR) | 0.00 | 0.00 | 0.00 | 5.10 | 123.41 | 0.00 | 4.25 | 5.11 | 6.81 | 23.44 | 1.74 | 9.52 | 9.52 |
| <i>Padoculina garcia-gomezi</i> (PED) | 0.00 | 32.03 | 47.11 | 13.22 | 125.43 | 427.31 | 0.00 | 0.00 | 8.92 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Pseudoprotella phasma minor</i> (PPM) | 0.00 | 18.97 | 32.98 | 13.36 | 14.37 | 7.94 | 2.44 | 11.34 | 1.62 | 3.31 | 3.36 | 0.00 | 0.00 |
| <i>P. phasma quadrispinis</i> (PPQ) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.37 | 9.03 | 0.00 |
| <i>P. phasma typica</i> (PPT) | 0.00 | 0.00 | 0.00 | 6.69 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 6.89 | 0.00 |
| <i>Peritonotus testudo</i> (PER) | 14.93 | 16.08 | 79.42 | 37.74 | 108.35 | 101.28 | 10.44 | 13.26 | 9.33 | 0.00 | 0.00 | 16.55 | 18.57 |
| <i>Harpinia</i> sp. (HAR) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.19 | 0.00 | 0.00 | 3.45 |
| <i>Platysca marina</i> (PHT) | 299.23 | 541.83 | 831.09 | 860.33 | 737.61 | 946.33 | 209.91 | 543.01 | 132.94 | 1.471.91 | 1.708.92 | 1.664.88 | 1.430.36 |
| <i>Podocerus variegatus</i> (POD) | 10.80 | 10.16 | 0.00 | 45.45 | 6.69 | 36.59 | 12.49 | 14.50 | 22.27 | 325.15 | 217.78 | 59.57 | 75.35 |
| <i>Stenothoe dolfusi</i> (STD) | 0.00 | 0.00 | 0.00 | 6.72 | 0.00 | 47.67 | 0.00 | 0.00 | 0.00 | 1.19 | 0.00 | 0.00 | 0.00 |
| <i>Stenothoe monoculoides</i> (STM) | 357.10 | 509.27 | 421.88 | 853.64 | 654.43 | 608.52 | 965.21 | 1.736.97 | 1.224.28 | 3.572.51 | 1.424.82 | 1.029.14 | 362.20 |
| <i>Hyle schmidti</i> (HYA) | 517.72 | 469.85 | 471.92 | 196.96 | 620.74 | 2.299.16 | 79.87 | 304.01 | 88.02 | 0.00 | 0.00 | 4.03 | 0.00 |
| Cumacea | | | | | | | | | | | | | |
| <i>Iphinoe</i> sp. (IPH) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.45 |
| <i>Vauchompsonia</i> sp. (VAU) | 0.00 | 0.00 | 8.98 | 0.00 | 3.35 | 0.00 | 0.00 | 0.00 | 6.14 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Cumella limitola</i> (CUL) | 77.69 | 122.31 | 65.78 | 30.49 | 56.09 | 27.58 | 109.01 | 42.45 | 69.29 | 1.356.10 | 659.74 | 244.89 | 234.78 |
| <i>Nannastacus urgiculatus</i> (NAN) | 124.47 | 153.60 | 227.21 | 166.13 | 364.15 | 284.71 | 63.08 | 59.36 | 61.52 | 8.28 | 0.00 | 24.85 | 49.41 |
| <i>Scherocumella longirostris</i> (SCH) | 3.25 | 3.11 | 11.76 | 6.17 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Decapoda | | | | | | | | | | | | | |
| <i>Athanas nitescens</i> (ATH) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.87 | 0.00 | 2.99 | 7.97 | 3.12 | 0.00 |
| <i>Dardanus</i> sp. (DAR) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 5.50 |
| <i>Hippolyte</i> sp. (HIP) | 43.70 | 71.31 | 22.18 | 121.11 | 123.35 | 121.63 | 155.28 | 30.87 | 30.41 | 20.30 | 23.51 | 74.29 | 291.91 |
| <i>Thorellus cranchii</i> (THO) | 0.00 | 5.79 | 5.16 | 22.10 | 6.38 | 17.83 | 2.09 | 8.43 | 5.50 | 0.00 | 3.12 | 11.77 | 11.43 |
| <i>Achaetus</i> sp. (ACH) | 0.00 | 5.11 | 2.31 | 8.96 | 26.80 | 0.00 | 6.72 | 7.31 | 18.12 | 19.73 | 30.95 | 11.82 | 20.16 |
| <i>Macropodita</i> sp. (MAC) | 0.00 | 0.00 | 0.00 | 9.53 | 1.77 | 0.00 | 4.08 | 0.00 | 18.17 | 11.39 | 11.18 | 0.00 | 9.48 |
| <i>Pisa carinimana</i> (PIC) | 2.33 | 0.00 | 0.00 | 8.14 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Pisa tereodon</i> (PTT) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.58 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Pugilidae</i> (PAG) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 37.28 |
| <i>Periclimenes</i> sp. (PEC) | 0.00 | 0.00 | 0.00 | 0.00 | 2.85 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Sirpus zariqueyi</i> (SIZ) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.94 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Phiumnus</i> sp. (PL) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 9.02 | 0.00 | 0.00 | 1.81 | 6.06 |

Table 3. *Continued.*

| Species | IP1 | IP2 | IP3 | SG1 | SG2 | SG3 | CU1 | CU2 | CU3 | GU1 | GU2 | CR1 | CR3 |
|---|--------|--------|----------|----------|----------|----------|----------|----------|----------|--------|----------|----------|--------|
| Isopoda | | | | | | | | | | | | | |
| <i>Astacilla cf. axelii</i> (ASA) | 7.72 | 14.72 | 20.65 | 17.08 | 26.74 | 0.00 | 0.00 | 1.24 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Astacilla</i> sp. (AST) | 0.00 | 0.00 | 8.96 | 0.00 | 0.00 | 0.00 | 0.00 | 0.04 | 7.52 | 15.19 | 0.00 | 0.00 | 0.00 |
| <i>Gnathia</i> sp. (GNA) | 7.33 | 3.05 | 9.80 | 0.00 | 17.73 | 0.00 | 1.82 | 2.04 | 3.75 | 0.00 | 0.00 | 1.12 | 6.07 |
| <i>Idotea</i> cf. <i>neglecta</i> (IDN) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.48 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Idotea</i> cf. <i>viridis</i> (IDV) | 1.44 | 1.34 | 4.52 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.16 | 7.83 | 0.00 | 0.00 | 0.00 |
| <i>Synisoma capito</i> (SYC) | 2.92 | 2.58 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.34 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Zenobiusiana prismatica</i> (ZEP) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.74 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Jaeropsis brevicornis</i> (JAE) | 12.94 | 5.27 | 4.88 | 2.71 | 2.85 | 0.00 | 1.87 | 5.13 | 1.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Janira maculosa</i> (JAN) | 11.39 | 5.22 | 44.84 | 12.27 | 0.00 | 0.00 | 0.00 | 1.37 | 0.00 | 0.00 | 0.00 | 9.03 | 0.00 |
| <i>Munna</i> sp. (MUN) | 4.12 | 7.34 | 17.77 | 0.00 | 33.57 | 23.72 | 0.00 | 9.26 | 0.00 | 21.42 | 8.47 | 2.59 | 11.69 |
| <i>Paranthura nigropunctata</i> (PTH) | 282.67 | 260.24 | 387.37 | 456.76 | 383.21 | 443.61 | 432.45 | 395.01 | 329.09 | 19.73 | 7.14 | 317.87 | 198.59 |
| <i>Cymodoce</i> sp. (CYM) | 20.62 | 28.01 | 9.90 | 53.73 | 71.86 | 20.62 | 66.53 | 61.10 | 32.02 | 41.80 | 3.50 | 230.63 | 326.31 |
| <i>Dynamene</i> sp. (DYN) | 84.06 | 14.46 | 4.72 | 39.31 | 14.62 | 6.68 | 158.60 | 324.38 | 141.68 | 401.06 | 54.08 | 26.97 | 151.39 |
| Leptostraca | | | | | | | | | | | | | |
| <i>Nebatia bipes</i> (NEB) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 10.36 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Mysidacea | | | | | | | | | | | | | |
| <i>Siriella</i> sp. (SIR) | 19.67 | 5.26 | 6.94 | 2.62 | 23.64 | 0.00 | 1.62 | 5.41 | 1.74 | 0.00 | 0.00 | 7.12 | 0.00 |
| Tanaidacea | | | | | | | | | | | | | |
| <i>Apseudes talpa</i> (APS) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 6.05 | 3.05 | 5.23 | 1.00 | 0.00 | 0.00 | 0.00 | 10.34 |
| <i>Parapselides latifrons</i> (PPS) | 0.00 | 0.00 | 9.76 | 3.06 | 17.73 | 22.03 | 3.70 | 3.05 | 0.00 | 124.95 | 294.84 | 13.43 | 0.00 |
| <i>Leprocheta dubia</i> (LCH) | 331.79 | 678.04 | 1,034.95 | 845.97 | 1,162.92 | 1,502.95 | 1,006.70 | 1,660.98 | 2,261.98 | 802.99 | 1,010.87 | 1,051.14 | 666.17 |
| <i>Tanais dulongji</i> (TAN) | 914.15 | 169.38 | 417.32 | 2,193.72 | 1,823.29 | 1,711.14 | 826.59 | 514.00 | 125.50 | 4.86 | 9.58 | 7.29 | 0.00 |
| <i>Zeuo cordensis</i> (ZEC) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 916.52 | 638.05 | 27.90 | 0.00 | 0.00 |
| <i>Zeuo normani</i> (ZEN) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 226.54 | 41.36 | 281.99 | 121.14 | |
| Insecta | | | | | | | | | | | | | |
| <i>Chironomidae</i> (CHI) | 36.16 | 27.87 | 30.76 | 41.11 | 24.69 | 18.52 | 1.76 | 2.02 | 3.57 | 8.80 | 0.00 | 2.02 | 3.45 |

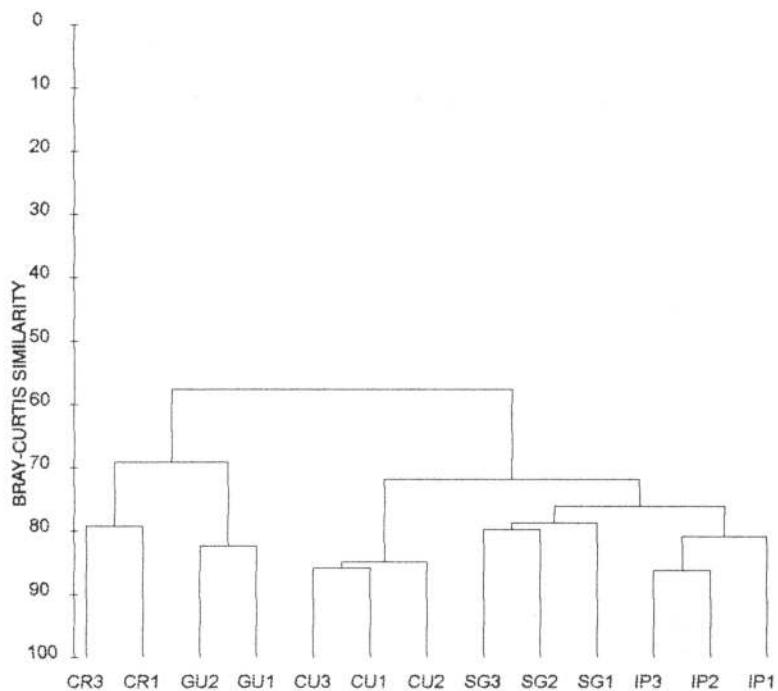


Figure 3. Dendrogram of similarity between the different stations in base to the annual average abundance of the species. (IP) Island of Las Palomas; (SG) San García; (CU) Cucareo; (GU) Guadarranque; (CR) Crinavis.

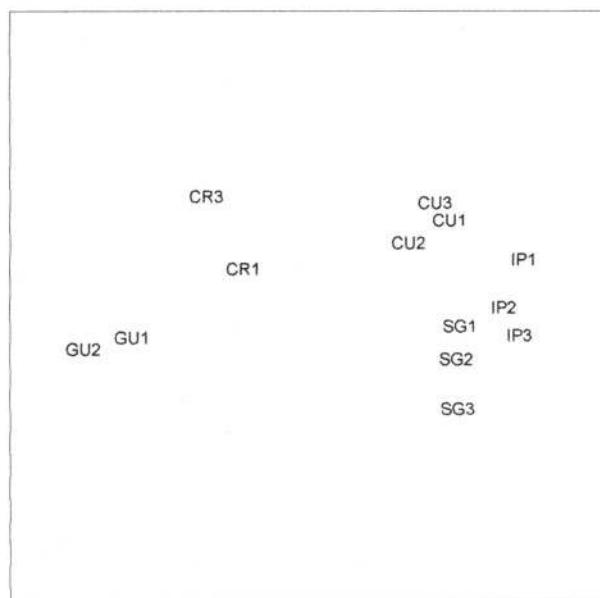


Figure 4. MDS ordination of the stations according to the annual average of abundance (Stress: 0,032). (IP) Island of Las Palomas; (SG) San García; (CU) Cucareo; (GU) Guadarranque; (CR) Crinavis.

have similar characteristics and are far from anthropogenic influence zones) and Cucareo (an area that suffers the influence of the nearby Port of Algeciras). Along the same line, the similarity of the stations within a specific area is greater among those located at more depth (IP2 and IP3; SG2 and SG3) than those more shallow (IP1 and SG1). In the case of the stations of Cucareo, these are found at a similar depth. With the two-dimensional representation of the ordination analysis MDS (Figure 4), the two groups and the different affinities among stations are confirmed.

With the SIMPER analysis it was possible to determine the species responsible for these groupings, those typical of a group as well as those discriminating among these groups. Among the typifying species of the external stations (Table 4), the isopod *Paranthura nigropunctata* stands out. Of relatively less importance, the amphipods *Microprotopus* sp., *Dexamine spinosa*, *Amphilochus neapolitanus*, *Stenothoe monoculoides* and *Aoridae* sp. and the pantopoda *Anoplodactylus angulatus*. At the internal stations (Table 5), the important species is the amphipod *Corophium acutum* and secondarily, the amphipods *Phtisica marina*, *Aoridae* sp. and *Aora spinicornis*, the tanaid *Leptochelia dubia* and the decapod *Macropodia* sp. The discriminating species (Table 6) are: *Amphilochus neapolitanus* and *Zeuxo normani*, as well as *Jassa marmorata* and *Microprotopus* sp.

In order to test the spatial differences, a one way ANOSIM was employed. In Table 7 the statistic values are shown at the spatial level and for each samplings. Generally, the differences are clear between the external and internal bay stations, which confirmed the results obtained by the previous

Table 4. Average abundance (\bar{Y}_A) of the most relevant species of the stations located in the external areas (Island of Las Palomas, San Garcia and Cucarao). Species are listed in decreasing order according to its contribution to the average of the similarity (\bar{S}_i) among stations until 75% of the accumulated total similarity ($\Sigma\bar{S}_i\%$). The total mean similarity among stations is 75.53%.

| Species | \bar{Y}_A | $SD(\bar{Y}_A)$ | \bar{S}_i | $SD(\bar{S}_i)$ | $\bar{S}_i/SD(\bar{S}_i)$ | $\bar{S}_i\%$ | $\Sigma\bar{S}_i\%$ |
|---------------------------------|-------------|-----------------|-------------|-----------------|---------------------------|---------------|---------------------|
| <i>Leptochelia dubia</i> | 1,165.14 | 574.31 | 3.0 | 0.32 | 9.17 | 3.91 | 3.91 |
| <i>Gammaropsis palmata</i> | 1,272.12 | 1,097.83 | 2.8 | 0.31 | 9.11 | 3.75 | 7.66 |
| <i>Stenothoe monoculoides</i> | 814.59 | 442.68 | 2.7 | 0.23 | 11.82 | 3.58 | 11.25 |
| Aoridae sp. | 847.77 | 498.58 | 2.7 | 0.27 | 10.01 | 3.54 | 14.78 |
| <i>Tanais dulongii</i> | 966.12 | 763.71 | 2.5 | 0.50 | 5.04 | 3.30 | 18.09 |
| <i>Corophium acutum</i> | 1,353.12 | 1,473.15 | 2.5 | 0.77 | 3.18 | 3.25 | 21.33 |
| <i>Apherusa bispinosa</i> | 492.49 | 243.87 | 2.4 | 0.25 | 9.78 | 3.21 | 24.55 |
| <i>Phtisica marina</i> | 566.92 | 299.50 | 2.4 | 0.34 | 7.07 | 3.18 | 27.72 |
| <i>Paranthura nigropunctata</i> | 374.49 | 69.92 | 2.4 | 0.09 | 28.68 | 3.17 | 30.90 |
| <i>Microprotopus</i> sp. | 291.71 | 76.69 | 2.2 | 0.11 | 19.87 | 2.94 | 33.83 |
| <i>Amphilochus neapolitanus</i> | 285.44 | 80.91 | 2.2 | 0.12 | 18.60 | 2.93 | 36.76 |
| <i>Amphithoe ramondi</i> | 427.75 | 313.63 | 2.2 | 0.41 | 5.28 | 2.86 | 39.62 |
| <i>Dexamine spinosa</i> | 265.84 | 99.47 | 2.1 | 0.11 | 19.52 | 2.83 | 42.45 |
| <i>Hyale schmidtii</i> | 560.92 | 679.77 | 2.1 | 0.39 | 5.42 | 2.81 | 45.25 |
| <i>Gammaropsis maculata</i> | 554.64 | 724.56 | 1.8 | 0.42 | 4.41 | 2.45 | 47.70 |
| <i>Achelia</i> sp. | 184.46 | 113.51 | 1.8 | 0.28 | 6.22 | 2.33 | 50.03 |
| <i>Nannastacus unglliculus</i> | 167.14 | 107.15 | 1.8 | 0.21 | 8.21 | 2.32 | 52.34 |
| <i>Callipallene tiberi</i> | 128.28 | 80.64 | 1.6 | 0.20 | 8.23 | 2.18 | 54.52 |
| <i>Aora spinicornis</i> | 145.13 | 113.46 | 1.6 | 0.23 | 7.08 | 2.15 | 56.67 |
| <i>Apherusa</i> sp. | 841.09 | 1,086.20 | 1.6 | 0.86 | 1.88 | 2.15 | 58.82 |
| <i>Cumella limicola</i> | 66.74 | 32.70 | 1.5 | 0.20 | 7.37 | 1.93 | 60.75 |
| <i>Hippolyte</i> sp. | 79.98 | 50.72 | 1.4 | 0.21 | 6.86 | 1.91 | 62.66 |
| <i>Dexamine spiniventris</i> | 53.84 | 40.79 | 1.3 | 0.18 | 7.48 | 1.75 | 64.41 |
| <i>Elasmopus</i> sp. | 59.97 | 45.53 | 1.3 | 0.24 | 5.45 | 1.73 | 66.14 |
| <i>Cymodoce</i> sp. | 40.49 | 22.95 | 1.2 | 0.17 | 7.30 | 1.64 | 67.78 |
| <i>Lembos websteri</i> | 59.32 | 70.98 | 1.2 | 0.24 | 4.81 | 1.54 | 69.32 |
| <i>Dynamene</i> sp. | 87.61 | 106.34 | 1.2 | 0.38 | 3.10 | 1.54 | 70.86 |
| <i>Perionotus testudo</i> | 43.43 | 41.25 | 1.1 | 0.18 | 6.17 | 1.49 | 72.36 |
| <i>Guernea coalita</i> | 121.45 | 168.07 | 1.1 | 0.44 | 2.51 | 1.46 | 73.82 |
| <i>Anoplodactylus angulatus</i> | 24.49 | 19.57 | 1.1 | 0.10 | 10.80 | 1.41 | 75.23 |

Table 5. Average abundance (\bar{Y}_B) of the most relevant species of the stations located in the internal areas (Guadarranque y Crinavis). Species are listed in decreasing order according to its contribution to the average of the similarity (\bar{S}_i) among stations until 78% of the accumulated total similarity ($\Sigma\bar{S}_i\%$). The total mean similarity among stations is 73%.

| Species | \bar{Y}_B | $SD(\bar{Y}_B)$ | \bar{S}_i | $SD(\bar{S}_i)$ | $\bar{S}_i/SD(\bar{S}_i)$ | $\bar{S}_i\%$ | $\Sigma\bar{S}_i\%$ |
|---------------------------------|-------------|-----------------|-------------|-----------------|---------------------------|---------------|---------------------|
| <i>Corophium acutum</i> | 4,186.67 | 869.27 | 4.6 | 0.14 | 33.00 | 6.32 | 6.32 |
| <i>Jassa marmorata</i> | 3,522.46 | 3,025.48 | 3.8 | 0.62 | 6.08 | 5.18 | 11.50 |
| <i>Phtisica marina</i> | 1,569.02 | 138.35 | 3.7 | 0.20 | 18.64 | 5.06 | 16.56 |
| Aoridae sp. | 1,739.83 | 588.56 | 3.6 | 0.28 | 12.74 | 4.91 | 21.46 |
| <i>Leptochelia dubia</i> | 882.79 | 180.77 | 3.1 | 0.24 | 13.03 | 4.29 | 25.75 |
| <i>Stenothoe monoculoides</i> | 1,597.17 | 1,387.98 | 3.0 | 0.60 | 5.06 | 4.17 | 29.92 |
| <i>Gammaropsis maculata</i> | 1,016.58 | 558.17 | 3.0 | 0.48 | 6.30 | 4.12 | 34.04 |
| <i>Aora spinicornis</i> | 674.97 | 443.18 | 2.7 | 0.16 | 16.92 | 3.69 | 37.73 |
| <i>Cumella limicola</i> | 623.88 | 526.77 | 2.5 | 0.39 | 6.38 | 3.37 | 41.10 |
| <i>Achelia</i> sp. | 379.27 | 139.91 | 2.4 | 0.27 | 9.11 | 3.35 | 44.44 |
| <i>Gammaropsis palmata</i> | 2,233.74 | 2,834.25 | 2.1 | 0.98 | 2.11 | 2.82 | 47.26 |
| <i>Erichtonius barsiliensis</i> | 1,055.42 | 1,131.73 | 1.9 | 1.34 | 1.39 | 2.54 | 49.81 |
| <i>Podocerus variegatus</i> | 169.46 | 125.84 | 1.8 | 0.33 | 5.49 | 2.46 | 52.27 |
| <i>Zeuxo normani</i> | 167.76 | 107.48 | 1.8 | 0.29 | 6.23 | 2.45 | 54.72 |
| <i>Lembos websteri</i> | 138.54 | 100.81 | 1.7 | 0.17 | 10.21 | 2.35 | 57.07 |
| <i>Microdeutopus chelifer</i> | 111.25 | 85.20 | 1.6 | 0.27 | 5.91 | 2.21 | 59.28 |
| <i>Callipallene tiberi</i> | 94.70 | 62.53 | 1.6 | 0.23 | 6.91 | 2.14 | 61.42 |
| <i>Dynamene</i> sp. | 158.38 | 170.38 | 1.6 | 0.29 | 5.34 | 2.14 | 63.57 |
| <i>Dexamine spinosa</i> | 72.72 | 44.58 | 1.5 | 0.24 | 6.03 | 2.01 | 65.58 |
| <i>Hippolyte</i> sp. | 102.50 | 128.67 | 1.4 | 0.15 | 9.32 | 1.86 | 67.43 |
| <i>Cymodoce</i> sp. | 150.56 | 153.57 | 1.3 | 0.54 | 2.238 | 1.75 | 69.18 |
| <i>Paranthura nigropunctata</i> | 135.83 | 149.58 | 1.3 | 0.42 | 3.01 | 1.73 | 70.91 |
| <i>Orchomenes</i> sp. | 42.71 | 31.69 | 1.2 | 0.26 | 4.64 | 1.68 | 72.59 |
| <i>Corophium sextonae</i> | 79.92 | 76.77 | 1.2 | 0.44 | 2.72 | 1.62 | 74.21 |
| <i>Achaeus</i> sp. | 20.15 | 7.96 | 1.2 | 0.11 | 10.23 | 1.60 | 75.81 |
| <i>Macropodia</i> sp. | 15.22 | 4.62 | 1.1 | 0.07 | 16.83 | 1.52 | 77.34 |

Table 6. Average abundance of the most relevant species of the stations located in the internal (\bar{Y}_i) and external areas (\bar{Y}_e). Species are listed in decreasing order according to its contribution to the average of the dissimilarity ($\bar{\delta}_i$) between the two groups until 40% of the accumulated total similarity ($\Sigma\bar{\delta}_i\%$). The total average dissimilarity among stations is 42.38%.

| Species | \bar{Y}_i | \bar{Y}_e | $\bar{\delta}_i$ | SD($\bar{\delta}_i$) | $\bar{\delta}_i/\text{SD}(\bar{\delta}_i)$ | $\bar{\delta}_i\%$ | $\Sigma\bar{\delta}_i\%$ |
|---------------------------------|-------------|-------------|------------------|------------------------|--|--------------------|--------------------------|
| <i>Jassa marmorata</i> | 3,522.46 | 5.57 | 2.00 | 0.56 | 3.58 | 4.71 | 4.71 |
| <i>Eriethonius brasiliensis</i> | 1,055.42 | 1.95 | 1.25 | 0.71 | 1.76 | 2.94 | 7.66 |
| <i>Hyale schmidtii</i> | 1.01 | 560.92 | 1.18 | 0.37 | 3.19 | 2.79 | 10.44 |
| <i>Amphilochus neapolitanus</i> | 0.00 | 285.44 | 1.18 | 0.08 | 14.35 | 2.79 | 13.23 |
| <i>Tanaid a dulongii</i> | 5.43 | 966.12 | 1.14 | 0.37 | 3.08 | 2.69 | 15.93 |
| <i>Apherusa</i> sp. | 0.81 | 841.09 | 1.11 | 0.66 | 1.69 | 2.62 | 18.55 |
| <i>Zeuxo normani</i> | 167.76 | 0.00 | 1.00 | 0.16 | 6.13 | 2.35 | 20.90 |
| <i>Microprotopus</i> sp. | 1.60 | 291.71 | 1.00 | 0.22 | 4.57 | 2.35 | 23.25 |
| <i>Zeuxo coralensis</i> | 395.62 | 0.00 | 0.95 | 0.68 | 1.40 | 2.24 | 25.49 |
| <i>Ampithoe ramondi</i> | 21.63 | 427.75 | 0.90 | 0.47 | 1.90 | 2.12 | 27.61 |
| <i>Caprella acanthifera</i> | 0.54 | 336.51 | 0.85 | 0.61 | 1.40 | 2.00 | 29.61 |
| <i>Apherusa bispinosa</i> | 43.62 | 492.49 | 0.80 | 0.43 | 1.86 | 1.88 | 31.49 |
| <i>Gammaropsis palmata</i> | 2,233.74 | 1,272.12 | 0.78 | 0.33 | 2.35 | 1.83 | 33.31 |
| <i>Guernea coalita</i> | 0.00 | 121.45 | 0.77 | 0.37 | 2.11 | 1.82 | 35.14 |
| <i>Corophium acutum</i> | 4,186.67 | 1,353.12 | 0.77 | 0.45 | 1.72 | 1.81 | 36.95 |
| <i>Corophium sextonae</i> | 79.92 | 3.57 | 0.61 | 0.32 | 1.91 | 1.45 | 38.40 |
| <i>Microdeutopus chelifer</i> | 111.25 | 26.81 | 0.59 | 0.39 | 1.51 | 1.40 | 39.80 |

Table 7. Values of the statistic R (A/OSIM) among the different stations during the period of study. P < 0.05; *: not meaningful. (I) Island of Las Palomas; (S) San García; (C) Cucareo; (G) Guadarranque; (CR) Crinavís.

| | SEP92 | DEC92 | MAR93 | JUN93 | SEP93 | | SEP92 | DEC92 | MAR93 | JUN93 | SEP93 |
|----------|-------|-------|-------|-------|-------|-----------|-------|-------|-------|-------|-------|
| (11-I2) | 0.28* | 0.00* | 0.82 | 0.24* | 0.46 | (S1-G2) | 0.94 | 1.00 | 0.96* | — | — |
| (11-I3) | 0.25* | 0.43 | 1.00 | 0.66 | 0.59 | (S1-CR1) | 1.00 | 1.00 | 0.68 | — | 1.00 |
| (11-S1) | 0.66 | 0.40 | 0.89 | — | 0.70 | (S1-CR3) | 0.98 | 1.00 | 0.73 | — | 1.00 |
| (11-S2) | 0.77 | 0.67 | 0.96 | — | — | (S2-S3) | 0.86 | 0.45* | — | — | — |
| (11-S3) | 0.90 | 0.70 | — | — | — | (S2-C1) | 1.00 | 0.69 | 0.76 | — | — |
| (11-C1) | 1.00 | 0.72 | 0.91 | 0.57 | 0.79 | (S2-C2) | 1.00 | 0.80 | 0.76 | — | — |
| (11-C2) | 1.00 | 0.70 | 1.00 | 0.66 | 0.84 | (S2-C3) | 1.00 | 0.80 | 0.90 | — | — |
| (11-C3) | 0.99 | 0.58 | 0.96 | 0.81 | 0.56 | (S2-G1) | 1.00 | 1.00 | 0.96 | — | — |
| (11-G1) | 1.00 | 0.99 | 1.00 | 1.00 | 1.00 | (S2-G2) | 0.95 | 1.00 | 1.00* | — | — |
| (11-G2) | 0.97 | 0.98 | 0.97 | 0.97 | — | (S2-CR1) | 1.00 | 1.00 | 0.92 | — | — |
| (11-CR1) | 1.00 | 0.96 | 1.00 | 1.00 | 1.00 | (S2-CR3) | 0.96 | 1.00 | 0.96 | — | — |
| (11-CR3) | 1.00 | 0.96 | 1.00 | 0.90* | 1.00 | (S3-C1) | 1.00 | 1.00 | — | — | — |
| (12-I3) | 0.28* | 0.25* | 0.49 | 0.29* | 0.18* | (S3-C2) | 1.00 | 0.98 | — | — | — |
| (12-S1) | 0.79 | 0.25* | 0.49 | — | 0.75 | (S3-C3) | 1.00 | 1.00 | — | — | — |
| (12-S2) | 0.80 | 0.49 | 0.84 | — | — | (S3-G1) | 1.00 | 1.00 | — | — | — |
| (12-S3) | 0.98 | 0.62 | — | — | — | (S3-G2) | 0.95 | 1.00 | — | — | — |
| (12-C1) | 1.00 | 0.62 | 0.51 | 0.74 | 0.83 | (S3-CR1) | 1.00 | 1.00 | — | — | — |
| (12-C2) | 1.00 | 0.64 | 0.68 | 0.68 | 0.93 | (S3-CR3) | 0.97 | 1.00 | — | — | — |
| (12-C3) | 1.00 | 0.63 | 0.80 | 0.80 | 0.70 | (C1-C2) | 0.77 | 0.79 | 0.21* | 0.08* | 0.96 |
| (12-G1) | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | (C1-C3) | 0.80 | 0.78 | 0.54* | 0.17* | 0.40 |
| (12-G2) | 0.96 | 1.00 | 0.97 | 0.99 | — | (C1-G1) | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| (12-CR1) | 1.00 | 0.96 | 1.00 | 1.00 | 1.00 | (C1-G2) | 0.96 | 1.00 | 0.95 | 0.98 | — |
| (12-CR3) | 0.98 | 0.96 | 1.00 | 0.90* | 1.00 | (C1-CR1) | 0.99 | 1.00 | 1.00 | 0.86 | 1.00 |
| (13-S1) | 0.98 | 0.42 | 0.43* | — | 0.84 | (C1-CR3) | 1.00 | 1.00 | 0.95 | 0.90* | 1.00 |
| (13-S2) | 0.89 | 0.84 | 0.77 | — | — | (C2-C3) | 0.27* | 0.61 | 0.55 | 0.35* | 0.56 |
| (13-S3) | 0.95 | 1.00 | — | — | — | (C2-G1) | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| (13-C1) | 1.00 | 1.00 | 0.93 | 0.87 | 1.00 | (C2-G2) | 0.96 | 1.00 | 0.95 | 0.97 | — |
| (13-C2) | 1.00 | 1.00 | 0.90 | 0.79 | 0.95 | (C2-CR1) | 1.00 | 1.00 | 1.00 | 0.90 | 1.00 |
| (13-C3) | 1.00 | 0.99 | 0.91 | 0.87 | 0.76 | (C2-CR3) | 0.96 | 1.00 | 0.95 | 0.90* | 1.00 |
| (13-G1) | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | (C3-G1) | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| (13-G2) | 0.96 | 1.00 | 0.97 | 0.99 | — | (C3-G2) | 0.94 | 1.00 | 0.96 | 0.96 | — |
| (13-CR1) | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | (C3-CR1) | 1.00 | 1.00 | 1.00 | 0.93 | 1.00 |
| (13-CR3) | 1.00 | 1.00 | 1.00 | 0.90* | 0.99 | (C3-CR3) | 0.94 | 1.00 | 1.00 | 0.90* | 0.99 |
| (S1-S2) | 0.87 | 0.36 | 0.04* | — | — | (G1-G2) | 0.01* | 0.81 | 0.90 | 0.77* | — |
| (S1-S3) | 0.98 | 0.57 | — | — | — | (G1-CR1) | 0.93 | 0.99 | 0.84 | 0.99 | 1.00 |
| (S1-C1) | 0.99 | 0.65 | 0.37 | — | 0.89 | (G1-CR3) | 0.95 | 0.98 | 0.99 | 0.90* | 0.97 |
| (S1-C2) | 1.00 | 0.62 | 0.46 | — | 0.84 | (G2-CR1) | 0.95 | 1.00 | 0.96 | 0.94 | — |
| (S1-C3) | 1.00 | 0.59 | 0.54 | — | 0.88 | (G2-CR3) | 0.92 | 1.00 | 0.92 | 0.94* | — |
| (S1-G1) | 1.00 | 1.00 | 0.95 | — | 1.00 | (CR1-CR3) | 0.31 | 0.69 | 0.61 | 0.70* | 0.71 |

Table 8. Results of the analysis BIO-ENV: variable combinations taking k every time, the higher values obtained in each combination being emphasized (in boldface the maximum value). (SED) sedimentation rate; (MSED) % organic matter of the sedimentation; (SUS) solid in suspension; (SUSM) organic matter in suspension; (HYD) hydrodynamism; (CI) compactness index; (INTV) interstitial volume.

| k | Harmonic Ranges of Spearman Correlation (p_w): Better Variable Combinations | | | | | | |
|---|---|---------|------------|---------|---------|---------|---------|
| | INTV | CI | HYD | SUS | SUSM | MSED | SED |
| 1 | (0.567) | (0.517) | (0.295) | (0.236) | (0.218) | (0.182) | (0.110) |
| 2 | HYD,INTV | HYD,CI | SUS,INTV | | | | |
| | (0.737) | (0.724) | (0.709) | | | | |
| 3 | HYD,CI,INTV | | SUS,HYD,CI | | | | |
| | (0.760) | | (0.696) | | | | |
| 4 | SUS,HYD,CI,INTV | | | | | | |
| | (0.749) | | | | | | |

ordination and classification analysis and this shows the discriminating power among localities that possess this animal group.

Relationships Between Biotic and Abiotic Variables

The environmental variables used for BIO-ENV were: sedimentation rate, % of organic matter of the sedimentation, hydrodynamism, solid and organic matter in suspension, interstitial volume and compactness index. These last two, considered as components of the habitat, together with the sedimentation rate were previously transformed ($\log x + 1$).

The better correlations obtained in the analysis are shown in Table 8. The maximum correlation is obtained for hydrodynamism, compactness and interstitial volume (0'760). Besides, good correlations of the combination of these parameters with the organic matter of the sedimentation and the solids in suspension are shown.

With the canonical correspondence analysis (CCA), it is possible to determine if these variables represent an environmental gradient throughout Algeciras Bay. In Figure 5 we observe that the stations of the Island Las Palomas are close to those of San García, while those of Cucareo are considerably more distant. Concerning the internal stations, Crinavis is relatively nearer to the external zone, due possibly to certain influences of the hydrodynamic forces (here, the tidal currents of the Strait converge during low tide). Table 9 shows that the gradient of the first axis of the analysis is practically determined by the compactness index. Hydrodynamism significantly influences distribution on the second axis, together with organic matter of the sedimentation.

With respect to the species composition, certain trends are indicated. In the area of Cucareo, there is a small group of species that are not abundant, such as *Caprella cf. mitis*, *Zenobiana prismatica* or *Pisa tetraedon*. But of more importance is the presence of typical soft-bottom species such as *Ampelisca cf. unidentata*, *Siphonoecetes* sp and *Apseudes talpa*; the motion of sands that are deposited on the algae making possible the settling of species of their characteristics.

Another noteworthy species is the amphipod *Apherusa* sp, characteristic of the shallow water stations of the external zones. It appears nearest to the stations of Cucareo. The zone

occupied by the stations of the Island Las Palomas and San García shows a numerous group of common species to both sites with a notable influence of compactness index (a larger number of spaces favors swimming species). Some examples are: *Cressa cristata*, *Guernea coalita*, *Eusiroides delavallei*, *Amphilochus neapolitanus*, *Hyale schmidtii*, *Apherusa bispinosa*, *Pereionotus testudo*, *Lembos viguieri*, *Pedoculina garcia-gomezi*, *Nebalia bipes*, etc. Another group, constituted by species living in moderate conditions of hydrodynamism, solids in suspension and compactness index are *Paranthura nigropunctata*, *Dexamine spinosa*, *Gammaropsis palmata*, *Thorlalus cranchii*, *Hippolyte* sp, *Leptochelia dubia*. Species with higher requirements of organic matter in suspension and low hydrodynamism are *Cymodoce* sp, *Tritraeta gibbosa*, *Corophium acutum*, *Aora spinicornis*, *Stenoiothe monocoloides* and *Orchomene* sp. Those of more hydrodynamism and a higher percentage of organic matter in the sediment are *Phtisica marina*, *Pariambus typicus*, *Ischyrocerus inexpectatus*, *Gammaropsis maculata*, etc. In the Crinavis area, the same as in Cucareo, there is a high number of typical soft-bottom species due to the fact that the algae are surrounded by a wide sandy bottom. Thus, *Corophium sextonae*, *Gammarella fucicola*, *Harpinia* sp, *Iphinoe* sp, *Dardanus* sp, etc are found. In the Guadarranque zone there is a predominance of deposit-feeder species such as *Jassa marmorata*, *Erithonius brasiliensis* and *Zeuxo normani* and *Z. coralensis*, characterized by the highest values of organic matter in the sedimentation and with little interfrond spaces.

According to the Monte-Carlo test, ordination is significant for the first axis ($F = 3'51$, $P = 0'05$).

DISCUSSION AND CONCLUSIONS

The determining factors of the composition of the epiphytic arthropods community are hydrodynamism, sedimentation rate, solids in suspension, levels of nutrients and algal morphology (HAGERMAN, 1966; BELLAN-SANTINI, 1969; MOORE, 1972, 1986; RUSSO, 1989; etc). With this animal group there is a clear separation of the different species according to the prevailing environmental conditions in each zone. According to FENWICK (1976), the variety of microhabitats is larger in exposed coasts than in protected, thus a greater number of species appear in the external zone. The quantitative dominances of external-zone species are more spread out than among the species of the internal zone where there are some species that are clearly dominant over the others, as in the case of the amphipods *Corophium acutum* and *Jassa marmorata*. Both species, even though they can also be abundant in high hydrodynamism conditions (DIVIACCO, 1980; DIVIACCO and ARATA, 1988; JIMENO, 1993), proliferate more in this zone because of the high available sediment on the algae for the construction of their tubes and as food, being two species typical of port environments (TARAMELLI and SCIPIONE, 1977; CANDELA *et al.*, 1983; PROCCACCINI and SCIPIONE, 1992). *J. marmorata* was also found in areas with high stress (LEDOYER and MENIOIU, 1983) and of high sedimentation (SCIPIONE *et al.*, 1981; FRANZ, 1989; MYERS, 1989). *C. acutum* was also abundant in the external zone, which confirms its resistance to strong hydrodynamism.

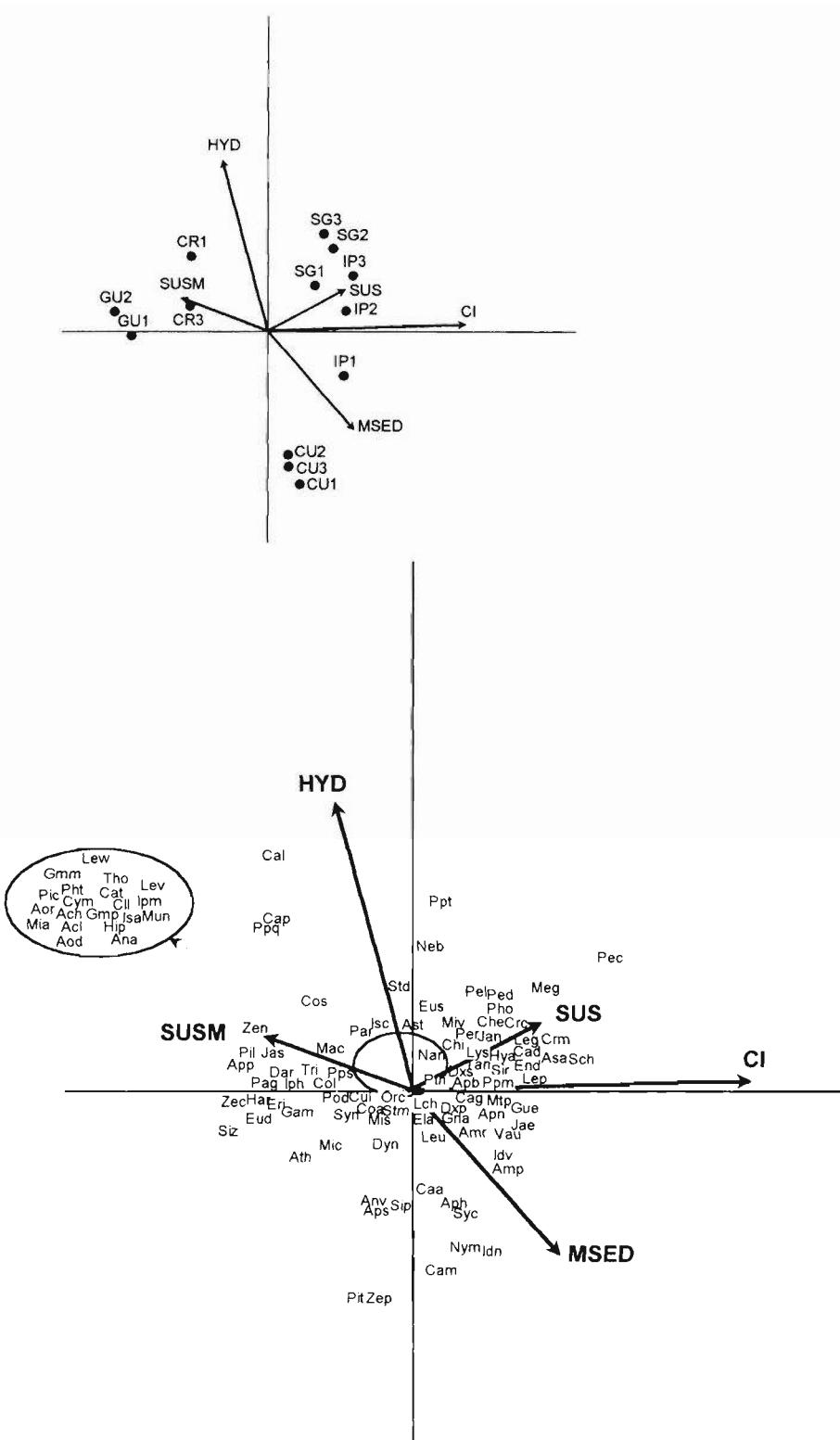


Figure 5. Graph representation of the stations and species with respect to the first two axes of the canonical correspondence analysis (CCA). (HYD) hydrodynamicism; (SUS) solid in suspension; (SUSM) organic matter in suspension; (MSED) organic matter of the sedimentation; (CI) compactness. (Species key in Table 3).

Table 9. Results of the canonical correspondence analysis.

| Intraset Values | Axis1 | Axis2 | Inflation |
|--|--------|-------|-----------|
| MSED | 0.442 | 0.466 | 1.3426 |
| SUS | 0.372 | 0.197 | 1.6461 |
| SUSM | -0.423 | 0.158 | 1.4838 |
| HYD | -0.219 | 0.819 | 1.1641 |
| CI | 0.966 | 0.028 | 1.7876 |
| | Axis1 | Axis2 | |
| Species-environment correl. | 0.905 | 0.914 | |
| Cumulative % variance of species data of species-environment relation | 33.4 | 45.7 | |
| | 51.1 | 69.9 | |

Another reason to explain the preference of this species for the internal stations can be because of the existence of competition. Some spatio-temporal differences in the distribution or abundance of species with apparent similar food or habitat requirements suggest competition, present or past, due to limited resources (DIAMOND, 1975). An important species in the external zone and practically non-existent in the internal is the amphipod *Ampithoe ramondi*. It is often a territorial species (BRAWLEY and ADEY, 1981), builder of tubes and of aggressive behavior compelling other tubicolous species such as *Corophium* to occupy the distal and basal portions of the algae (GUNNILL, 1984). It is a species of relative large size and whose preference in our study was for algae with a lot of space and larger size. *C. acutum* is located in opposite conditions, together with other tubicolous species such as *Ericthonius brasiliensis*.

There are a great number of species under the same conditions as those of *A. ramondi*, with moderate and high hydrodynamism and a lot of space which should also favor swimming species (CRISP and MWAISEJE, 1989) such as *Apherusa bispinosa*, *Apherusa* sp., *Lembos vigueri*, *Amphilochus neapolitanus* or *Hyale schmidtii* (this species sustains strong mechanical stress according to KRAPF-SCHICKEL, 1975). In intermediate conditions there are species such as the isopod *Paranthura nigropunctata*, abundant on macroalgae and seagrass and regularly associated with the sediment fraction (WAGELE, 1982; ARRONTES and ANADON, 1990); the decapods *Hippolyte* sp. and *Thoralus cranchii* are both swimming species and with need of space; etc. Between the species with requirements of low hydrodynamism and high percentages of organic matter in suspension we find the isopod *Dynamene* sp., which tends to be tolerant of pollution (BELLAN-SANTINI, 1969; DESROSIERS *et al.*, 1986), and the amphipod *Stenothoe monoculoides*, that seems to be indifferent to hydrodynamism (DOMMASNES, 1968; MOORE, 1986). In zones with higher incidence of currents but smaller suspension, for example, the amphipod *Pariambus typicus* stands out, in spite of being a typical soft bottoms species and sheltered areas as the *Caulerpa* and fanerogames meadows (SCIPIONE and CHESSA, 1986) with tolerance for high sedimentation rates (PROCCACCINI and SCIPIONE, 1992), and *Ischyrocerus inexpectatus* like most of the species of the Ischyroceridae family adjust well to high hydrodynamism (MOORE, 1973;

VADER, 1983). At the opposite extreme, in conditions of smaller compactness index and more protected areas, the typical deposit-feeder species appear as mentioned above, *Jassa marmorata* and *Ericthonius brasiliensis* (TARAMELLI and PEZZOLI, 1986) or the tanaids *Zeuxo normani* and *Z. coralensis*, characteristic of shallow waters and higher temperatures (SANZ, 1993).

In conclusion, the different environmental conditions from each zone of the Bay of Algeciras determine the final composition of the community. Therefore, we can assert that this taxonomic group, thanks to its high richness in species with a wide environmental spectrum and its great quantitative importance in the epiphytic communities shows in a clear way the spatio-temporal differences among the different localities. Thus, it can be extremely useful in characterization studies of environmental quality of coastal waters.

ACKNOWLEDGEMENT

Appreciation to CEPSA (Compañía Española de Petróleos, S.A.), Sevillana de Electricidad, Excmo. Ayuntamiento de los Barrios, Mancomunidad de Municipios del Campo de Gibraltar for financial support of this work.

LITERATURE CITED

- AOKI, M., 1988. Factor affecting population fluctuations of caprellid amphipods inhabiting *Sargassum patens* bed (Preliminary report). *Bulletin of the Japanese Association of Benthology*, 32, 42-49.
- AOKI, M. and KIKUCHI, T., 1990. Habitat adaptations of caprellid amphipods and the importance of epiphytic secondary habitats in a *Sargassum patens* bed in Amakusa, southern Japan. *Publications Amakusa Marine Biology Laboratory*, 10(2), 123-133.
- ARRONTES, J. and ANADON, R., 1990. Distribution of intertidal isopods in relation to geographical changes in macroalgal cover in the Bay of Biscay. *Journal of Marine Biological Association of UK*, 70, 283-293.
- BAILEY-BROCK, J., 1979. Sediment trapping by chaetopterid polychaetes on a Hawaiian fringing reef. *Journal of Marine Research (USA)*, 37(4), 643-656.
- BELLAN-SANTINI, D., 1964. Influence de la pollution sur quelques peuplements superficiels de substrat rocheux. *Comm. Int. Explor. Sci. Mer. Médit. Symp. Pollut. Mar. Par. Microorgan. Prod. Petrol.*, pp. 127-131.
- BELLAN-SANTINI, D., 1966. Influence des eaux polluées sur la flore et la faune marines benthiques dans la région marseillaise. *Techniques et Sciences Municipales*, 61 Année, 7, 285-292.
- BELLAN-SANTINI, D., 1969. Contribution à l'étude des peuplements infralittoraux sur substrats rocheux (étude qualitative et quantitative de la frange supérieure). *Reel. Travail Station marine d'Endoume*, 47(63), 5-294.
- BIBILONI, M.A., 1981. Estudio faunístico del litoral de Blanes: V. Sistemática de moluscos y artrópodos (crustáceos y piezognóstidos). *Miscellánea Zoológica*, vol. VII, pp. 43-52.
- BRAWLEY, S.H. and ADEY, W.H., 1981. The effect of micrograzers on algal community structure in a coral reef microcosm. *Marine Biology*, 61, 167-177.
- BRAY, J.R. and CURTIS, J.T., 1957. An ordination of the upland forest communities of Southern Wisconsin. *Ecological Monograph*, 27, 325-349.
- CANDELA, A.; SCONFIELLI, R., and TORELLI, A., 1983. Ricerche sperimentali sulla dinamica stagionale delle zoocenosi intermareali della Laguna di Venezia. *Bollettino Museo Civico di Storia Naturale Verona*, 33, 23-72.
- CASTELLO, J.; PORTAS, F., and ISERN-ARUS, J., 1987. Contribución al conocimiento de los Crustáceos Decápodos algicolas de las islas Baleares. *Investigaciones Pesqueras*, 51 (supl. 1), 293-300.

- CLARKE, K.R. and AINSWORTH, M., 1993. A method of linking multivariate community structure to environmental variables. *Marine Ecology Progress Series*, 92, 205–219.
- CLARKE, K.R. and GREEN, R.H., 1988. Statistical design and analysis for a "biological effects" study. *Marine Ecology Progress Series*, 46, 213–226.
- COEN, L.D.; HECK, K.L. JR., and ABELE, L.G., 1981. Experiments on competition and predation among shrimps of seagrasses meadows. *Ecology*, 62(6), 1484–1493.
- CONRADI, M., 1995. Distribución espaciotemporal de los peracáridos (Crustacea) asociados a *Bugula neritina* (L., 1758) en la Bahía de Algeciras. Aspectos faunísticos y zoogeográficos. Doctoral Thesis. University of Cádiz, 210p.
- COSTELLO, M.J. and MYERS, A.A., 1987. Amphipod fauna of the sponges *Halichondria panicea* and *Hymeniacidon perleve* in Lough Hyne, Ireland. *Marine Ecology Progress Series*, 41, 115–121.
- CRISP, D.J. and MWAISEJJE, B., 1989. Diversity in intertidal communities with special reference to the *Corallina officinalis* community. In: ROS, J.D. (ed.), *TOPICS IN MARINE BIOLOGY. SCIENTIA MARINA*, 53(2-3), 365–372.
- DAHL, E., 1948. On the smaller Arthropoda of marine algae, especially in the polyhaline waters off the Swedish west coast. Dissertation, Lund (Undersökn. över Öresund, 35), 193p.
- DESRÖSIERS, G.; BELLAN-SANTINI, D., and BRETHES, J.C., 1982. Evolution spatio-temporelle des peuplements de substrats rocheux superficiels dans un golfe soumis à de multiple pollutions (Golfe de Fos, France). *Tethys*, 10, 245–253.
- DESRÖSIERS, G.; BELLAN-SANTINI, D., and BRETHES, J.C., 1986. Organisation trophique de quatre peuplements de substrats rocheux selon un gradient de pollution industrielle (Golfe de Fos, France). *Marine Biology*, 91, 107–120.
- DESRÖSIERS, G.; BELLAN-SANTINI, D.; BRETHES, J.C., and WILLEIE, A., 1990. Variability in trophic dominance of crustaceans along a gradient of urban and industrial contamination. *Marine Biology*, 105, 137–143.
- DIAMOND, J.M., 1975. Assembly of species communities. In: CODY, M.L. and DIAMOND, J.M., (eds.), *Ecology and Evolution of Communities*. Cambridge, Massachusetts: Belknap pp. 342–444.
- DIVIACCO, G., 1980. Ecologia e distribuzione dei crostacei Anfipodi nella laguna di Orbetello. *Bollettino Museo Civico di Storia Naturale Verona*, 7, 303–317.
- DIVIACCO, G., and ARATA, P., 1988. Inventory of the amphipod crustaceans of the Italian Peninsula coastal lagoons. *Rapp. Proc. verb. Comm. int. Mer Médit.*, 3(2), 53.
- DOMMNASNE, A., 1968. Variations in the meiofauna of *Corallina officinalis* L. with wave exposure. *Sarsia*, 34, 117–124.
- DOMMNASNE, A., 1969. On the fauna of *Corallina officinalis* L. in western Norway. *Sarsia*, 38, 71–86.
- EDGAR, G.J., 1990a. The influence of plant structure on the species richness, biomass and secondary production of macrofaunal assemblages associated with Western Australian seagrass beds. *Journal of Experimental Marine Biology and Ecology*, 137, 215–240.
- EDGAR, G.J., 1990b. Predator-prey interactions in seagrass beds. I. The influence of macrofaunal abundance and size-structure on the diet and growth of the western rock lobster *Panulirus cygnus* George. *Journal of Experimental Marine Biology and Ecology*, 139, 1–22.
- FENWICK, G.D., 1976. The effect of wave exposure on the amphipod fauna of the alga *Caulerpa brownii*. *Journal of Experimental Marine Biology and Ecology*, 125, 117–136.
- FRANZ, D.R., 1989. Population density and demography of a fouling community amphipod. *Journal of Experimental Marine Biology and Ecology*, 125, 117–136.
- GAMBI, M.C.; BUJA, M.C.; CASOLA, E., and SCARDI, M., 1989. Estimates of water movement in *Posidonia oceanica* beds: a first approach. *International Workshop on Posidonia Beds*, 2, 101–112.
- GARCIA, J., 1986. Variabilidad del nivel del mar en el Estrecho de Gibraltar. Mareas y oscilaciones residuales. Doctoral Thesis, University of Málaga.
- GARCIA-RASO, J.E., 1988. Consideraciones generales sobre la taxocenosis de crustáceos decápodos de fondos de concrecionamiento calcáreo superficial del alga *Mesophyllum lichenoides* (Ellis and Sol.) Lemoine (Corallinaceae) del mar de Alborán. *Investigaciones Pesqueras*, 52(2), 245–264.
- GRAHAME, J. and HANNA, F.S., 1989. Factors affecting the distribution of the epiphytic fauna of *Corallina officinalis* (L.) on an exposed rocky shore. *Ophelia*, 30(2), 113–129.
- GUÑILU, F.C., 1984. Differing distributions of potentially competing amphipods, copepods and gastropods among specimens of the intertidal alga *Pelvetia fastigiata*. *Marine Biology*, 82, 277–291.
- HACKER, S.D. and STENECK, R.S., 1990. Habitat architecture and the abundance and body-size-dependent habitat selection of a phytal amphipod. *Ecology*, 71(6), 2269–2285.
- HAGERMAN, L., 1966. The macro- and microfauna associated with *Fucus serratus* L. with some ecological remarks. *Ophelia*, 3, 1–43.
- HOLMLUND, M.B.; PETERSON, C.H., and HAY, M.E., 1990. Does algal morphology affect amphipod susceptibility to fish predation? *Journal of Experimental Marine Biology and Ecology*, 139, 65–83.
- JIMENO, A., 1993. Contribución al estudio de los Anfípodos de las costas mediterráneas catalanas. Estudio faunístico, ecológico, biológico y biogeográfico. Doctoral Thesis. Univ. Barcelona, 573p.
- KRAPP-SCHICKEL, G., 1975. Revision of Mediterranean *Leucothoe* species (Crustacea, Amphipoda). *Bollettino Museo Civico di Storia Naturale Verona*, 11, 91–118.
- KRUSKAL, J.B. and WISH, M., 1978. *Multidimensional Scaling*. Beverley Hills, California: Sage Publications.
- LEDROYER, M. and MENIOU, M., 1983. Considerations sur la répartition du Gammarien (Crustacea: Amphipoda) *Jassa falcata* (Montagu, 1808). *Bulletin Institute Scientifique. Rabat*, 7, 93–114.
- MOORE, P.G., 1972. The kelp fauna of Northeast Britain. I. Introduction and the physical environment. *Journal of Experimental Marine Biology and Ecology*, 13, 97–125.
- MOORE, P.G., 1973. The kelp fauna of Northeast Britain. II. Multivariate classification: turbidity as an ecological factor. *Journal of Experimental Marine Biology and Ecology*, 13, 127–163.
- MOORE, P.G., 1986. Levels of heterogeneity and the amphipod fauna of kelp holdfasts. In: MOORE and SEED (eds.), *The Ecology of Rocky Coasts*, cap. XX, 274–289.
- MUTUS, B.J., 1968. A field method for measuring "exposure" by means of plaster balls. *Sarsia*, 34, 61–68.
- MYERS, A.A., 1989. The family Isaeridae. In: RUFFO, S. (ed.), *The Amphipoda of the Mediterranean. Part II. Gammaridea (Haustoriidae to Lysianassidae)*. Mémoires Institute Océanographique. Monaco, 13, 395–431.
- NELSON, W.G., 1979a. Experimental studies of selective predation on amphipods: consequences for amphipod distribution and abundance. *Journal of Experimental Marine Biology and Ecology*, 38, 225–245.
- NELSON, W.G., 1979b. An analysis of structural pattern in an eelgrass (*Zostera marina* L.) amphipod community. *Journal of Experimental Marine Biology and Ecology*, 39, 231–264.
- POORE, A.G.B., 1994. Selective herbivory by amphipods inhabiting the brown alga *Zonaria angustata*. *Marine Ecology Progress Series*, 107, 113–123.
- PROCACCINI, G. and SCIOPONE, M.B., 1992. Observations on the spatio-temporal distribution of crustaceans amphipods in the Fusaro coastal lagoon (Central Tyrrhenian Sea, Italy) and some notes on their presence in Mediterranean lagoons. *P.S.Z.N.I.: Marine Ecology*, 13(3), 203–224.
- RUSSO, A.R., 1989. Fluctuations of epiphytic gammaridean amphipods and their seaweed hosts on an Hawaiian algal reef. *Crustaceana*, 57(1), 25–37.
- SÁNCHEZ-MOYANO, J.E., 1996. Variación espacio-temporal en la composición de las comunidades animales asociadas a macroalgas como respuestas a cambios en el medio. Implicaciones en la caracterización ambiental de las áreas costeras. Doctoral Thesis, University of Sevilla, 407p.
- SÁNZ, C., 1993. *Contribución al estudio de los Crustáceos litorales del Mediterráneo Occidental*. Doctoral Thesis. University of Barcelona, 688p.
- SCHNEIDER, F.J. and MANN, K.H., 1991. Species specific relationships of invertebrates to vegetation in a seagrass bed. II. Experiments on the importance of macrophyte shape, epiphyte cover and

- predation. *Journal of Experimental Marine Biology and Ecology*, 145, 119–139.
- SCIPIONE, M.B.; CARNEVALE, G.; FINELLI, F.; FRESI, E.; MAZZELLA, L.; PONTICELLI, M.P., and TARAMELLI, E., 1981. Ricerche sui popolamenti bentonici di substrato duro del porto d'Ischia. Infralitorale fotofilo. III. Anfipodi. *Quaderno Laboratorio Tecnologia Pesca*, 3(1), 505–517.
- SCIPIONE, M.B. and CHESSA, L.A., 1986. The benthic fauna of Palau Bay (North-East Sardinia): Amphipoda. *Rapp. Comm. int. Mer Médit.*, 30(2), 9.
- SNEATH, P.H.A. and SOKAL, R.R., 1973. *Numerical Taxonomy*. San Francisco: Freeman.
- STONER, A.W., 1983. Distributional ecology of amphipods and tanaidaceans associated with three seagrass species. *Journal of Crustacean Biology*, 3(4), 505–518.
- STRICKLAND, J.D.H. and PARSON, T.R., 1969. A practical handbook of sea water analysis. *Fisheries Research Board, Canadian Bulletin*, 167.
- TARAMELLI, E. and PEZZOLI, 1986. Ecologia e distribuzione dei crostacei Anfipodi nei laghi costei della provincia di Latina. *Bullettino Museo Civico di Storia Naturale Verona*, 13, 295–318.
- TARAMELLI, E. and SCIPIONE, M.B., 1977. Gli Anfipodi del porto di Civitavecchia. In: CINELLI, F.; FRESI, E., and MAZZELLA, L. (eds.), *Atti IX Congr. S.I.B.M., Lacco Ameno d'Ischia*, 427–438.
- TER BRAAK, C.J.F., 1986. Canonical correspondence analysis: a new eigenvector technique for multivariate direct gradient analysis. *Ecology*, 67(5), 1167–1179.
- TER BRAAK, C.J.F., 1990. Interpreting canonical correlation analysis through biplots of structure correlations and weights. *Psychometrika*, 55(3), 519–531.
- VADER, W., 1983. Prehensile pereapods in Gammaridean amphipoda. *Sarsia*, 68, 134–148.
- WAGELE, J.W., 1982. Redescription and comparison of the Mediterranean species of *Paranthura* (Crustacea, Isopoda, Anthuridea). *P.S.Z.N.I.: Marine Ecology*, 3, 109–132.

[] SUMMARY []

Se estudia la variación espacial de las comunidades de artrópodos asociadas al alga *Halopteris scoparia* en relación a la influencia de factores ambientales en la Bahía de Algeciras (Sur de España). Por medio de diversos análisis multivariantes, podemos constatar una clara diferencia en la composición de las comunidades de las áreas externas e internas de la bahía, siendo el hidrodinamismo y la morfología algal los principales factores determinantes de esta composición. En la zona externa se aprecia un mayor número de especies cuyas dominancias cuantitativas están más repartidas que entre las especies de la zona interna donde hay algunas que dominan claramente sobre las demás como es el caso de los anfípodos *Corophium acutum* y *Jassa marmorata*. Como conclusión, podemos afirmar que este grupo animal, con una elevada riqueza de especies con un amplio espectro ambiental y con una gran importancia cuantitativa en las comunidades epifitas, muestra claramente las diferencias entre las distintas localidades, por lo que pueden ser de gran utilidad en estudios de caracterización de la calidad ambiental de las aguas litorales.