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1	The bioelements, the elementome and the "biogeochemical niche".
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32 Abstract

Every living creature on Earth is made of atoms of the various bioelements that are harnessed in 33 34 the construction of molecules, tissues, organisms and communities, as we know them. Organisms need these bioelements in specific quantities and proportions to survive and grow. Distinct 35 species have different functions and life strategies, and have therefore developed distinct 36 37 structures and adopted a certain combination of metabolic and physiological processes. Each species is thus also expected to have different requirements for each bioelement. We therefore 38 propose that a "biogeochemical niche" can be associated with the classical ecological niche of 39 40 each species. We show from field data examples that a biogeochemical niche is characterized by a particular elementome defined as the content of all (or at least most) bioelements. The 41 42 differences in elementome among species are a function of taxonomy and phylogenetic distance, 43 sympatry (the bioelemental compositions should differ more among coexisting than among noncoexisting species to avoid competitive pressure), and homeostasis with a continuum between 44 45 high homeostasis/low plasticity and low homeostasis/high plasticity. This proposed biogeochemical niche hypothesis has the advantage relative to other associated theoretical niche 46

47	hypotheses that it can be easily characterized by actual quantification of a measurable trait: the
48	elementome of a given organism or a community, being potentially applicable across taxa and
49	habitats. The changes in bioelemental availability can determine genotypic selection and
50	therefore have a feedback on ecosystem function and organization, and, at the end, become
51	another driving factor of the evolution of life and the environment.
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56	1. The elements of life and the construction of molecules, tissues, organisms and
57	communities
57 58	communities All living beings are made of atoms of various bioelements (elements used by living organisms).
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68	These six bioelements, however, are not the only bioelements that have important roles in
69	organisms (Slade, 2006). Other bioelements, normally present in low concentrations such as
70	potassium (K), magnesium (Mg), iron (Fe), calcium (Ca), molybdenum (Mo), manganese (Mn)
71	and zinc (Zn) have functions that are more specific. They are cofactors in catalytic reactions in
72	several biochemical and physiological processes. For example, in plants, K, Mg, Fe, S, Ca, and
73	Zn are involved in photosynthesis, respiration, regulation of ion balance in chloroplasts and
74	vacuoles, transport of sugar into the phloem, secondary metabolism, maintenance of internal
75	cellular osmotic or pH equilibria and balances, and strategies for stomatal control and avoidance
76	of water stress (Knight et al. 1991, Bourgis et al. 1999, Shaul 2002, Tripler et al. 2006, Peñuelas
77	and Sardans 2009, Baribault et al. 2012, Naples and Fisk 2010, Rivas-Ubach et al. 2012,
78	Peñuelas et al. 2013, Sardans and Peñuelas 2015). All these bioelements thus contribute
79	essentially to organismic functions, albeit to different degrees (Figure 1).
80	The concentrations and ratios of C, N and P have been the basis of ecological
81	stoichiometry (Elser et al. 2000a,b, 2001, Sardans et al. 2015, 2016). However, to link
82	bioelemental composition and stoichiometry with ecological traits in all types of organisms and
83	communities will thus require information about bioelemental composition beyond the C:N:P
84	concentrations and ratios (Peñuelas and Sardans 2009). The other bioelements such as S, K, Ca
85	or Mg are also crucial and can be limiting in some ecosystems (Naples and Fisk 2010, Baribault
86	et al. 2012, Lapenis et al. 2013). These other bioelements have been introduced in ecological
87	stoichiometric studies to link the elemental composition of individuals and species with their
88	ecological traits, such as drought resistance (related to K), light environment (related to Mg) or

Sardans et al., 2015). All these bioelements should thus be considered in ecologicalstoichiometric studies.

92

93 2. Elementome and biogeochemical niche.

94 Species use these bioelements to different degrees based on their ecological strategy (example 95 for plants in Figure 2). Classic Hutchinson's (1957) description of species niche has been largely applied to the main taxa from a point-of-view of distinct species in a site position in a "n-96 97 dimensional hypervolume" formed by the values of determined environmental variables and/or organisms traits. For instance, plant species have been classified in a hypervolume formed by 98 three axes, i.e. corresponding to Grime's 3 strategies: competitor, stress-tolerant and ruderal 99 100 (Grime 1977). In animals, the species-specific niche has been determined and located in function of hypervolumes for n-axis of different values of energy expenditure (Wilson et al., 101 2011), distinct diet sources (Araujo et al., 2007; Bearhop et al., 2004) or different elemental 102 composition (González et al., 2017, 2018). In microbial communities, genotype-niches have also 103 been described in function of the different availabilities of different nutrients (Pereira and Berry, 104 105 2017) or as a function of the values of diverse functional traits (Lennon et al., 2012) thus 106 describing multivariant spaces with a determined combination of nutrients availability or 107 functional traits where a determined genotype can exist. Any living species should thus tend to 108 optimize its overall function within a determined and space within this volume generated by a combination of these axes describing strategies or biotic and/or abiotic variables while avoiding 109 the overlap with other coexisting species. Following these niche approach, we can hypothesize 110 that distinct species have different functions and life strategies and have therefore developed 111 distinct structures and adopted a certain combination of metabolic and physiological processes. 112

113	Therefore, each species (of all taxa types) also has different requirements for each bioelement,
114	and should thus tend to have its own elemental composition (Yu et al. 2011; Kattge et al. 2011;
115	Sardans and Peñuelas 2013, 2014; Sardans et al. 2015; Pierce et al. 2016; Bartrons et al., 2018).
116	We postulate that the more functionally separated two species are, the more different the
117	content of bioelements should be. This led us to propose a "biogeochemical niche" (BN) defined
118	as the multidimensional space defined by the concentrations of bioelements in individuals of a
119	given species (Peñuelas et al., 2008). This BN has been used in the recent years by other authors
120	such as Gonzalez et al. (2017) who proposed the "multidimensional stoichiometric niche" as the
121	specific position of different taxonomic/trophic groups in the three-dimensional space
122	determined by C, N, and P, and other elements as additional axes in their application of the BN
123	to animals (Gonzalez et al., 2018), as previously proposed for plants by Peñuelas et al. (2008),
124	and for plants and animals together by Bartrons et al (2018). BN is thus based on the species-
125	specific needs and use of bioelements in different amounts and proportions resulting from the
126	organismic structure and function in the given abiotic and biotic environmental space where the
127	species has evolved (Peñuelas et al. 2008, 2010). Each species should thus have an optimal (at
128	maximum fitness) equilibrium of bioelemental composition, resulting from its functional and
129	structural adaptations, equivalent to the optimum "elemental phenotype" (Jeyasingh et al., 2014).
130	The BN represents thus the species-specific strategies of growth and resource uptake and the
131	differences in soil-space occupation.

We propose that the processes underlying this species-specific elementome and the BN are based on three basic complementary rules (Figures 3). First, distinct taxonomic groups exhibit different elementomes, with larger differences as taxonomic distance and evolutionary time increase. Second, at equilibrium, coexisting species tend to have distinct elementomes to

136 minimize competitive pressure. Third, trade-offs between adaptation to being competitive in a stable environment versus being successful in more fluctuating environments cause homeostasis 137 and plasticity differences between species in a continuum of strategies. The BN distances among 138 139 species should thus be a function of taxonomic difference, sympatry (coexistence), and homeostasis/plasticity and sympatry (Figure 3). The BN of taxa i is assumed to be the result of 140 141 its taxonomical evolutionary determination and its capacity to respond to changes in external conditions, which is also partly taxonomically-determined. External changes include nutrient 142 availability (including the level of atmospheric CO₂), changes in growing season duration, 143 144 disturbance regimes, and, on longer time scales, changes in soil properties. This capacity to respond to environmental changes is, in turn, a function of taxonomy, sympatry, and 145 146 homeostasis:

147
$$BN_{i} = BN \text{ of } taxa_{i} = Elementome_{i1-n} = (C: N: P: K: S: Mg: Ca: ...)_{i1-n} = BNtax_{i} +$$

148
$$\frac{\partial BNtax_{i}}{\partial Env} = BNtax_{i} + f(Tax_{i}, Symp_{i}, Hom_{i})$$
(Eq. 1)

149 where:

BNtax_i = Biogeochemical niche explained by taxonomy of taxa i (measured in n
individuals of taxa i). A n-dimensional hypervolume describes the fundamental BN of a species.
The domain of this hypervolume function is restricted to a finite region corresponding to the
possible values of bioelemental combinations.

154 $\frac{\partial BNtax_i}{\partial Env}$ = Changes in BNtax_i of taxa i explained by environmental changes 155 tax_i = taxonomy of taxa i; Symp_i sympatry of taxa i with other competing taxa (j), and 156 Hom_i = elemental homeostasis-plasticity of the taxa i

157
$$BN_i = BN \tan_{i=1} Elementom_{i_{1-n}} = (C: N: P: K: S: Mg: Ca...)_i$$

158
$$BN_j = BNtax_j + \frac{\partial BNtax_j}{\partial Env} = BNtax_{j1-n} + f(Tax_j, Symp_j, Hom_j)$$
 (Eq. 2)

159 The BN distance between taxa i and j, ΔBN_{i-j} , will be:

160
$$\Delta BN_{i-j} = (BNtax_i - BNtax_j) + \frac{\partial BN_{taxi} - \partial BN_{taxj}}{\partial Env} = (BNtax_i - BNtax_j) +$$

161
$$[f(Tax_i, Symp_i, Hom_i) - f(Tax_j, Symp_j, Hom_j)$$
 (Eq. 3)

The segregation of the BNs of the dominant species of a holm-oak evergreen 162 163 Mediterranean forest (Figure 4a) and their displacement when exposed to drought in a climatic field experiment (Figure 4b) constitute a consistent confirmation of these two first BN rules. 164 Very similar results for segregation and displacement were found using different calculation 165 166 approaches, from the additional use of ratios as explanatory variables to the use of discriminant and other multivariate analyses (Appendix S1: Fig. S1). Pairs of species with more overlap in 167 their distributions had greater differences in their BNs as estimated from PCA scores (PCA based 168 169 on N,P, K, S, Ca and Mg concentrations) in a set of the 12 most representative Mediterranean tree species of South Europe (Figure 4c), confirming the third rule. The BN hypothesis is an 170 171 extension of the ecological niche concept (Schoener 1986) to the bioelemental components of life. The BN aims to establish the morphological and functional "adaptation" to the abiotic and 172 biotic traits that define the niche of each species-genotype from simple determination of the 173 174 elemental composition. It is based on the different contribution of each bioelement to the different organismal functions and structures. Its main advantage over other theoretical classical 175 176 niche concepts is that it can be characterized by the measurable bioelemental composition, the 177 elementome, of the organism, species, community or even biome (Figure 5). This BN

measurement is not biased by the most abundant bioelements since the analysis is conducted fora multivariate space defined by the standardized elemental concentrations (Figures 4 and 6).

180 The quantitative BN framework can then be used to improve projections of how species will respond to environmental changes, provided elementome homeostasis/plasticity (Hom) on a 181 given time horizon is known. This Hom can be defined for example as the coefficient of variance 182 183 of the scores for the different individuals of a species in the PCA. It results from both intraspecific variation and phenotypic individual responses. These two responses can be 184 185 differentiated when the hypervolume occupied by the individuals of one species is compared with the hypervolume generated after changing the environmental conditions, for instance in 186 experiments in field conditions where environmental conditions are changed. 187

188 BN aims to refocus the use of the hypervolume in ecological niche studies. A n-189 dimensional hypervolume was originally proposed by Hutchinson (1957) and defined as a subset 190 of the space, i.e. an n-dimensional geometrical shape, to describe the fundamental niche of a species. Hutchinson suggested that the axes correspond to requirements of the species (e.g. light, 191 temperature, food size) and that the boundaries of the shape indicate the conditions that permit 192 193 the growth and reproduction of the individuals of a species. This description of the species niche 194 has been widely used and discussed (Begon et al, 2006 Chase and Leibold 2003, Colwell and 195 Rangel 2009, Holt 2009, Peterson et al 2011). There is now a growing set of statistical methods 196 to operate this hypothesis of species niche with data (Blonder 2017). BN should be a tool to detect the niche shifts at short-, medium- and long-term scales in response to all type of 197 198 environmental changes (drought, species invasion, pests, new biotic relationships, etc.). The 199 mathematical description of each specific BN position in a multi-space in function of the 200 concentrations of several bioelements can be done by using essential multivariate statistical

201 analyses such as principal components analyses (PCA) or functional/general discriminant analysis (F/GDA) (Sardans et al. 2015; Urbina et al. 2017), or other analyses such as kernel 202 density estimation, support vector machine, Bayesian ellipse, etc. (Blonder 2017). BN distances 203 and the level of overlap/separation among different groups can be computed with the score 204 205 means (centroids) of the main axes (in the PCAs; Figure 4, Appendix S1, Fig. S1) or with 206 squared Mahalanobis distances (in the discriminant analyses; Appendix S1: Fig. S2). These metrics provide numerical evidences of the "distances" between the BN of different species and 207 of the BN displacement (if any) when comparing the BN of different populations of the same 208 209 species along space or time. Correlations between matrices provide a way of comparing the distances between hypervolumes, for instance the correlation matrix of the genome differences 210 211 or of the differences in some ecological or ecophysiological variables in a particular set of 212 species with the corresponding correlation matrix of the BN distances of the same set of species. Statistical tests such as Mantel tests (Mantel, 1967) can be used with this aim. All these 213 214 mathematical tools to estimate the BN allow all levels of comparisons, from species (Sardans et al. 2015) to different taxa and trophic groups (Bartrons et al. 2018), also including the 215 216 comparison of different populations of the same species growing under distinct environmental 217 conditions (Urbina et al., 2015, 2017).

These studies have shown the facility and consistency in the use of different types of multivariate analyses to detect the BN in spatial and temporal gradients/scales. Moreover, the different levels of homeostasis/plasticity of different studied groups can be calculated by the area/volume in that the individuals of each group are spread in the multidimensional space generated. The temporal shifts of the BN of a species or populations of one species can also be calculated and thus their evolution can be followed. BN also allows the description and

224 quantification of niche appearance (i.e. when new species colonize the ecosystem), expansion, contraction, and extinction. Therefore, BN constitutes a quantifiable tool to detect, quantify and 225 thus better comprehend the mechanisms and processes underlying community evolution and 226 species turn-over (Figure 7). Niche construction vs niche expansion or niche contraction vs niche 227 228 extinction are apparently difficult to differentiate because of their similar outcomes, and niche 229 contraction (and extinction) could be the result of environmental filtering, whereas niche expansion could be due to addition of species to a community. Phenomena within the niche 230 theory such as niche convergence (Pianka et al., 2017) can be followed by monitoring the species 231 232 BN hyper-dimension along spatial and temporal gradients and focusing on the changes in the level of overlap/distance in their BN. 233 234 3. Evidence from observational data 235 236 3.1- Taxonomy 237 The BN hypothesis is supported by strong evidence in recent studies. Observational data

(Sardans and Peñuelas 2014; Bartrons et al., 2018) provide evidence that plant and animal 238 species have their own elementome, their own BN. For example, different groups of species in 239 distinct climatic biomes have different foliar bioelemental compositions and stoichiometries 240 (Sardans et al. 2014, 2015, 2016). Foliar bioelemental composition of forest tree species, for 241 242 example, has been positively correlated with phylogenetic distance between taxa at different 243 geographic extents, from local to continental (Sardans et al. 2014, 2015, 2016). Bioelemental 244 composition and stoichiometry thus tend to differ more between phylogenetically distant than 245 between phylogenetically close tree species (Sardans and Peñuelas 2014, Sardans et al. 2015, 246 2016). These positive correlations between elemental composition differences and phylogenetic

247 distances have been also observed in animals (Gonzalez et al. 2018) in a further demonstration of the first process underlying the BN hypothesis. The stoichiometry of an individual under a given 248 climatic and edaphic condition is determined in part by the uptake and the efficiency with which 249 250 the individual uses each bioelement, all of which depend on the trade-offs among the various 251 individual functions that maximize individual fitness under that specific climatic situation. Each 252 species and taxon have its own space in the two-dimensional plot formed by the two first components of the PCA of individual composition (N, P, S, K, Ca and Mg concentrations and 253 their ratios) (See example in Figure 4A and see Appendix S1: Table S1 for the scores of 254 255 additional PCs).

256

257 3.2- Sympatry

Larger differences in the stoichiometry of foliar composition among sympatric than allopatric 258 species have recently been reported both in plant communities (Sardans et al. 2015, 2016) and in 259 260 animal communities (González et al., 2018) as implied in the third process involved in the BN. The overlapping frequency of all pairwise comparisons among many tree species is correlated 261 262 with the elementome distances of the corresponding tree pairs (Figure 4c) (Sardans et al. 2015), 263 thus supporting this third process. Different grass and shrub species growing under different climatic conditions in communities with distinct species compositions maintain distinct species-264 specific foliar compositions, although there may be displacement, expansion or concentration of 265 the biogeochemical space occupied (Figure 6) depending on the coexisting species (Urbina et al. 266 2015, 2017). The species bioelemental composition changes in parallel with the shifts of 267 environmental conditions, in accordance with the new optimal functions that require new 268 proportional use of different bioelements. Contraction, expansion, and displacement at the short-269

term would thus be complemented by construction (allopatric speciation), extinction,

segregation, and equalization at the long-term both for estenoic (narrow range of ecological andenvironmental conditions) and euroic (wide range) niches (Figure 7).

In-depth investigation of these processes constitutes a challenge for ecologists. Some 273 studies have already demonstrated different biogeochemical niche space (euroic-estenoic) among 274 275 different coexisting species-genotypes in plants (Urbina et al., 2015, 2017, Sardans et al., 2015), animals (González et al., 2017, 2018; Bartrons et al., 2018) and microbes (Pereira and Berry, 276 2017). Other studies have observed how the same species displaced their biogeochemical niche 277 depending on the intraspecific and intraspecific competition (Urbina et al., 2015, 2017; Sardans 278 et al., 2016). However, observational field studies along successional stages, field manipulation 279 studies of biotic (introducing new species, changes in density,..), and abiotic (changing climate 280 281 conditions or nutrient availability) followed along time with similar studies in greenhouse, micro- and mesocosm conditions would be necessary to further test the consistency of the new 282

holistic view of ecosystem processes provided by the BN.

284

285

286 3.3- Homeostasis/Plasticity

Changes in environmental supply (for example soil or water bioelemental composition) can affect the bioelemental composition of organisms, as observed in microbes (Frenken et al 2017) and plants (He et al. 2016, Salvagiotti et al. 2017). A trade-off between adaptations to be competitive in stable environments versus being successful in less stable environments should underpin the differences in the continuum of homeostasis-plasticity strategies in bioelemental composition and stoichiometry among individuals, species and communities (Figure 3, Figure

293 4B). Higher N and P concentrations and lower N:P ratios coincide with higher stoichiometric plasticity in opportunistic (r-strategy) than in competitor (K-strategy) plant species (Yu et al. 294 2011). Climate can also account for some of the variability in the bioelemental composition of 295 plant tissues at intra- and interspecific levels (Sardans et al. 2015, 2016, Zhang et al. 2012) 296 297 indirectly though climate controls on bioelemental availability and directly as a forcing of traits 298 linked to the elementome. Environments with fluctuating nutrient availabilities relative to the species' lifetime would have to favor organisms with greater stoichiometric plasticity. 299 Stoichiometric plasticity tends to decrease in response to fertilization as environmental nutrient 300 301 richness increases, in both terrestrial and aquatic systems (Sistla et al. 2015), thus suggesting a different position in the homeostatic-plasticity continuum between species in nutrient-rich versus 302 303 nutrient-poor environments. The BN hypothesis allows to detect the plasticity at two levels: at 304 the individual level (phenotypic plasticity) by showing how its individual elemental composition can vary when environmental conditions shift, and at the intraspecific level as a result of 305 306 intraspecific variability (individual genotypic differences, different ontogenic stage, distinct sex,..) in elemental composition of a set of individuals of the same species living under the same 307 308 environmental conditions. This dual source of variability due to natural intra-population 309 variability and to individual capacity to shift BN in response to environmental changes can be at 310 least partially detected experimentally with multivariate analyses working with populations of 311 the same genotype growing in different environmental conditions across space or time.

312

313 3.4 Evolutionary change

Species have stoichiometric strategies to adapt to heterogeneous environments (Sun et al. 2017)
and are able to adapt their elemental compositions to those of the media after several generations

316 (Turner et al. 2017). The stoichiometry of phytoplankton appears to reflect oceanic stoichiometry over geological timescales, providing even stronger support to this link between organisms' 317 composition and media nutrients availability (Quigg et al. 2003; Reinhard et al. 2017). Other 318 319 studies of terrestrial ecosystems found that microbial community N:P ratios were correlated with 320 plant-community and soil N:P ratios (Kerkhoff and Enquist 2006, Cleveland and Liptzin 2007). 321 Close relationships between microbial and soil C:N ratios across a soil C:N gradient have been also observed, e.g. in Scandinavian boreal forests (Högberg et al. 2006). Evidence suggests that 322 variation in the environmental supply of P affects the expression of highly conserved genes (e.g. 323 324 those of the phosphate-transporter system) and that the acquisition, assimilation and allocation of P is genetically variable (Jevasingh and Weider 2007). Transcriptomic studies of algae 325 326 (Grossman 2000), microbes (Baek and Lee 2007), terrestrial plants (Morcuende et al. 2007) and 327 more recently herbivores (Jeyasingh et al. 2011) have also reported that the availability of P influences genetic expression. Such effects of the supply of P should impinge on fitness and 328 drive evolutionary change. Jeyasingh et al. (2009) observed that genetic variation in 329 phosphoglucose isomerase (Pg) indicated that Pg heterozygotes of Daphnia pulicaria 330 331 outcompeted Pg homozygotes under conditions of low C:P ratios, whereas the opposite was 332 observed under conditions of high C:P ratios.

Availability and stoichiometry of bioelements can determine transcriptional expression and metabolomic structure (Rivas-Ubach et al. 2012) and phenotypic selection (Chowdhury et al. 2015), and can affect genomes and thereby become evolutionary drivers for organisms and ecosystems on the longest timescales (Acquisti et al. 2009). For example, the mechanisms controlling P acquisition are genetically variable, and the expression of the genes involved is strongly affected by its environmental supply (Grossman 2000, Jeyasingh and Weider 2007,

Jeyasingh et al. 2011, 2014). Selection in environments with very low N availability can even
favor new genes richer in N-poor nucleotides (Acquist et al. 2009). Evolution could also be
controlled by directly affecting the quality of matter, e.g. different N:P ratios available to
genomes for the optimal allocation of N and P to DNA/RNA for maximizing fitness (Sardans et
al., 2012b).

344

345 3.5 BN of communities

346 Communities should be more capable to adjust their stoichiometry to additional nutrient inputs and varying stoichiometrical conditions than single species (Danger et al. 2008). A simple and 347 comprehensive example, only considering N and P concentrations and stoichiometry, can be 348 349 observed when P is added to a freshwater community. In such a case, a shift in community 350 composition typically occurs in favor of N-fixing organisms with a posterior cascade of 351 composition changes towards a new equilibrium in community composition (Davidson et al. 352 2007). Environmental changes (anthropogenic or natural) such as perturbations or shifts in climate, N deposition, species invasion or the increase in atmospheric CO₂ can probably generate 353 354 changes in the availability of various nutrients. Species of the target community can then adapt, 355 depending on their phenotypic flexibility. For example, species with a BN better adapted to new 356 environmental conditions can be favored under large long-term loadings of N by N deposition, 357 but species less well adapted can disappear. Species can also turnover, when possible, with the entrance of new species with traits better adapted to life in the newly generated niches (with 358 different optimal elemental compositions). This sequence of events has been observed in several 359 360 field studies where shifts in the elemental composition of the media (water or soil) due to changing conditions (anthropogenic or natural) have been translated into changes in the species 361

composition of species of the community (Sterner and Elser, 2002). The species with more
favorable elemental compositions gain then importance and/or there is species turnover
(Novotny et al. 2007; Yu et al. 2011, Poxleitner et al. 2016, Du 2017). Changes in N:P have ben
related to substantial changes in species composition of plant communities (Peñuelas et al 2013)
and changes in P loads to adjusted C:P ratios in zooplankton communities (Teurlink et al., 2017).

367

368 4- Feedbacks on environment

The relationships between evolution and the bioelemental composition of ecosystems are
bidirectional, because nutrient supply can affect evolutionary processes, while the changes
imposed by evolutionary processes can also affect nutrient supply (Durston and El-Sabaawi
2017). As a result, ocean and soil stoichiometries may have been also partly controlled by
organisms over long time scales (Tyrrell 1999, Lenton and Klausmeier 2007, Hatton et al. 2015,
Zederer et al. 2017).

375 The strongest support of Redfield ratio comes from the fact that the N:P ratio in the open 376 sea areas of all major oceans is remarkably similar to the N:P ratio of plankton (16:1). This is 377 due to the residence time of N and P in the ocean (104 years) relative to the ocean's circulation 378 time (103 years) (Falkowski and Davis 2004). However, in addition to the geochemical causes, 379 biological processes are also underlying Redfield ratios. N₂ fixation can play an important role in 380 ocean N:P ratio homeostasis (Tyrrell 1999, Lenton and Watson 2000, Ganeshram et al. 2002, Kenesi et al. 2009). When ocean N:P ratios fall too low, nitrogen fixation increases by 381 382 augmenting the biomass of N₂-fixers and when these N₂-fixers decompose or are eaten, their N rich organic matter turns to dissolved ammonium and nitrate in the water increasing water N:P 383

ratios. Conversely, when the ratio rises too high, N₂ fixation inputs decrease by P limitation of
 N₂-fixation, thus restoring lower water N:P ratio. This process would confer to P primary control
 of ocean production (Tyrrell 1999, Lenton and Klausmeier 2007).

Soil elemental composition is also controlled by the species of plants (Zederer et al. 2017), the soil microbiota (Hartman et al. 2017), and the interactions between plants and soil fungal communities. For example, *Quercus* species control nutrient dynamics by determining the composition and activity of the forest floor fungal community (Chávez-Vergara et al. 2016) and the litter type (Hatton et al. 2015). Conifers acidifying soil to reduce soil nutrient availabilities (base cations and P) to outcompete species with a different BN provide another consistent example.

394 We should thus expect a constant feedback between organisms and their environment. 395 Understanding the processes underlying species shifts in bioelemental composition and their genetic basis will help us understand the responses of organisms to environmental changes 396 397 (Yamamichi et al. 2015, Leal et al. 2017) and in turn the effects of organisms on ecosystem functioning and services (Frisch et al. 2014, Leal et al. 2017). We are currently still moving 398 towards understanding and identifying the feedbacks between species evolution and shifts in the 399 400 bioelemental composition of ecosystems and how these feedbacks interact and determine 401 ecosystem and evolutionary processes. Available data seems to indicate that the long-term 402 availability and use of bioelements can determine individual selection and therefore ecosystem 403 function and organization and evolution of life.

404

405 5- The biogeochemical niche hypothesis and related approaches

406 Functional traits such as leaf mass area, seed size or plant height have been also used to build hypervolumes to identify species-specific niche (Violle and Jiang, 2009; Violle et al., 2014; 407 Lamanna et al., 2014). However, taxonomical groups (e.g. animals, plants, fungi,...) have 408 different functional traits, preventing the use of these approaches to different taxonomy clades. 409 Instead, all the organisms, independently of their taxonomical groups, are constituted by the 410 411 same basic bio-elements, thus providing a general and easy to measure variable, the elementome, with which to establish the species-specific niche in plants (Sardans et al., 2015), in animals 412 (Gonzalez et al., 2017) and in sets of plants and animals analyzed altogether (Bartrons et al., 413 414 2018). The use of organismal elemental stoichiometry has been proven to be related to the species functional traits in aquatic and terrestrial ecosystems (Meunier et al., 2017, Sardans et al. 415 416 2012c, Peñuelas et al. 2013). This also allows us to establish relationships between observed 417 shifts in ecosystem N:P medium (water or soil) with the shift in species dominance or community and/or in trophic web composition and structure. 418 We first defined the BN in 2008 for plant communities (Peñuelas et al. 2008) and we 419 have since then developed it in several papers (Peñuelas et al. 2008, Sardans and Peñuelas 2014, 420 421 Frisch et al. 2014, Sardans et al. 2015, 2016, Urbina et al. 2017). Other groups have also used 422 this approach (De and Bhattacharyya 2009, Menge and Hedin 2009, Chimphango et al. 2015, Zamora et al. 2017), obtaining results that have provided solid experimental consistency to BN 423 hypothesis. The BN that we define as species elemental composition is wide in scope and simple 424 425 in determination, and reflects the structure and function of each species-genotype resulting from its unique evolutionary history (Sardans et al. 2008, 2015, Urbina et al. 2017). 426 427 BN aimed to progress beyond the classical ecological stoichiometric studies (Elser et al. 2000a,b, Sterner and Elser 2002, Elser and Kyle 2009, Elser et al. 2009, 2010, Rivas-Ubach et al. 428

429 2012, Sardans et al. 2012b). These ecological stoichiometric studies have linked several 430 important ecosystem processes with ecosystem and organismal C:N:P ratios. Several organismal functions and responses to environmental shifts, though, cannot be explained only by the 431 changes in these three main bioelements (Sardans et al. 2012b,c). As stated above, other 432 433 bioelements (such as K, S, Ca, Mg and Fe) are needed in ecological stoichiometric studies to 434 provide more complete and global information correlating shifts in organismal elemental composition with changes in organismal morphology and function in response to environmental 435 conditions (Peñuelas et al. 2008, Peñuelas and Sardans 2009, Sardans and Peñuelas 2014, 436 437 Sardans et al. 2015). The BN hypothesis is thus based on the entire elemental composition or at least on the composition of as many elements as possible. The BN hypothesis has afterwards also 438 been used to refer to the concentrations and availabilities of nutrients in soil or water that favor 439 or determine the corresponding spatial occupation by a species or taxon (De and Bhattacharyya 440 2009, Menge and Hedin 2009, Chimphango et al. 2015, Zamora et al. 2017). 441 442 A similar hypothesis of a "multidimensional stoichiometric niche" was recently

introduced by Gonzalez et al. (2017). These authors defined a "multidimensional stoichiometric 443 niche" for taxonomic/trophic groups as their position in the three-dimensional space determined 444 445 by C, N, and P following Peñuelas et al. (2008). Gonzalez et al (2017) also interestingly opened the possibility to incorporate more elements, and also element fluxes and transformation rates as 446 additional axes. In their studies, Gonzalez et al. (2017, 2018) proved the suitability of this 447 approach for animals. This "multidimensional stoichiometric niche" also tried to incorporate 448 rates of nutrient cycling, but these are much more difficult to quantify than elemental 449 concentrations and ratios. Thus, according with the observations in plants, animals, microbes, 450 451 and also in combination of different taxa altogether (animals and plants) (Peñuelas et al., 2008;

452 Sardans et al., 2015; González et al., 2017; Bartrons et al., 2008), the BN multidimensional space 453 proposed here should allow the establishment and quantification of the links between the species 454 BN distances and their taxonomic/phylogenetic distances and between homeostatic capacities 455 and sympatries, and thus to provide information on species/taxa evolutionary processes. The BN 456 thus also aims to refocus the use of the hypervolume in ecological and evolutionary niche studies 457 as discussed above in section 2.

Another similar approach to BN, ionomics, i.e. the profiles of mineral elements in plants, 458 has been proposed mostly focused on the genetics of plant mineral nutrition. Ionomics analyzes 459 460 the elemental composition of organisms in relation to soil concentrations of nutrients and trace elements and also plant functional traits (Salt et al., 2008; Buescher et al., 2010; Baxter et al., 461 462 2012; Huang and Salt, 2016). Ionomics is mostly used to study the content of bioelements with 463 electrical charge, mostly absorbed from soils, and mostly within a context of the impact on plant health, food quality and/or identification of hyper-accumulation species. It comes from the 464 confluence of metabolomics and plant nutrition (Salt et al., 2008) that, with the incorporation of 465 modern analytical platforms such as ICP techniques, allows to determine most elements 466 467 altogether. The fusion of this information with bioinformatics and genetic tools, such as genomic 468 sequencing, allows the identification of the genes that control uptake, storage, and use of soil nutrients and trace elements in plants (Baxter et al., 2012; Lowry et al., 2012). This further 469 470 allows to analyze the effects of genes and environment on plant nutrition (Buescher, et al., 2010; 471 Lowry et al., 2012; Neugebauer et al., 2018) and plant physiological status (Baxter et al., 2008). BN instead considers all, or as many as possible, bioelements, i.e. also including those such as C 472 473 or N that are initially from atmospheric, not ionic, forms (e.g. N comes mainly from atmospheric N₂; nitrates are rare in bedrock). BN focuses on organisms and on the total elemental 474

475 composition and stoichiometry as a proxy of their optimal function and morphology, whereas ionomics focuses on the composition of the medium (soil) to describe different ionomic niches in 476 the abiotic dimension of an ecosystem. Ionomics has aimed to explain shifts in ionic composition 477 by identifying specific genes and their control in changing environmental conditions, thus 478 479 providing knowledge and information to improve the nutrition of crop species. BN is instead 480 based on the idea that the entire genome of each genotype/species determines species-specific functionality and morphology (more distinct as genomic differences increase) and thus a distinct 481 use of different bioelements, which in turn contribute asymmetrically to different functions and 482 483 morphological structures. BN can thus be used without knowing about the control of gene expression or about gene function. Moreover, whereas ionomic approach has been developed for 484 plants, BN is developed to be equally applicable to all type of taxa. 485

Even though these ionomic studies are mostly agriculturally oriented, and thus the link with the classical theory of ecological niche is not specifically considered, they are consistent with the BN hypothesis, at least at the level to correlate elemental composition with phylogenetic distance. They constitute strong evidence in favor of BN consistency and constitute a very interesting contribution to the BN.

491 7. Concluding summary and perspectives

Here we propose a BN of genotypes/species/communities that should provide an improved and
quantifiable proxy of the general ecological niche. The BN is characterized by a particular
hypervolume of their individual elementomes based on their specific needs and use of
bioelements in different amounts and proportions for their structure and functioning. Several
organismal structures and functions and the responses to environmental shifts cannot be
explained only by the changes in the three main bioelements, C, N and P. We propose that other

498 bioelements such as K, S, Ca, Mg, Fe, Mn and Mo, among others, must be incorporated to fully describe the BN. The BN extends the classical ecological niche concept but has the advantage 499 that it can be easily, directly and quantitatively analyzed. 500 501 The biogeochemical distances, i.e. the differences in elementome, among species, are a function of taxonomy (phylogenetic distance), sympatry (the bioelemental compositions should differ 502 503 more among coexisting than among non-coexisting species to avoid competitive pressure), and 504 homeostasis (within the continuum between high homeostasis/low plasticity and low 505 homeostasis/high plasticity). Further studies are warranted to discern the ecological and evolutionary processes involved in the BN of all types of individuals, taxa and ecosystems. The 506 changes in bioelemental availability and use at long timescales should determine phenotypic 507 selection and therefore also ecosystem function and organization, and, at the end, the evolution 508 of life and the environment. 509

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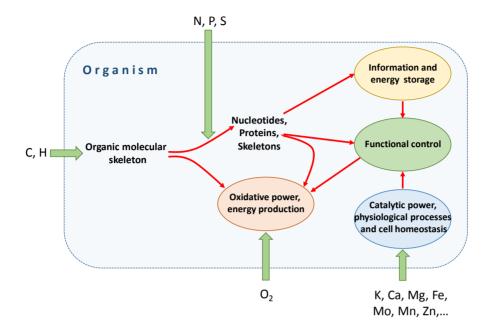
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804 Figure captions



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806 Figure 1. Schematic of the use of bioelements by organism in different functions. C, H, and O are the basis of organismic structure, whereas N, P and S enter in the various biomolecules 807 allowing biochemical reactions. The molecules that control life processes (e.g. RUBISCO or 808 809 chlorophyll), that carry information (e.g. DNA and RNA), and that control the storage and release of energy when and where necessary (e.g. ATP and NAD) all incorporate N, P and in 810 811 lower amounts S. Other bioelements, normally present in low concentrations such as potassium (K), magnesium (Mg), iron (Fe), calcium (Ca), molybdenum (Mo), manganese (Mn) and zinc 812 (Zn) are cofactors in catalytic reactions in several biochemical and physiological processes 813 814 including for example photosynthesis, respiration, or cellular ion balance. The arrows indicate the processes in which each bioelement is involved. Green arrows indicate the main organismal 815 functions to which each bioelement contributes. The red arrows indicate the relationships of the 816 817 different bio-molecules with basic organismal structures and functions.

	Feelesieel	Ecological variables					Stoichiometry traits				
	Ecological - strategy		Growth rate	Body size	Reproductive strategy	Other	Stoichiometry flexibility	[C]	[N]	[P]	[K]
	K-strategy (competitor)	Long	Medium	Large	Few seeds but with highly caloric reserves	Strong investment in structure	Medium	_	t	t	_
6 38	A-strategy (stress- tolerator)	Medium	Slow	Medium -small	Medium	Strong investment in C-rich secondary compounds	Low	t	ţ	ţ	t
	r-strategy (ruderal)	Short	Fast	Small	Many seeds but with low caloric reserves	General trend to develop belowground vegetative & reproductive structures	High		t	† †	_

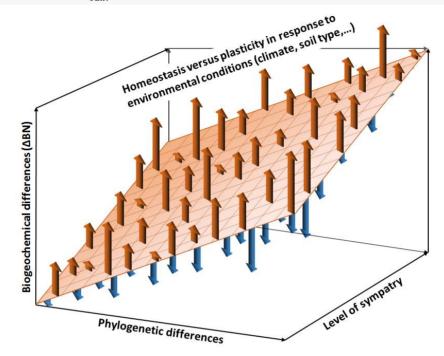
The example of plants

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Figure 2. The example of plants. Hypothesized asymmetric uses of bioelements in evolution,
depending on the ecological strategy of each species. Vertical arrows indicate low or high
concentration values (double arrows indicate especially high values). Dashes indicate not clear
patterns. There is a different use of all these elements depending on the ecological strategy of
each species (Grime 1977).

 $BN_{i} = BN \text{ of } taxa_{i} = Elementome_{i1-n} = (C: N: P: K: S: Mg: Ca: ...)_{i1-n} = BNtax_{i} + \frac{\partial BNtax_{i}}{\partial Env} = BNtax_{i} + f(Tax_{i}, Symp_{i}, Hom_{i})$ $BN_{j} = BN taxa_{j=} Elementome_{j1-n} = (C: N: P: K: S: Mg: Ca ...)_{j}$ $BN_{j} = BNtax_{j} + \frac{\partial BNtax_{j}}{\partial Env} = BNtax_{j1-n} + f(Tax_{j}, Symp_{j}, Hom_{j})$

 $\Delta BN_{i-j} = (BNtax_i - BNtax_j) + \frac{\partial BN_{tax_i} - \partial BN_{tax_j}}{\partial Env} = (BNtax_i - BNtax_j) + [f(Tax_i, Symp_i, Hom_i) - f(Tax_j, Symp_j, Hom_j)]$



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Figure 3. Theoretical distribution of the biogeochemical differences between species as a 826 function of (1) the taxonomic and phylogenetic distances between them, (2) the level of 827 828 sympatry (overlap frequency in their corresponding areas of distribution) and of (3) the 829 homeostatic or flexible response to current environmental conditions (e.g. climate, soil traits). The surface represents the biogeochemical distance resulting from phylogenetic distance and 830 level of sympatry, and the brown and blue arrows the positive and negative residuals of the BN 831 distance induced by variable current environmental conditions (e.g. climate, soil traits, 832 833 neighbors).

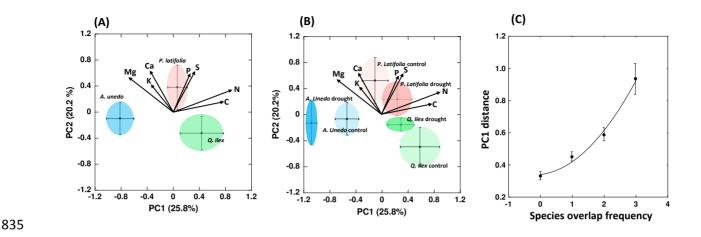


Figure 4. (A) Biogeochemical niche segregation among coexisting plant species. We analyzed 836 837 the foliar elementomes of the three dominant trees (Quercus ilex, Arbutus unedo and Phillyrea 838 *latifolia*) in a field experiment of climatic manipulation in an evergreen Mediterranean forest in 839 the Prades Mountains (Catalonia). We plotted the plant scores for the first two principal components of the principal component analysis (PCA) conducted with foliar N, P, K, S, Ca, Mg 840 841 concentrations as variables. (B) Shifts in biogeochemical niche in a drought experiment. The mean \pm confidence intervals (95%, represented by the ellipses) of the species \times climate 842 interaction groups are depicted in the plot thus comparing control plants to those submitted to a 843 decrease of 20 % of water availability during six years in field conditions. The other PCAs (from 844 PC3 to PC7) and the corresponding other axes are now shown in the Appendix S1: Table S1. 845 (C) Score distances for PC1 of the PCA of the foliar stoichiometry (foliar N, P, K, S, Ca and Mg 846 concentrations and their ratios) in a set of the 12 most representative Mediterranean species of 847 South Europe as a function of the frequency of pairwise species overlap (coincidence in the same 848 849 community): six broadleaf species, Quercus ilex, Q. faginea, Q. pyrenaica, Q. suber, Castanea sativa and Eucalyptus globulus, and six needleleaf species, Pinus halepensis, P. pinaster, P. 850 nigra, P. pinea, Juniperus oxycedrus and J. thurifera). 0, no overlap; 1, occasional overlap; 2, 851

moderate overlap; 3, frequent overlap. Based on Rivas-Martinez et al. (2001) De Bolos and Vigo

853 (2001), CREAF, Catalonia cover map, (2013) and Folk et al. (1984).



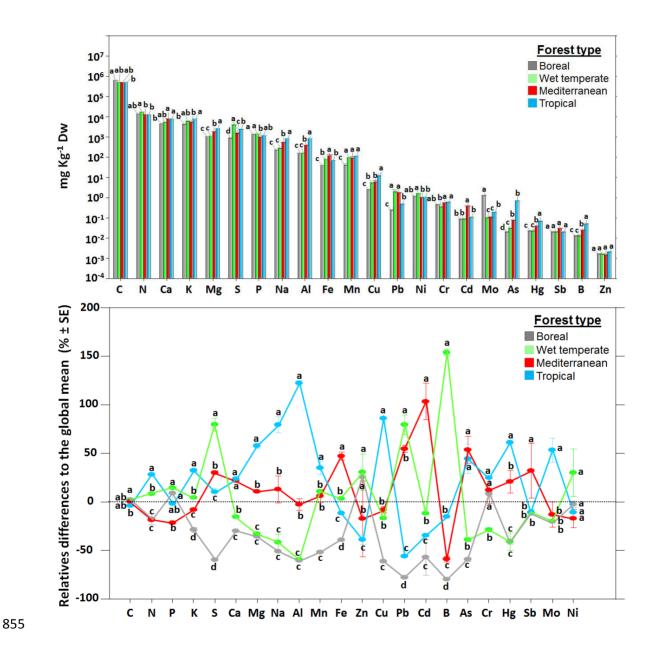


Figure 5. Elementome, full spectrum of bioelemental content, of boreal, wet temperate,
Mediterranean and tropical forests. (A) Elementomes calculated from web of science data

858 incorporated to our GEU-CREAF dataset (Boreal forest: 473 sites and 1890 trees; Wet temperate

- forests: 2882 sites and 12665 trees; Mediterranean forests: 2001 sites and 2901 trees; Tropical
- 860 wet forest: 1100 sites and 1821 trees). The values in plot B show the difference of each forest
- type relative to the global mean ($\% \pm SE$). We have used data from the main available data bases
- 862 (e.g. TRY and ICP forest) and from an exhaustive search for woody plants from non-
- 863 manipulated areas in WEB of Science

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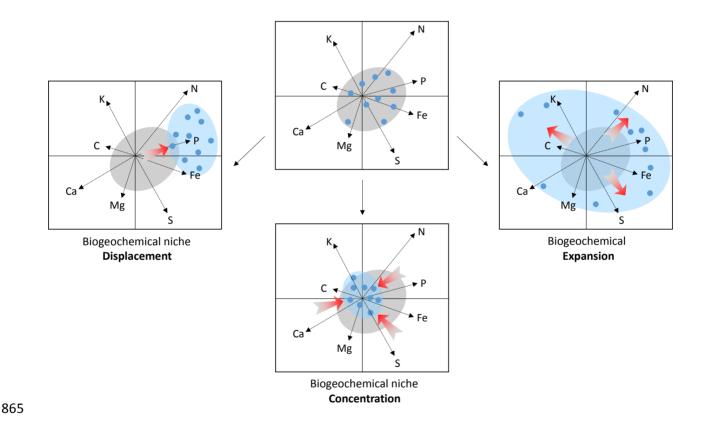
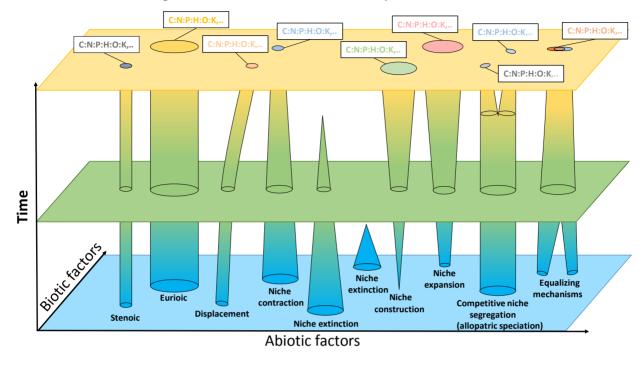
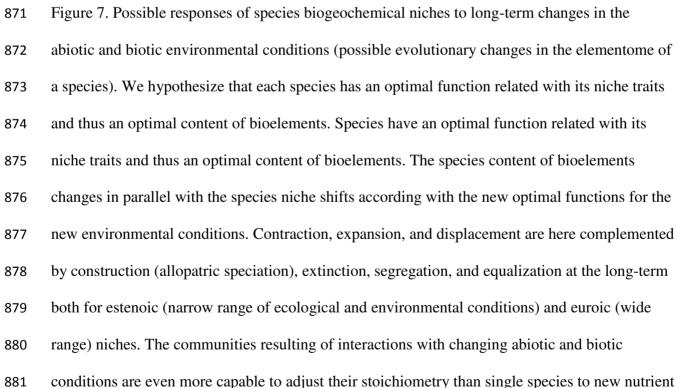


Figure 6. Three possible biogeochemical niche responses to short-term changes in the
environmental conditions: Contraction, Expansion or Displacement. Blue dots represent
individuals of a species, grey and blue ellipses represent the initial and final BN space.



Biogeochemichal niches defined by the different elementomes



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supply and stoichiometrical conditions. Each particular niche constitutes a singular functional

and physical position in the ecosystem that can vary with time in response to changes in abiotic

- variables (climate, perturbations such as volcanic eruptions, etc.) and to changes in the impacts
- and feed-backs of the other organisms living in it (biotic factors). In this context, each niche
- should determine a specific and particular content of different bioelements in each moment. This
- should also determine a singular elementome accompanying the niche characteristics of each
- genotype and species at both micro- and macro-evolutionary timescales.