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Review



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The other ocean acidification problem: CO_2 as a resource among competitors for ecosystem dominance

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Predictions concerning the consequences of the oceanic uptake of increasing atmospheric carbon dioxide (CO2) have been primarily occupied with the effects of ocean acidification on calcifying organisms, particularly those critical to the formation of habitats (e.g. coral reefs) or their maintenance (e.g. grazing echinoderms). This focus overlooks direct and indirect effects of CO2 on non-calcareous taxa that play critical roles in ecosystem shifts (e.g. competitors). We present the model that future atmospheric [CO₂] may act as a resource for mat-forming algae, a diverse and widespread group known to reduce the resilience of kelp forests and coral reefs. We test this hypothesis by combining laboratory and field CO₂ experiments and data from 'natural' volcanic CO₂ vents. We show that mats have enhanced productivity in experiments and more expansive covers in situ under projected near-future CO2 conditions both in temperate and tropical conditions. The benefits of CO2 are likely to vary among species of producers, potentially leading to shifts in species dominance in a high CO₂ world. We explore how ocean acidification combines with other environmental changes across a number of scales, and raise awareness of CO2 as a resource whose change in availability could have wide-ranging community consequences beyond its direct effects.

1. Introduction

Ecosystem collapses often go unanticipated because their drivers are unrecognized, have indirect effects or combine in unexpected ways to alter interactions between key species [1,2]. To redress this uncertainty, much work has been carried out to identify and quantify stressors and their cumulative effects by observing historical change (review on ecosystem collapses [3]), experimentally manipulating drivers (review of synergies [4]) and modelling their outcomes (alternate stable states [5]). By combining such lines of evidence, ecologists have demonstrated the tractability of identifying drivers of change [6] to develop frameworks for anticipating or managing change [7].

Ocean acidification, a result of rising carbon dioxide (CO₂) [8], is a particularly vexing stressor to assess as a potential driver of future ecosystem change [9]. This is because ocean acidification represents a series of changes in seawater chemistry, with each alteration representing a potential driver of change [10]. The diminishing availability of carbonate ion (CO₃^{2–}), and ensuing reduction in calcium carbonate (CaCO₃) saturation states are widely reported to reduce calcification in a wide range [11,12] of, but not all, calcifying organisms [13,14]. To date, much of the focus of ocean acidification research has been on the response of calcifiers, both algae and invertebrates, to the changing carbonate system, with a particular preoccupation on one property: the hydrogen ion concentration [H⁺], which is frequently reported as pH owing to the relative ease of its measurement. Initial insights were derived from manipulative experiments, particularly single-factor manipulations of CO_2 levels (i.e. pH) that assess change from a physiological and often single-species perspective. These assessments highlight the vulnerabilities of calcifying organisms (e.g. review [15]) and consider the potential extent of ecological change [16].

While there is recognition that ocean acidification may alter net primary production, via the increased solubility of biogenic calcareous structures and reduced survival of calcifying species that consume algae [11,17], it may also alter production directly, especially under elevated temperature [18]. Several reviews now recognize that ocean acidification can increase carbon fixation rates in some photosynthetic organisms (review [19,20]); however, the ecological implications of this increase are largely untested. Thus, CO₂ may not only act as a stressor but also act as a resource. CO2 may not only limit primary productivity but also limit the growth rates of a population. Because such effects are variable among species, there is an enormous potential for shifts in species dominance, as some species gain a relative advantage from their response to elevated CO₂. Indeed, alteration of resource availability has a fundamental role in regulating the productivity of individuals, populations and, ultimately, communities [21].

2. Carbon dioxide as a carbon resource

 CO_2 can act as a resource by increasing carbon fixation rates in some photosynthetic organisms. The degree to which this occurs is dependent on the carbon capture strategies and the degree to which carbon is limiting (review [20]). The relationship between aqueous CO_2 , photosynthesis and growth is not simple, because not all photosynthesizing species require environmental CO_2 for their source of carbon (C). The majority of marine algae have carbon concentrating mechanisms (CCMs) that facilitate the active influx of CO_2 and/or bicarbonate ions (HCO₃⁻) and elevate C concentrations at the site of C fixation (i.e. Rubisco), with few algae being CO_2 -only users [22,23]. For these reasons, there has been uncertainty regarding whether algae with CCMs will benefit from CO_2 enrichment [24].

Despite the prevalence of CCMs, evidence suggests that many algae do respond positively to increasing CO₂ [12]. Indeed, the ability of some algae with CCMs to benefit from enriched CO₂ lends insights into the potential mechanisms for CO₂ effects. Species with CCMs can shift away from HCO_3^- towards aqueous CO₂ when CO₂ levels are high [25]. Thus, aqueous CO₂ may facilitate CCM C-acquisition or make it less energetically costly, and this capacity varies among species [26]. Such variation among species may be due to species-specific variation in CCMs themselves [27,28] or the ability of species to acquire other limiting resources (e.g. light or nutrients) [29].

As human activities modify environmental conditions, and therefore resource availability, some species of algae may be released from carbon limitations while others are not [30]. This mismatch has the potential to affect competitive abilities and alter community structure. Moreover, the effects of these shifts would be particularly profound if key functional groups, whose interactions structure entire communities, experience contrasting resource limitations.

Falkenberg *et al.* [31] reveal how variability in resource limitation may play a substantial role in 'kelp-to-turf'

phase-shifts. Kelps were shown to be limited by a single type of resource (e.g. only nutrients and not CO_2), whereas their alternate state (characterized by turf algae) was co-limited by both nutrients and CO_2 . This difference in carbon limitation demonstrates the potential for elevated CO_2 to influence turf expansion [32], especially when amplified by human activities that also increase nutrient loads [33]. This response is characteristic of co-limitation by multiple resources [34]. Therefore, ocean acidification is unlikely to act alone, but instead acts in concert with other environmental conditions (e.g. nutrient pollution) and primary consumers (i.e. herbivores).

Indeed, there are a large number of resources that constrain the abundance of marine [35] and terrestrial plants [36] and determine the composition of space-holding species. Combinations of two or, sometimes, three of these limiting factors are often incorporated into models to account for the diversity and composition of plant communities (review [37]). Thus, interspecific variation in resource requirements is likely to affect relative species abundance in complex communities. Changes in any of these constraints could alter the probability of phase-shifts, particularly as many resources are undergoing large, rapid changes because of human activities [38].

Mat-forming algae (here defined as low-profile groundcovering macroalgal and turf communities) are fast-growing and can be effective competitors for space in kelp forests and on coral reefs. The hypothesis that phase-shifts towards matforming algae are likely to be more common under conditions of high [CO_2] is therefore of particular interest. If this model has validity, then enhanced CO_2 should cause mats to increase their extent and productivity in both temperate and tropical systems. Supporting evidence requires field observations of natural variation in [CO_2] to provide insights into ocean acidification effects at the ecosystem level (e.g. CO_2 seeps), and field and laboratory manipulations to establish the physiological reasons for such an effect [39]. Until now, such a combination of evidence has not been strongly pursued.

3. Ocean acidification as an indirect agent

While categorizing the potential effects of ocean acidification into direct negative effects (i.e. on calcifiers) and positive effects (i.e. on non-calcifying algae) provides a conceptual starting point, it is manifestly overly simplistic to forecast community persistence or change. While the detection of *direct* effects is readily achievable, it will not necessarily shed light on the drivers that will play predominant roles in shaping future communities relative to indirect effects [40]. Indirect effects yield not only 'unexpected results' but also some of the strongest ecological effects (i.e. phase-shifts) that have been regarded as 'catastrophes' on coral reefs [41] and 'collapses' of kelp forests [42]. They are often unanticipated, because the impact of one component (e.g. ocean acidification) on another (e.g. kelp decline) requires knowledge of a third species (e.g. kelp-competitors) or mediating factors (e.g. interactions among stressors), which are poorly understood (figure 1).

While the indirect climate effects on species interactions were initially surprising [43], we are starting to learn that they commonly lag behind the more intuitive and easily detectable direct effects, and can even reverse the direct effects of climate [16,44]. To date, most research on ocean acidification has focused on the direct environmental impacts on individual 2

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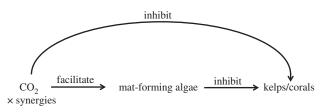


Figure 1. Diagram showing the indirect effect of enriched CO_2 on habitatforming species (i.e. kelps and corals) via mat-forming algae. CO_2 facilitates the growth of mat-forming algae, which, in turn, inhibit the recruitment of kelps and corals. The indirect effects of CO_2 may rival the direct effects of CO_2 , particularly where CO_2 combines with other stressors to accelerate the expansion of mats.

species of iconic status (e.g. corals), but the capacity for indirect effects to drive change may rival the known direct effects [40].

The indirect effects of ocean acidification may be especially important in communities where CO_2 can act as a resource and change competitive outcomes. Competitive ability is contingent on the relative availability of resources and the tolerance to environmental stress. Dominance may switch between subordinate species and their normally dominant competitors, or even change competitive interactions into facilitated interactions [45]. Studies of volcanic CO_2 vents reveal the potential for altered competitive balances between space-holders at projected near-future levels of CO_2 (750 ppm [46]). In particular, these studies reveal the potential for increased dominance of mat-forming macroalgae in low pH environments in temperate [47] and tropical zones [48].

Indirect effects that account for phase-shifts can be classified into two broad classes: those driven by strong consumer effects (i.e. trophic cascades) and those driven by strong producer effects (i.e. dominance of space; review [40]). The former class has attracted considerable attention in both tropical and temperate systems [49], but there is profound and widespread variation in its explanatory power (reviews of phase-shifts on coral reefs [50] and kelp forests [51]). While there is merit in understanding the future role of acidification on consumers and potential trophic cascades, there is also merit in examining the breadth of direct and indirect producer effects (figure 1). This review provides an alternative framework for forecasting community-level impacts of ocean acidification by focusing on producers' responses through the lens of phase-shifts in kelp forests and coral reefs. By 'kelp forests', we refer to canopies of brown seaweeds, including Laminariales and Fucales.

Our present understanding of the physiological effects of ocean acidification is primarily based on results obtained from closed experimental systems, particularly small laboratory experiments. Beyond physiological predictions, larger field mesocosm studies enable tests of direct and indirect effects that can modify the persistence of species and functional relationships. Together, such experimental knowledge can be compared against the pattern observed in natural systems. Such investigative frameworks have been established in the discipline of macroecology [52] in response to the persistent need to scale from small to larger experimental units and ultimately ecosystems [53]. For ocean acidification research, this challenge is particularly vexing, because the logistic constraints on manipulating large volumes of aqueous CO2 place upper limits on tests of combinations of multiple stressors and species. Thus, there is considerable discussion on methods that integrate experimental research across laboratory, mesocosm and field systems to unravel the complexity of functional relationships under future environmental conditions [9]. Nevertheless, the key question for ecosystems is not based on just the direct effects and fate of individual species, but on the stability and persistence of the system as a whole.

4. Mat-forming algae as competitive dominants

We focus on a group of fast-growing algae with quick rates of invasion and growth and wide physiological tolerance. We classify these relatively low-lying algae in a single term, 'mats' or 'mat-forming algae', to emphasize their similar physiological and ecological properties across the temperate and tropical communities. We propose that if we can identify similarities that occur between kelp forests of the temperate zone and coral reefs of the tropical zone, then we open a broader avenue of enquiry about the potential effects of environmental change via altered productivity.

Mats are characteristically small, with high surface-areato-volume ratios and high demand for resources relative to surrounding kelps and corals. Physiologically, they require increased resource availability (e.g. nutrients [54]) to enable their normally ephemeral status to become competitively superior to perennial species of kelps and corals. The taxonomic identity potentially includes many tens to hundreds of species [55] from many lineages. Despite their taxonomic diversity, similarities in their biology have been found to be sufficiently large that authors have consistently referred to them as 'turf-forming algae' or 'turfs', 'epilithic algal matrix' or 'mats' [55]. These terms emphasize the carpet-like nature of these algae which we call 'mats'.

The 'coral–algal phase-shift' occurs as a function of reduced herbivory and water quality that independently, or in combination, enhances net production of algae [56]. Ephemeral algae that form dense carpet-like mats are a key mechanism that creates physically stressful conditions for corals [57,58]. On the Great Barrier Reef, 'mats' of algae (10–50 cm high) that bloom across large areas of coral reefs on coastal and inshore fringing reefs are known to overgrow established corals and prevent coral recruitment [58,59]. In the Caribbean, greater emphasis is placed on turf-forming algae comprising filaments of benthic algae and cyanobacteria (less than 1 cm height) that can also prevent settlement of coral planulae [60].

Kelps and corals form three-dimensional structures that are episodically disturbed and readily reassemble after natural disturbances. The loss of corals and kelps can occur when mats of algae colonize bare substrata after a disturbance, hence preventing recruitment by occupying purchases for attachment and smothering remnant corals. Many matforming species are normally subordinate to kelps and corals (e.g. ephemeral or early successional species), but increased resource availability enables their physiology and life history to become competitively superior [33,61].

The 'kelp-turf phase-shift' occurs on coasts associated with water pollution, because the altered water conditions favour a suite of turf-forming species, which due to their physiology (i.e. fast uptake of nutrients) and life history (ability to withstand high sediment loads) are well suited for polluted environments [33,62–64]. On Caribbean coral reefs, nutrient overloading also causes turfs to expand over extensive areas, leading to coral recruitment failure [65], similar to that observed in kelp forests [61].

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Table 1. The total alkalinity (TA) and pCO_2 were increased, and pH decreased, in treatments where CO_2 was elevated. The effect on algae was similar among the methods (i.e. vents, versus field versus laboratory) and climes (i.e. temperate versus tropical). Measured: total alkalinity (TA) and pH; calculated, concentration (ppm) of pCO_2 ; calculated: effect-size (ES) of the treatments on algae. The magnitude of the effect-size (*r*) was calculated using Cohen's *d* for two independent groups (i.e. the difference between means divided by standard deviation (Cohen's *d* [68]).

	temperate				tropical			
	ТА	рН	pC02	ES	ТА	рН	pCO 2	ES
vents				0.70				0.82
elevated	2560	7.76	805		2302	7.83	748	
ambient	2563	8.01	305		2297	8.00	401	
field mesocosm				0.88				0.81
elevated	2357	7.91	647		2320	7.87	779	
ambient	2276	8.15	307		2320	8.10	416	
laboratory				0.82				0.73
elevated	2585	7.96	557		2260	7.84	828	
ambient	2263	8.14	300		2258	8.16	342	

5. Lines of evidence: carbon dioxide increases net productivity

(a) Comparison among methods: vents versus field

mesocosms versus laboratory

Here, we compare the effects of CO2 enrichment on matforming algae among several projects, combining findings from laboratory experiments and mesocosms with those of natural field studies. For field studies, we use observations of benthic communities around volcanic seeps that have been shaped by decades-long exposure to fine bubble streams of CO₂ seeping from the seabed, which cause local alterations in the seawater chemistry [47,48]. These seep sites complement and improve the interpretive value of laboratory experiments. While laboratory studies can be carefully controlled, the range of testable ecological interactions is quite limited. Conversely, field studies at CO2 vents include interactions within acclimatized natural communities, but spatial and temporal variation in pH does not behave exactly the same as future ocean conditions [66]. The limitations of laboratory and field assessments suggest that combining both approaches will provide a more informed model.

Our comparisons focus on benthic communities exposed to similar elevations in CO₂ concentrations that are within the forecasted range of change by the end of this century [46,67]. Relative to ambient conditions (8.0–8.1 pH_T), two levels of projected future pH were compared in warm temperate (Mediterranean [47]) and tropical (Papua New Guinea [48]) regions (table 1). The future pH conditions were chosen to represent the projected near-future conditions anticipated for the end of the twenty-first century (pCO_2 approx. 580–1080, depending on emission scenarios [69]). Near-future conditions were represented as 7.7–7.8 pH_T (temperate) and 6.8–7.8 pH_T (tropical). Experimental mesocosms in temperate [26] and tropical zones (i.e. the Coral Proto-Free Ocean Carbon Experiment [70]) are compared with the vents and their laboratory counterparts (figure 2).

Experimental and observational results are not directly comparable because they estimate responses from mats

grown over different periods of time using different techniques. Laboratory experiments used similar techniques [18], but mesocosm experiments used floating docks in temperate conditions [26] and a Free Ocean Carbon Experiment in tropical conditions [70]. Observations at volcanic vents quantified the percentage cover of mats within space that would otherwise be suitable for recruitment (e.g. coral planualae settlement on space lacking live coral). At tropical vents, mats were quantified within 25 × 25 cm quadrats randomly assigned among reef without live coral. At temperate vents, mats were quantified after 3.5 months of development on uncolonized 15 × 15 cm volcanic rock tiles [47].

(b) General carbon dioxide effects on the net production of mats

Mats responded positively to conditions of ocean acidification among all studies reviewed here. Mats covered nearly 40% more space at temperate vents and 50% more space at tropical vents (figure 2a,b). In both cases, mats expanded their covers from a minority space holder (i.e. approx. 10–20% cover) to a majority space holder (i.e. approx. 60%), regardless of sources of loss.

The effect of enriched CO_2 in field mesocosms was to increase rates of growth (mg per day dry weight) by two to three times that of ambient conditions (figure 2*c*,*d*). In the laboratory, net productivity was substantially greater, probably owing to fewer sources of loss (e.g. herbivory), but the increased growth rate under enriched treatments was about twice that of ambient conditions (figure 2*e*,*f*). These consistently strong effects suggest that CO_2 -induced increases in productivity may be quite general among a class of algae renowned for its taxonomic diversity, variation in morphology and life-history characteristics.

An unexpected and preliminary result is the similarity in magnitude of response between climes and experimental protocols (i.e. effect-sizes based on relative change in growth or percentage cover between ambient and elevated CO_2 conditions ranged from 0.70 to 0.88; table 1). We caution over interpretation of these similarities, because the responses

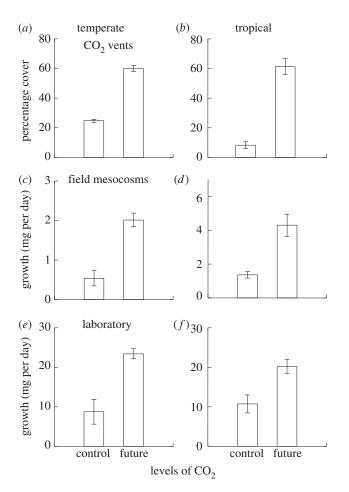


Figure 2. Elevated net productivity and cover of mat-forming algae under enriched CO₂ conditions. In the natural environment, the percentage cover of algal mats (standardized by available space, see §5*c*) is greater at CO₂ vents in temperate (*a*. Ischia, Mediterranean) and tropical seas (*b*. Papua New Guinea). Growth rates are greater under CO₂-enriched treatments within temperate (*c*) and tropical (*d*) field manipulations and temperate (*e*) and tropical (*f*) laboratory manipulations. 'Control' represents current CO₂ conditions (approx. $300-410 \ p$ CO₂) and 'future' represents end of this century conditions (approx. $550-830 \ p$ CO₂).

are derived from mats grown over different periods of time (electronic supplementary material, S1). In addition, the magnitude of CO_2 -induced increases in productivity is likely to be contingent on the availability of nitrogen and light as well as species identity. Nevertheless, the consistency in direction of strong effects is clearly suggestive that mat-forming algae have the potential to respond positively to predicted levels of CO_2 .

(c) Carbon dioxide effects on community composition

Mat-forming algae appeared to expand (i.e. higher percentage cover) under projected near-future conditions at CO₂ seeps, regardless of any differences in herbivory (figure 2). At the temperate vents, estimates of percentage cover of mats on the natural substrate were similar among ambient and low pH zones (mean percentage cover was 26% and 33%, respectively [47]). However, the differences among these zones were readily apparent in a study of succession, which monitored the development of rocky reef assemblages on uncolonized tiles of rock in each pH zone. During succession, calcareous taxa (primarily crustose coralline algae and barnacles) maintained similar percentage cover in both the ambient and low pH zones for the early stages of succession (until approx. 3.5 months) but were overgrown by mat-forming algae (figure 2) in the low pH zone during the later stages [47]. At the tropical vents, overall abundance of non-calcareous macroalgae was twofold greater at the low pH than ambient zones (mean percentage cover of 3.33% and 1.60%, respectively [48]). When standardized by available space (hard substratum without live corals), however, the difference between these zones was more apparent, with mats of filamentous turf– algae being far more expansive under near-future conditions than at control sites (figure 2), whereas calcareous algae showed the opposite pattern [48].

In both tropical and temperate regions, shifts in the benthic communities surrounding the vents were consistent with proposed direct and indirect effects of ocean acidification on matforming algae. Mat-forming algae at temperate vents may limit the percentage cover of calcified taxa, which at the end of a year of development was 26% in the ambient pH zones and 17% in the low pH zones [47]. Hence, mat-forming algae appear to inhibit other taxa [47]. Among the tropical vents, at a pH below 7.7 (i.e. conditions beyond those expected for the end of the twenty-first century), no coral reef development was found, and the benthos was dominated by seagrasses, ephemeral macroalgae and volcanic sand, and a few individual colonies of robust coral species (massive *Porites* and *Favites pentagona*) [48].

(d) Herbivores

As anticipated, the densities of calcareous benthic grazers (e.g. sea urchins and gastropods) can be sparser in extremely low pH conditions [47,71,72]. However, in the projected near-future conditions presented in these analyses (figure 2), sea urchin densities are similar to control sites at tropical vents (K.E.F. 2013, unpublished data) and temperate vents [72,73]. In addition, the abundance of small mobile grazers (primarily gastropods) does not differ among the low pH and control sites at the temperate vents [74]. Differences in herbivorous fish populations are unknown at both these vents; however, herbivorous fish can easily access all areas of the temperate and tropical vents and did not appear to avoid areas of low pH ([74], K.E.F. 2013, personal observation). The increased probability that calcareous benthic grazers are sparser at low pH levels suggests that there is merit in understanding the conditions in which herbivory may be altered by enriched CO₂.

6. Discussion

(a) Ocean acidification and resource availability

Human-driven environmental changes are producing regional combinations of environmental conditions that may push many ecological systems outside the environmental envelope in which they evolved [75]. The relative abundance of sessile plants often reflects resource limitations, but many of these constraints are undergoing large, rapid changes capable of causing phase-shifts [76]. These new environmental conditions appear to favour species with fast rates of colonization, growth and short generation times that can competitively displace slower growing and longer-living space-holders when resource availability is increased [33,38,56].

The effects of CO_2 enrichment that we synthesized here were derived from substantially different species, regions

and experimental protocols (electronic supplementary material, S1). The consistency of positive effects on productivity alone suggests that previous reports of CO2 effects on expansive covers of algae on temperate and tropical rocky coasts [71] and increasing dominance of mats over corals [30] may not represent an idiosyncratic set of case examples. Naturally, ephemeral mats may not only displace recovering kelp forests under elevated nutrients [33,61], but this loss may be strengthened by future CO_2 concentrations (i.e. nutrient \times CO_2 synergy [77]). Similarly, mats may have variable effects on corals recovering from disturbance under current conditions [78], but CO2 may strengthen their negative effects on corals via increased growth rates of mats [30] coupled with decreased growth rates of corals [79]. Together, these studies provide insights into ecological drivers that to date have attracted comparatively little attention, but represent potentially profound drivers of change to the competitive dominance of kelps and corals within their respective latitudes.

(b) Indirect effects on herbivores

The role of herbivores in the trends presented here is a current gap in understanding. The studies from the CO₂ seeps report that herbivores are present in end-of-century pH conditions, and at least some species are in similar abundance between the low pH and control sites. This suggests the increased dominance of mats in the projected near-future conditions is not predominantly due to the absence of calcareous grazers. However, decades of research have highlighted the importance of grazers in mediating competitive interactions between benthic space holders [80,81], and research is required to assess whether an increased supply of resources (e.g. CO₂) can change these ecosystems from predominant consumer control to predominant resource control (e.g. nutrients [82]).

There are several ways in which herbivores could contribute to or counter the trends reported here that deserve further attention. For example, changes in abundance or behaviour (resulting in reduced per capita herbivory) could contribute to the dominance of mat-forming algae. However, laboratory experiments reveal that gastropods reared on turfs grown in various CO₂ concentrations removed more algae per herbivore under elevated CO_2 ; this effect was driven by the indirect effect of CO2 via changes to C:N ratios in the algae rather than the direct effect on the herbivore [83]. Thus, increased herbivory could compensate for increased growth by mats, although this was not apparent at the CO₂ seeps. Furthermore, some diatoms produce tissue with lower nutrient concentrations in response to elevated CO₂ [84]. If acidification were to similarly affect the nutrient status of benthic algae, individual herbivores would need to increase *per capita* rates of consumption to acquire sufficient energy for growth and development. Conversely, reduced nutrient content of algae could cause herbivores to grow more slowly and suffer higher mortality in the process, or even selectively graze certain species of algae. Elevated CO₂ can also affect the production of chemical herbivore deterrents in marine plants [85], which could further influence the interaction between herbivores and algae. Last, the effects of ocean acidification on herbivores are likely to be influenced by concurrent changes in temperature [86] that can affect population size and per capita rates of consumption [87]. While there is a multitude of potential interactions and indirect effects of acidification on herbivores that require further attention, the

results from CO₂ seeps suggest that an increased dominance of mat-forming algae is a likely emergent effect.

(c) Multiple stressors: ocean acidification as a non-additive stressor

Most ecological systems are exposed to multiple stressors. The projected rates of CO₂ absorption in the future [88] suggest that ocean acidification represents a relatively slow driver of future ecological change. However, the gradual trend in ocean acidification will be overlaid on additional stressors to marine ecosystems (e.g. temperature, fishing and nutrients) as well as short-term perturbations that occur at local to regional scales (e.g. storms, coral bleaching, coral disease and mass urchin mortalities). Many stressors combine in nonadditive ways (i.e. synergies or antagonisms). Recent reviews reveal synergistic effects in about one-third of factorial experiments involving two stressor combinations [89], which increases to three-quarters when three stressors are considered [4]. Indeed, it is probably a simple function of sample size (*n*) that the frequency of synergistic and antagonistic effects will increase with an increase in number of stressors. Hence, the potential for acidification-related phase-shifts in response to gradual acidification is likely to increase with additional stressors [90].

Not surprisingly, the potential for non-additive effects between multiple stressors increases the probability of phase-shifts [56]. When combined with elevated nutrients, elevated CO₂ can release mats from co-limitation with nitrogen [31] to multiply the rate of expansion [77]. In addition, projected warming is likely to enhance mat productivity, because mats are not only better suited to the increasing availability of CO₂ [31], but these effects are also enhanced under warmer temperatures [18]. Whether additive or synergistic, such combinations of CO₂ with other stressors increase the probability of phase-shifts.

A tractable aspect of synergisms with local stressors is that local management may ameliorate the effects of CO₂ enrichment by reducing local stressors. For example, although CO₂ and eutrophication can yield multiplicative responses [77], a reduction of local stressors (e.g. nutrients) can negate this synergistic effect [26]. These results suggest that the management of local stressors (e.g. water pollution and overfishing) may have a greater contribution in determining the ecosystem effects of ocean acidification than current thinking allows. Such findings empower local managers because they show that reducing local stressors (e.g. nutrient pollution and overfishing) can reduce the effects of global stressors not under their governance (e.g. ocean acidification and temperature).

(d) Conclusion

Ocean acidification is often considered in terms of its direct negative effects on the growth and calcification of organisms with calcareous shells or skeletons. We argue that this focus overlooks the important role of ocean acidification as a resource, which can enhance the productivity of algae known to influence the status of kelp forests and coral reefs (i.e. mat-forming algae or mats). We have highlighted how ocean acidification can indirectly tip the competitive balance towards dominance by mats through mechanisms that generate new space (e.g. disturbance or storm events), which enables 6

colonization and persistence of mats rather than the original kelp or coral state.

Ocean acidification, therefore, has the capacity to act as a resource that shifts the status of subordinates into dominant competitors. Consequently, human activities that alter the availability of resources have important implications for the relative competitive abilities of major ecosystem components. We suggest that additional stressors will influence the effect of ocean acidification on producers, and that many cumulative impacts may reflect multiplicative rather than additive interactions. Contrary to conventional wisdom, we argue that if these synergies involve local stressors, then environmentally mediated ecosystem shifts may be greatly ameliorated by managing local stressors. Nevertheless, there are few assessments of whether management of local processes can weaken the feedbacks that reinforce altered state and enable the

reversibility of phase-shifts. Importantly, we suggest that in the face of changing climate (e.g. ocean acidification and temperature), effective management of local stressors (e.g. water pollution and overfishing) may have a greater contribution in determining ecosystem states than currently anticipated. Thus, we highlight how ocean acidification has the potential to influence competitive abilities via changes in resource availability, with implications for the stability and persistence of the system as a whole.

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References

- Wootton JT. 1994 The nature and consequences of indirect effects in ecological communities. *Annu. Rev. Ecol. Syst.* 25, 443–466. (doi:10.1146/annurev. es.25.110194.002303)
- Suding KN, Hobbs RJ. 2009 Threshold models in restoration and conservation: a developing framework. *Trends Ecol. Evol.* 24, 271–279. (doi:10. 1016/j.tree.2008.11.012)
- Jackson BC. 2001 What was natural in the coastal oceans? Proc. Natl Acad. Sci. USA 98, 5411–5418. (doi:10.1073/pnas.091092898)
- Crain CM, Kroeker K, Halpern BS. 2008 Interactive and cumulative effects of multiple human stressors in marine systems. *Ecol. Lett.* **11**, 1304–1315. (doi:10.1111/j.1461-0248.2008.01253.x)
- Scheffer M, Carpenter SR. 2003 Catastrophic regime shifts in ecosystems: linking theory to observation. *Trends Ecol. Evol.* 18, 648–656. (doi:10.1016/j.tree. 2003.09.002)
- Hughes TP *et al.* 2007 Phase shifts, herbivory, and the resilience of coral reefs to climate change. *Curr. Biol.* **17**, 360–365. (doi:10.1016/j.cub.2006. 12.049)
- Lebel L, Anderies JM, Campbell B, Folke C, Hatfield-Dodds S, Hughes TP, Wilson J. 2006 Governance and the capacity to manage resilience in regional socialecological systems. *Ecol. Soc.* 11, 19.
- Godbold JA, Solan M. 2013 Long-term effects of warming and ocean acidification are modified by seasonal variation in species responses and environmental conditions. *Phil. Trans. R. Soc. B.* 368, 20130186. (doi:10.1098/rstb.2013.0186)
- Russell BD, Harley CDG, Wernberg T, Mieszkowska N, Widdicombe S, Hall-Spencer JM, Connell SD. 2012 Predicting ecosystem shifts requires new approaches that integrate the effects of climate change across entire systems. *Biol. Lett.* 8, 164–166. (doi:10.1098/rsbl.2011.0779)
- Doney SC, Fabry VJ, Feely RA, Kleypas JA. 2009 Ocean acidification: the other CO₂ problem. *Annu. Rev. Mar. Sci.* **1**, 169–192. (doi:10.1146/annurev. marine.010908.163834)

- Hoegh-Guldberg O *et al.* 2007 Coral reefs under rapid climate change and ocean acidification. *Science* **318**, 1737–1742. (doi:10.1126/science. 1152509)
- Kroeker KJ, Kordas RL, Crim RN, Singh GG. 2010 Meta-analysis reveals negative yet variable effects of ocean acidification on marine organisms. *Ecol. Lett.* 13, 1419–1434. (doi:10.1111/j.1461-0248.2010. 01518.x)
- Ries JB, Cohen AL, McCorkle DC. 2009 Marine calcifiers exhibit mixed responses to CO₂-induced ocean acidification. *Geology* **37**, 1131–1134. (doi:10.1130/g30210a.1)
- Findlay HS, Wood HL, Kendall MA, Spicer JI, Twitchett RJ, Widdicombe S. 2011 Comparing the impact of high CO₂ on calcium carbonate structures in different marine organisms. *Mar. Biol. Res.* 7, 565–575. (doi:10.1080/17451000.2010.547200)
- Byrne M. 2011 Impact of ocean warming and ocean acidification on marine invertebrate life history stages: vulnerabilities and potential for persistence in a changing ocean. *Oceanogr. Mar. Biol., An Annu. Rev.* 49, 1–42.
- Wootton JT, Pfister CA, Forester JD. 2008 Dynamic patterns and ecological impacts of declining ocean pH in a high-resolution multi-year dataset. *Proc. Natl Acad. Sci. USA* **105**, 18 848–18 853. (doi:10. 1073/pnas.0810079105)
- Orr JC *et al.* 2005 Anthropogenic ocean acidification over the twenty-first century and its impact on calcifying organisms. *Nature* **437**, 681–686. (doi:10.1038/nature04095)
- Connell SD, Russell BD. 2010 The direct effects of increasing CO₂ and temperature on non-calcifying organisms: increasing the potential for phase shifts in kelp forests. *Proc. R. Soc. B* **277**, 1409–1415. (doi:10.1098/rspb.2009.2069)
- Harley CDG, Anderson KM, Demes KW, Jorve JP, Kordas RL, Coyle TA, Graham MH. 2012 Effects of climate change on global seaweed communities. *J. Phycol.* 48, 1064–1078. (doi:10.1111/j.1529-8817.2012.01224.x)

- Koch M, Bowes G, Ross C, Zhang XH. 2013 Climate change and ocean acidification effects on seagrasses and marine macroalgae. *Glob. Change Biol.* 19, 103–132. (doi:10.1111/j.1365-2486.2012.02791.x)
- Harpole WS *et al.* 2011 Nutrient co-limitation of primary producer communities. *Ecol. Lett.* 14, 852–862. (doi:10.1111/j.1461-0248.2011.01651.x)
- 22. Beardall J, Giordano M. 2002 Ecological implications of microalgal and cyanobacterial CO_2 concentrating mechanisms, and their regulation. *Funct. Plant Biol.* **29**, 335–347. (doi:10.1071/pp01195)
- Giordano M, Beardall J, Raven JA. 2005 CO₂ concentrating mechanisms in algae: mechanisms, environmental modulation, and evolution. *Annu. Rev. Plant Biol*, 56, 99–131. (doi:10.1146/annurev. arplant.56.032604.144052)
- Hurd CL, Hepburn CD, Currie KI, Raven JA, Hunter KA. 2009 Testing the effects of ocean acidification on algal metabolism: considerations for experimental design. *J. Phycol.* 45, 1236–1251. (doi:10.1111/j.1529-8817.2009.00768.x)
- Cornwall CE, Hepburn CD, Pritchard D, Currie KI, McGraw CM, Hunter KA, Hurd CL. 2012 Carbon-use strategies in macroalgae: differential responses to lowered pH and implications for ocean acidification. *J. Phycol.* 48, 137–144. (doi:10.1111/j.1529-8817. 2011.01085.x)
- Falkenberg LJ, Connell SD, Russell BD. 2013 Disrupting the effects of synergies between stressors: improved water quality dampens the effects of future CO₂ on a marine habitat. J. Appl. Ecol. 50, 51–58. (doi:10.1111/1365-2664.12019)
- Rost B, Riebesell U, Burkhardt S, Sultemeyer D. 2003 Carbon acquisition of bloom-forming marine phytoplankton. *Limnol. Oceanogr.* 48, 55–67. (doi:10.4319/lo.2003.48.1.0055)
- Beardall J, Raven JA. 2004 The potential effects of global climate change on microalgal photosynthesis, growth and ecology. *Phycologia* 43, 26–40. (doi:10.2216/i0031-8884-43-1-26.1)
- 29. Hepburn CD, Pritchard DW, Cornwall CE, McLeod RJ, Beardall J, Raven JA, Hurd CL. 2011 Diversity of

7

carbon use strategies in a kelp forest community: implications for a high CO₂ ocean. *Glob. Change Biol.* **17**, 2488–2497. (doi:10.1111/j.1365-2486.2011. 02411.x)

- Diaz-Pulido G, Gouezo M, Tilbrook B, Dove S, Anthony KRN. 2011 High CO₂ enhances the competitive strength of seaweeds over corals. *Ecol. Lett.* 14, 156–162. (doi:10.1111/j.1461-0248.2010. 01565.x)
- Falkenberg LJ, Russell BD, Connell SD. 2013 Contrasting resource limitations of marine primary producers: implications for competitive interactions under enriched CO₂ and nutrient regimes. *Oecologia* 172, 575-583. (doi:10.1007/s00442-012-2507-5)
- Connell SD, Russell BD, Turner DJ, Shepherd SA, Kildea T, Miller DJ, Airoldi L, Cheshire A. 2008 Recovering a lost baseline: missing kelp forests from a metropolitan coast. *Mar. Ecol. Prog. Ser.* 360, 63-72. (doi:10.3354/meps07526)
- Gorman D, Russell BD, Connell SD. 2009 Land-tosea connectivity: linking human-derived terrestrial subsidies to subtidal habitat change on open rocky coasts. *Ecol. Appl.* **19**, 1114–1126. (doi:10.1890/ 08-0831.1)
- Allgeier JE, Rosemond AD, Layman CA. 2011 The frequency and magnitude of non-additive responses to multiple nutrient enrichment. J. Appl. Ecol. 48, 96-101. (doi:10.1111/j.1365-2664.2010.01894.x)
- Schiel DR, Foster MS. 1986 The structure of subtidal algal stands in temperate waters. *Oceanogr. Mar. Biol., An Annu. Rev.* 24, 265–307.
- Chapin FS, Bloom AJ, Field CB, Waring RH. 1987 Plant-responses to multiple environmental factors. *Bioscience* 37, 49–57. (doi:10.2307/1310177)
- Venterink HO, Wassen MJ, Verkroost AWM, de Ruiter PC. 2003 Species richness-productivity patterns differ between N-, P-, and K-limited wetlands. *Ecology* 84, 2191–2199. (doi:10.1890/01-0639)
- Tilman D, Lehman C. 2001 Human-caused environmental change: impacts on plant diversity and evolution. *Proc. Natl Acad. Sci. USA* 98, 5433–5440. (doi:10.1073/pnas.091093198)
- Katsikatsou M, Anestis A, Portner HO, Vratsistas A, Aligizaki K, Michaelidis B. 2012 Field studies and projections of climate change effects on the bearded horse mussel *Modiolus barbatus* in the Gulf of Thermaikos, Greece. *Mar. Ecol. Prog. Ser.* 449, 183–196. (doi:10.3354/meps09550)
- Connell SD, Russell BD, Irving AD. 2011 Can strong consumer and producer effects be reconciled to better forecast 'catastrophic' phase-shifts in marine ecosystems?. J. Exp. Mar. Biol. Ecol. 400, 296–301. (doi:10.1016/j.jembe.2011.02.031)
- Hughes TP. 1994 Catastrophes, phase-shifts, and large-scale degradation of a Caribbean coral-reef. *Science* 265, 1547–1551. (doi:10.1126/science.265. 5178.1547)
- Jackson JBC *et al.* 2001 Historical overfishing and the recent collapse of coastal ecosystems. *Science* 293, 629–638. (doi:10.1126/science.1059199)
- Sanford E. 1999 Regulation of keystone predation by small changes in ocean temperature. *Science* 283, 2095–2097. (doi:10.1126/science.283.5410.2095)

- Suttle KB, Thomsen MA, Power ME. 2007 Species interactions reverse grassland responses to changing climate. *Science* **315**, 640–642. (doi:10.1126/ science.1136401)
- Kawai T, Tokeshi M. 2007 Testing the facilitationcompetition paradigm under the stress-gradient hypothesis: decoupling multiple stress factors. *Proc. R. Soc. B* 274, 2503–2508. (doi:10.1098/rspb. 2007.0871)
- Caldeira K, Wickett ME. 2003 Anthropogenic carbon and ocean pH. *Nature* 425, 365. (doi:10.1038/ 425365a)
- Kroeker KJ, Micheli F, Gambi MC. 2012 Ocean acidification causes ecosystem shifts via altered competitive interactions. *Nat. Clim. Change* 3, 156–159. (doi:10.1038/nclimate1680)
- Fabricius KE *et al.* 2011 Losers and winners in coral reefs acclimatized to elevated carbon dioxide concentrations. *Nat. Clim. Change* 1, 165–169. (doi:10.1038/nclimate1122)
- Krebs CJ. 2009 Ecology: the experimental analysis of distribution and abundance, 6th edn. 688 p. San Francisco, CA: Benjamin Cummings.
- Bruno JF, Sweatman H, Precht WF, Selig ER, Schutte VGW. 2009 Assessing evidence of phase shifts from coral to macroalgal dominance on coral reefs. *Ecology* **90**, 1478–1484. (doi:10.1890/ 08-1781.1)
- Foster MS, Schiel DR. 2010 Loss of predators and the collapse of southern California kelp forests (?): alternatives, explanations and generalizations. *J. Exp. Mar. Biol. Ecol.* 393, 59–70. (doi:10.1016/j. jembe.2010.07.002)
- Connell SD, Irving AD. 2009 The subtidal ecology of rocky coasts: local-regional-biogeographic patterns and their experimental analysis. In *Marine macroecology* (eds JD Witman, R Kaustuv), pp. 392–417. Chicago, IL: University of Chicago Press.
- Levin SA. 1992 The problem of pattern and scale in ecology. *Ecology* **73**, 1943 – 1967. (doi:10.2307/ 1941447)
- Pedersen MF, Borum J. 1996 Nutrient control of algal growth in estuarine waters. Nutrient limitation and the importance of nitrogen requirements and nitrogen storage among phytoplankton and species of macroalgae. *Mar. Ecol. Prog. Ser.* **142**, 261–272. (doi:10.3354/meps142261)
- Connell SD, Foster MS, Airoldi L. In press. What are algal turfs? Towards a better description of turfs. *Mar. Ecol. Prog. Ser.*
- Bellwood DR, Hughes TP, Folke C, Nystrom M. 2004 Confronting the coral reef crisis. *Nature* 429, 827–833. (doi:10.1038/nature02691)
- Martinez JA, Smith CM, Richmond RH. 2012 Invasive algal mats degrade coral reef physical habitat quality. *Estuar. Coast Shelf Sci.* 99, 42–49. (doi:10.1016/j.ecss.2011.12.022)
- Hauri C, Fabricius KE, Schaffelke B, Humphrey C. 2010 Chemical and physical environmental conditions underneath mat- and canopy-forming macroalgae, and their effects on understorey corals. *PLoS ONE* 5, e12685. (doi:10.1371/journal.pone. 0012685)

- Jompa J, McCook LJ. 2002 Effects of competition and herbivory on interactions between a hard coral and a brown alga. J. Exp. Mar. Biol. Ecol. 271, 25–39. (doi:10.1016/s0022-0981(02)00040-0)
- Vermeij MJA, Sandin SA. 2008 Density-dependent settlement and mortality structure the earliest life phases of a coral population. *Ecology* 89, 1994–2004. (doi:10.1890/07-1296.1)
- Gorman D, Connell SD. 2009 Recovering subtidal forests on human-dominated landscapes. J. Appl. Ecol. 46, 1258–1265. (doi:10.1111/j.1365-2664. 2009.01711.x)
- Benedetti-Cecchi L, Pannacciulli F, Bulleri F, Moschella PS, Airoldi L, Relini G, Cinelli F. 2001 Predicting the consequences of anthropogenic disturbance: large-scale effects of loss of canopy algae on rocky shores. *Mar. Ecol. Prog. Ser.* 214, 137–150. (doi:10.3354/meps214137)
- Eriksson BK, Johansson G, Snoeijs P. 2002 Longterm changes in the macroalgal vegetation of the inner Gullmar Fjord, Swedish Skagerrak coast. *J. Phycol.* 38, 284–296. (doi:10.1046/j.1529-8817. 2002.00170.x)
- Airoldi L. 2003 The effects of sedimentation on rocky coast assemblages. *Oceanogr. Mar. Biol., An Annu. Rev.* 41, 161–236.
- 65. Vermeij MJA, van Moorselaar I, Engelhard S, Hornlein C, Vonk SM, Visser PM. 2010 The effects of nutrient enrichment and herbivore abundance on the ability of turf algae to overgrow coral in the Caribbean. *PLoS ONE* 5, e14312. (doi:10.1371/ journal.pone.0014312)
- Hofmann GE *et al.* 2011 High-frequency dynamics of ocean pH: a multi-ecosystem comparison. *PLoS ONE* 6, e28983. (doi:10.1371/journal.pone.0028983)
- Feely RA, Sabine CL, Hernandez-Ayon JM, Ianson D, Hales B. 2008 Evidence for upwelling of corrosive 'acidified' water onto the continental shelf. *Science* 320, 1490–1492. (doi:10.1126/science.1155676)
- Cohen J. 1988 Statistical power analysis for the behavioural sciences, 2nd edn. Hillsdale, NJ: Lawrence Erlbaum Associates.
- 69. IPCC. 2007 Climate change 2007: the physical science basis. Contribution of Working Group 1 to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge, UK: Cambridge University Press.
- Kline DI *et al.* 2012 A short-term in situ CO₂ enrichment experiment on Heron Island (GBR). *Sci. Rep.* 2, 413. (doi:10.1038/srep00413)
- Hall-Spencer JM, Rodolfo-Metalpa R, Martin S, Ransome E, Fine M, Turner SM, Rowley SJ, Tedesco D, Buia MC. 2008 Volcanic carbon dioxide vents show ecosystem effects of ocean acidification. *Nature* 454, 96–99. (doi:10.1038/nature07051).
- Calosi P, Rastrick SPS, Graziano M, Thomas SC, Baggini C, Carter HA, Hall-Spencer JM, Milazzo M, Spicer JI. In press. Distribution of sea urchins living near shallow water CO₂ vents is dependent upon species acid – base and ion-regulatory abilities. *Mar. Pollut. Bull.* (doi:10.1016/j.marpolbul.2012.11.040)
- 73. Kroeker KJ, Gambi MC, Micheli F. In press. Community dynamics and ecosystem simplification

in a high-CO₂ ocean. *Proc. Natl Acad. Sci.* (doi:10. 1073/pnas.1216464110)

- Kroeker KJ, Micheli F, Gambi MC, Martz TR. 2011 Divergent ecosystem responses within a benthic marine community to ocean acidification. *Proc. Natl Acad. Sci. USA* **108**, 14 515–14 520. (doi:10.1073/ pnas.1107789108)
- Duarte CM, Hendriks IE, Moore TS, Olsen YS, Steckbauer A, Ramajo L, Carstensen J, Trotter JA, McCulloch M. 2013 Is ocean acidification an openocean syndrome? Understanding anthropogenic impacts on seawater pH. *Estuar. Coasts* 36, 221–236. (doi:10.1007/s12237-013-9594-3)
- Vitousek P, Mooney H, Lubchenco J, Melillo J. 1997 Human domination of earth's ecosystems. *Science* 277, 494–499. (doi:10.1126/science.277. 5325.494)
- Russell BD, Thompson JI, Falkenberg LJ, Connell SD. 2009 Synergistic effects of climate change and local stressors: CO₂ and nutrient driven change in subtidal rocky habitats. *Glob. Change Biol.* **15**, 2153–2162. (doi:10.1111/j.1365-2486.2009.01886.x)
- Bender D, Diaz-Pulido G, Dove S. 2012 Effects of macroalgae on corals recovering from disturbance. *J. Exp. Mar. Biol. Ecol.* **429**, 15–19. (doi:10.1016/ j.jembe.2012.06.014).

- Schneider K, Erez J. 2006 The effect of carbonate chemistry on calcification and photosynthesis in the hermatypic coral *Acropora eurystoma. Limnol. Oceanogr.* **51**, 1284–1293. (doi:10.4319/lo.2006.51. 3.1284)
- Lubchenco J. 1978 Plant species diversity in a marine intertidal community: importance of herbivore food preference and algal competitive abilities. *Am. Nat.* **112**, 23–39. (doi:10.1086/ 283250)
- Carpenter RC. 1986 Parttitioning herbivory and its effects on coral-reef algal communities. *Ecol. Monogr.* 56, 345–363. (doi:10.2307/1942551).
- Worm B, Lotze HK. 2006 Effects of eutrophication, grazing, and algal blooms on rocky shores. *Limnol. Oceanogr.* **51**, 569–579. (doi:10.4319/lo.2006.51. 1_part_2.0569)
- Falkenberg LJ, Russell BD, Connell SD. In press. Future herbivory: the indirect effects of enriched CO₂ may rival its direct effects. *Mar. Ecol. Prog. Ser.*
- Rossoll D, Bermudez R, Hauss H, Schulz KG, Riebesell U, Sommer U, Winder M. 2012 Ocean acidification-induced food quality deterioration constrains trophic transfer. *PLoS ONE* 7, e34737. (doi:10.1371/journal.pone.0034737)

- Arnold T, Mealey C, Leahey H, Miller AW, Hall-Spencer JM, Milazzo M, Maers K. 2012 Ocean acidification and the loss of phenolic substances in marine plants. *PLoS ONE* 7, e35107. (doi:10.1371/ journal.pone.0035107)
- O'Connor MI. 2009 Warming strengthens an herbivore-plant interaction. *Ecology* **90**, 388–398. (doi:10.1890/08-0034.1)
- Russell BD, Connell SD, Findlay HS, Tait K, Widdicombe S, Mieszkowska N. 2013 Ocean acidification and rising temperatures may increase biofilm primary productivity but decrease grazer consumption. *Phil. Trans. R. Soc. B* 368, 20120438. (doi:10.1098/rstb.2012.0438)
- Feely RA, Sabine CL, Lee K, Berelson W, Kleypas J, Fabry VJ, Millero FJ. 2004 Impact of anthropogenic CO₂ on the CaCO₃ system in the oceans. *Science* **305**, 362–366. (doi:10.1126/science.1097329)
- Darling ES, Côté IM. 2008 Quantifying the evidence for ecological synergies. *Ecol. Lett.* **11**, 1278–1286. (doi:10.1111/j.1461-0248.2008.01243.x)
- Hughes TP, Linares C, Dakos V, van de Leemput IA, van Nes EH. 2013 Living dangerously on borrowed time during slow, unrecognized regime shifts. *Trends Ecol. Evol.* 28, 149–155. (doi:10.1016/j.tree. 2012.08.022)