

1 **Understanding environmental change through the lens of trait-based,**
2 **functional and phylogenetic biodiversity in freshwater ecosystems**

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4 Janne Alahuhta^{1*}, Tibor Erős², Olli-Matti Kärnä¹, Janne Soininen³, Jianjung Wang^{3,4,5}, Jani Heino⁶

5

6 1 Geography Research Unit, University of Oulu, P.O. Box 3000, FI-90014 Oulu, Finland

7 2 Balaton Limnological Institute, MTA Centre for Ecological Research, Tihany, Hungary

8 3 Department of Geosciences and Geography, University of Helsinki, P.O. Box 64, FI-00014

9 Helsinki, Finland

10 4 Nanjing Institute of Geography and Limnology, Chinese Academy of Sciences, 73, East Beijing

11 Road, 210008 Nanjing, China

12 5 University of Chinese Academy of Sciences, 380 Huaibeizhuang, Huairou, 101408 Beijing, China

13 6 Finnish Environment Institute, Biodiversity Centre, Paavo Havaksen tie 3, FI-90530 Oulu, Finland

14

15

16 * Janne Alahuhta, Geography Research Unit, University of Oulu, P.O. Box 3000, 90014 Oulu,

17 Finland. Email: janne.alahuhta@oulu.fi, Fax: +358 8 344 084

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19 **Running head:** Environmental change and freshwater biodiversity

20 **Paper type:** Research Review

21

22 **Abstract**

23

24 In the era of the Anthropocene, environmental change is accelerating biodiversity loss across
25 ecosystems on Earth, among which freshwaters are likely the most threatened. Different biodiversity
26 facets in the freshwater realm suffer from various environmental changes that jeopardize the
27 ecosystem functions and services important for humankind. In this work, we examine how
28 environmental changes (e.g. climate change, eutrophication or invasive species) affect trait-based,
29 functional and phylogenetic diversity of biological communities. We first developed a simple
30 conceptual model of the possible relationships between environmental change and these three
31 diversity facets in freshwaters, and secondly, systematically reviewed articles where these
32 relationships had been investigated in different freshwater ecosystems. Finally, we highlighted
33 research gaps from the perspectives of organisms, ecosystems, stressors and geographical locations.
34 Our conceptual model suggested that both natural factors and global change operating at various
35 spatial scales influence freshwater community structure and ecosystem functioning. The relationships
36 between biodiversity and environmental change depend on geographical region, organism group,
37 spatial scale and environmental change gradient length. The systematic review revealed that
38 environmental change impacts biodiversity patterns in freshwaters, but there is no single type of
39 biodiversity response to the observed global changes. Natural stressors had different, even
40 contradictory effects (i.e., multiple, negative and positive) on biodiversity compared with
41 anthropogenic stressors. Anthropogenic stressors more often decreased biodiversity, although
42 eutrophication and climate change affected freshwater ecosystems in a complex, more
43 multidimensional way. The research gaps we identified were related, for example, to the low number

44 of community-based biodiversity studies, the lack of information on true phylogenies for all
45 freshwater organism groups, the missing evaluations whether species traits are phylogenetically
46 conserved, and the geographical biases in research (i.e., absence of studies from Africa, Southern
47 Asia and Russia). We hope that our review will stimulate more research on the less well-known facets
48 and topics of biodiversity loss in highly vulnerable freshwater ecosystems.

49

50 **Keywords:** Community ecology, Diversity index, Functional diversity, Global change, Lakes,
51 Phylogenetic diversity, Rivers, Species traits, Streams

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66 **Introduction**

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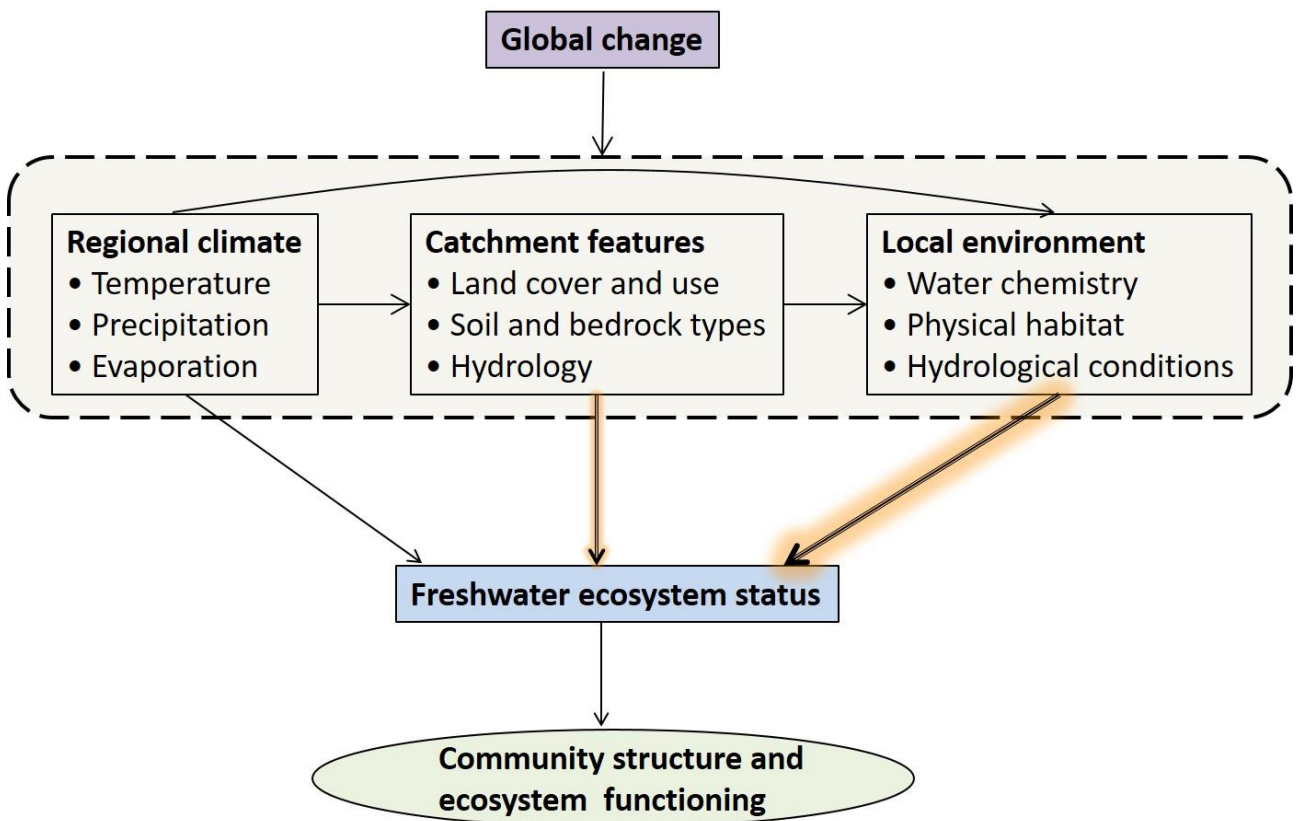
68 Environmental change affects biodiversity, but its influence varies in time and space, including within
69 and across ecosystems (Hooper et al. 2012; Dornelas et al. 2014). In the era of the Anthropocene, the
70 general understanding is that biodiversity loss is accelerating, for example, due to increased
71 atmospheric greenhouse gases, land use alteration, environmental pollution including eutrophication,
72 overexploitation of species and invasion of exotic species (McGill et al. 2015; Maxwell et al. 2016).
73 Such undesirable progress affecting biodiversity is also jeopardizing ecosystem functions and
74 services vital to human well-being (Cardinale et al. 2012). In this sense, perhaps the most threatened
75 ecosystems exposed to environmental changes are freshwaters (Dudgeon et al. 2006; Vörösmarty et
76 al. 2010; Wiens 2016; Vilmi et al. 2017). This is because many freshwater species have limited ability
77 to disperse in the face of changing environmental conditions (Heino et al. 2009) and they are subject
78 to multiple anthropogenic pressures acting simultaneously (Woodward et al. 2011). In addition,
79 freshwaters are not often part of the biodiversity conservation programs.

80

81 Although freshwaters account for only ca. 1% of the Earth's total surface area, they are especially
82 important ecosystems, because they 1) are hosting relatively larger proportion of biodiversity
83 compared to terrestrial systems and 2) constitute a source for many of the essential but threatened
84 ecosystem services, such as drinking water supplies, aquaculture and climate change mitigation
85 (Dudgeon et al. 2006; Cardinale et al. 2012). In addition, freshwater and terrestrial ecosystems are
86 fundamentally interrelated through the movement of energy, nutrients and other materials (Soininen
87 et al. 2015). For example, organic matter within a catchment area and terrestrial organisms enter lentic

88 and lotic systems, whereas aquatic insects emerge and fly to surrounding riparian zones, where they
89 are eaten by terrestrial predators. Thus, freshwater ecosystems depend on multiple environmental
90 characteristics operating at various spatial scales (Fig. 1). These issues not only highlight the
91 importance to maintain and protect the taxonomic diversity of ecological communities, but also other
92 facets of biodiversity in the freshwater realm at various spatial scales.

93



94

95 Fig. 1. Conceptual illustration of the relationships between environmental change and freshwater
96 community structure and ecosystem functioning. Freshwater (abiotic) ecosystem status is influenced
97 by different environmental variables, ranging in an increasing order of importance from regional
98 climate and catchment features to local environmental features. Ecological status of surface waters
99 per se comprises of many water quality variables such as nutrient status and oxygen levels.

100

101 Community ecologists have measured various aspects of biodiversity concurrently within species
102 assemblages, including trait-based, functional and phylogenetic diversity. In general, a species trait
103 is any single feature or quantifiable feature of an organism that affects its performance or fitness in
104 relation to abiotic and biotic factors (McGill et al. 2006). A set of species traits is related to a site
105 where a species can actually live, how species interact with each other, the strength of competition or
106 consumption efficiency of a predator, and the contribution of species to ecosystem functioning
107 (McGill et al. 2006; Cadotte et al. 2011). Functional diversity is traditionally defined as the diversity
108 of species traits in ecosystems and measures how an ecosystem operates or functions without
109 necessarily considering organisms' evolutionary history (Petchey and Gaston, 2006; Schleuter et al.
110 2010). Phylogenetic diversity, on the other hand, comprises the differences in evolutionary history of
111 species in a community and can possibly be used as a proxy for functional diversity if the species
112 traits considered are phylogenetically conserved (Winter et al. 2012). Phylogenetic diversity captures
113 various species traits, but is not informative for identifying what they might be (Flynn et al. 2011).
114 These alternative approaches may provide better generality in understanding and predicting the
115 assembly of ecological communities and ecosystem functions than more traditional approaches based
116 on species taxonomic identity (Devictor et al. 2010; Schleuter et al. 2010; Gagic et al. 2015).
117 Although more research is being devoted to understanding and measuring these aspects of
118 biodiversity, our knowledge of their response to environmental change is still limited in freshwater
119 ecosystems (Vaughn 2010; Woodward et al. 2010).

120

121 To better understand how environmental change affects trait-based, functional and phylogenetic
122 diversity of freshwater assemblages, we a) developed a conceptual model of the possible relationships
123 between environmental change and these three diversity facets in freshwaters, and b) systematically
124 reviewed articles where these relationships have been studied in different freshwater ecosystems. Our
125 study focused exclusively on the investigations of diversity of biological communities where a trait-

126 based, a functional or a phylogenetic index was used to indicate how environmental change has
127 altered freshwater ecosystems. For the systematic review, we specifically investigated which i)
128 biodiversity facets and ii) organism groups have been under investigation, and iii) which
129 environmental stressors (i.e., natural vs. anthropogenic) have impacted freshwater biodiversity. In
130 addition, to provide a general picture of what kind of changes in freshwater biodiversity have already
131 been studied, we highlighted research gaps from the perspectives of organisms, ecosystems, stressors
132 and geographical locations.

133

134 **Local communities, biodiversity patterns and ecosystem functioning**

135

136 In a freshwater community, species functional traits are likely to be more important than species
137 richness in maintaining ecosystem functioning (Mouillot et al. 2012). Papers investigating the
138 relationships between species traits, ecosystem functioning and the environment in freshwaters
139 consider various ecosystems and biological groups (Jones et al. 2002; Vaughn et al. 2007; Bruder et
140 al. 2015). For example, increasing and more frequent drying of river channels is expected due to the
141 climate change (Datry et al. 2017; Mustonen et al. 2018), and Bruder et al. (2011) found that drying
142 influenced both fungal decomposers and the decomposition rate of broad-leaved tree litter. However,
143 most studies on the relationship between freshwater biodiversity and ecosystem functioning have
144 been done using a single species trait or functional groups until recent years, possibly resulting in
145 underestimation of species' roles in ecosystem functions (Vaughn 2010).

146

147 A local freshwater community not only consists of different taxonomic assemblages but also
148 comprises species with various traits. The foundations of a local community come from the global
149 and regional species pools, from which species with suitable traits are filtered by the biotic and abiotic

150 environment to determine species that can successfully colonize and co-exist at a local site (e.g., Poff
151 et al. 1997). In addition, for a given regional species pool, species may respond to environmental
152 gradients in different ways, affecting the distribution of different biodiversity measures over different
153 spatial and temporal scales and generating spatial mismatch among taxonomic, functional and
154 phylogenetic diversities (Devictor et al. 2010).

155

156 Dispersal is an essential natural process influencing local freshwater communities, as well as regional
157 species pools and ecosystem functioning (Fig. 2). Dispersal may mask the importance of
158 environmental conditions affecting local communities, because very high or low dispersal rates may
159 restrict species sorting, disassociating the otherwise strong relationship between local communities
160 and local environmental characteristics (Leibold et al. 2004; Winegardner et al. 2012). In addition to
161 dispersal, speciation-extinction rate is a major relatively long-term driver of local communities that
162 should be acknowledged in order to understand the evolutionary processes driving diversity patterns
163 (Mittelbach and Schemske 2015). Biotic interactions among species, especially competitive
164 interactions, are also important drivers of local community structure that are, at least partly, mediated
165 by species functional traits (Edwards et al. 2011). Ecosystem disturbance often enhances mortality
166 rates and decreases reproduction rates for the species present, causing density-dependent competition
167 to have a weaker effect on taxonomic community structure than on functional community structure
168 (Mouillot et al. 2012). Moreover, global change effects can exclude species with certain traits or
169 strongly decrease their abundance in a community. As a result, trait differences between species can
170 mediate interspecific differences in relation to global change, thus influencing ecosystem functioning
171 in freshwaters (Haddad et al. 2008).

172

173 Global change has also other impacts on local community structure and ecosystem functioning (Fig.
174 2). Climate change affects not only taxonomically-defined communities, but also causes shifts in
175 functional space occupation by driving species with traits poorly fitted to the new environment to
176 extinction (Mouillot et al. 2012). In freshwaters, this would affect especially species having traits
177 suitable for coping with cold climates, where species may be severely affected by climate warming
178 (Heino et al. 2009). Climate change also allows colonization of species with better-fitting traits to
179 remove cold-tolerant species from high-latitude and high-elevation freshwaters (Angeler et al. 2013;
180 Boersma et al. 2016; Garcia-Raventos et al. 2017), showing a negative trend between biodiversity
181 and climate change (Fig. 3). In addition, non-native species can change the functional structure of a
182 given community through altering functional space occupied by native species, for example, through
183 competition (Olden et al. 2006; Mouillot et al. 2012). Although native and non-native species may
184 possess similar functional traits, a competitive advantage may allow non-native species to establish
185 and finally even outcompete native species. Finally, non-native species can function as consumers to
186 diminish native species abundances until they are threatened with extinction (Mouillot et al. 2012).

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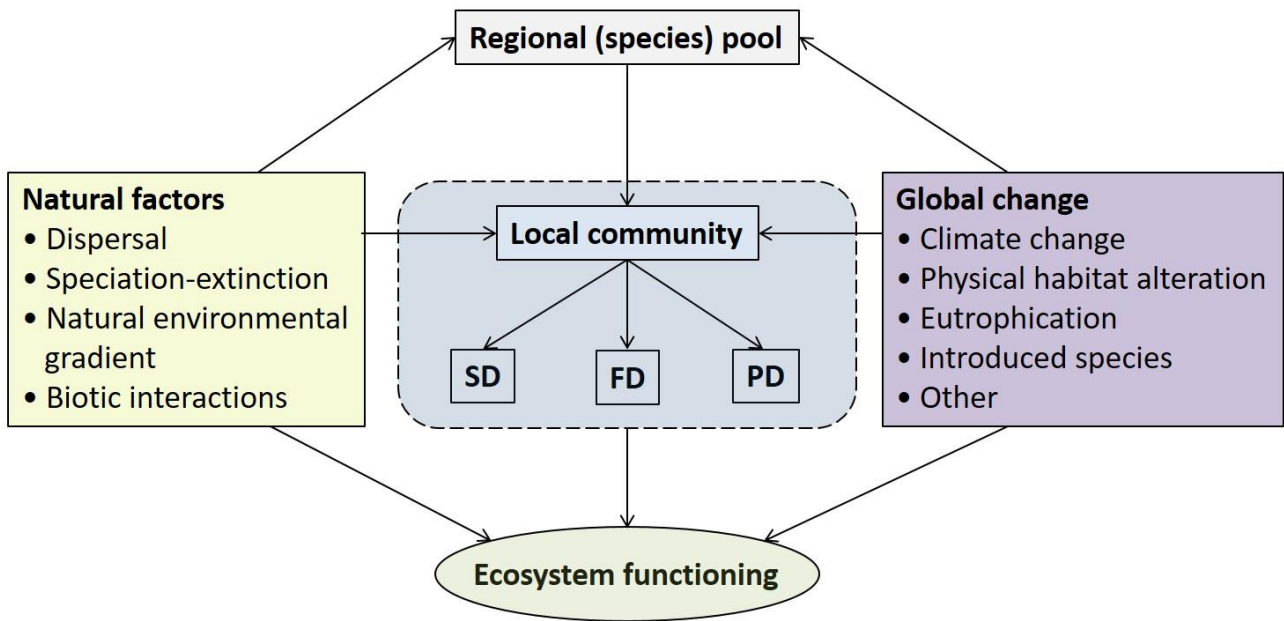
188 Eutrophication is a major problem in many freshwater ecosystems across the world. In addition,
189 climate change likely boosts the harmful effects of eutrophication, because warming temperatures
190 and enhanced carbon dioxide concentrations increase eutrophication symptoms (Moss et al. 2011).
191 As a result, trait-based, functional and phylogenetic diversity are likely to be reduced (Fig. 3), because
192 the combined effects of global change filter out species located in different parts of the functional
193 space or even act additively, leading to rapid extinctions when their effects intersect in functional
194 space (Statzner and Beche 2010; Mouillot et al. 2012). However, the influence of eutrophication
195 likely varies according to the original background ecosystem status (Fig. 3). In mainly oligotrophic
196 systems, the relationship between biodiversity and nutrient enrichment can even be positive (Erős et
197 al. 2009; Leira et al. 2009), whereas mesotrophic freshwaters may show a unimodal response to

198 eutrophication (Nevalainen and Luoto 2016), and a negative relationship is found especially in high-
199 nutrient ecosystems due to competitive exclusion (Peru and Doleddec, 2010; Fernandez et al. 2014).
200 In some cases, biodiversity measures may not respond to the measured and anticipated disturbance,
201 leading to a non-significant relationship. This kind of pattern has especially been found for taxonomic
202 distinctness (e.g., Heino et al. 2007; Vilmi et al. 2016), which has been used as a proxy for
203 evolutionary relationships among species when no true phylogeny is available (Clarke and Warwick
204 2001).

205

206 Physical habitat alterations in freshwaters are typically related to damming of rivers, leading to loss
207 or change of hydrological connections, channelization, water level regulation in lakes and rivers,
208 degradation of the riparian zone by land use along both lakes and rivers, and drought events. As
209 hydrological conditions fundamentally govern the establishment, growth, reproduction, dispersal and
210 extinction of many, if not most, freshwater organisms (Poff et al. 1997), changes in physical habitat
211 have profound effects on biodiversity patterns in freshwaters. Species with poor dispersal abilities
212 and/or intolerant traits against rapid short-term habitat changes are in a jeopardy to be removed from
213 a given freshwater ecosystem suffering from water level fluctuations, and temporally dynamic flood
214 and drought events (Silver et al. 2012; Abgrall et al. 2017). In addition, long-lasting changes in
215 physical habitats due to dam construction or channel modification and destruction of the riparian zone
216 force species to evolve new traits as adaptations to new environmental conditions unless they go to
217 extinct or disperse to new habitats (Bhat and Maguirran 2006; Espanol et al. 2015).

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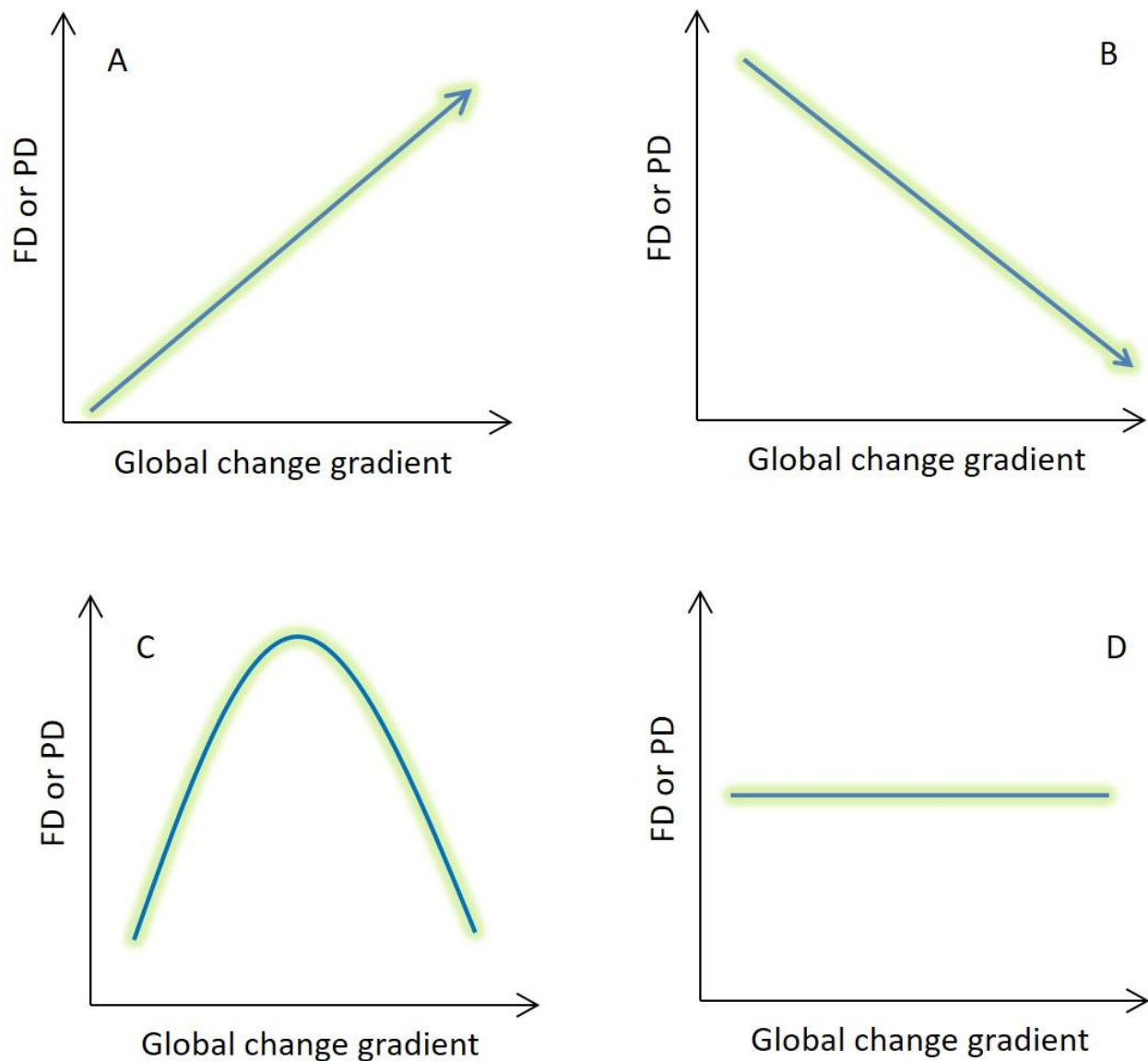


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220 Fig. 2. The relationships between a local community and its environment in relation to ecosystem
 221 functioning. Local communities consist of a subset of species with suitable traits from the regional
 222 (species) pool that have passed through environmental filters (i.e., natural factors and global changes).
 223 Both natural factors and global change affect regional (species) pool, local communities and
 224 ecosystem functioning. SD: species diversity, FD: functional diversity, PD: phylogenetic diversity.

225

226



227

228 Fig. 3. Hypothesised relationships between functional diversity (FD) or phylogenetic diversity (PD)
 229 and an environmental change gradient. Depending on the length of the gradient and geographical
 230 location of study region, these relationships could be different. Environmental change may enhance
 231 diversity in less-disturbed regions situated, for example, in high latitudes (A), where increased
 232 nutrient inputs to freshwaters or higher temperatures can boost functional and phylogenetic diversity.
 233 On the other hand, the relation between diversity and environmental change is often negative in more
 234 human-impacted regions (B), where eutrophication, invasive species or increased temperatures may
 235 strongly affect local (native) communities by decreasing functional and phylogenetic diversity. When

236 the focus is on a full environmental change gradient, such as at global scale, the relationship is
237 expected to be unimodal (C). Functional diversity is first enhanced by increased environmental
238 change effects, but the relationship becomes negative when the environmental change pressures
239 increases. In some cases, environmental changes may not have any detectable influence on functional
240 and phylogenetic diversity (for example in the case of short environmental gradients or when species
241 are functionally redundant), resulting in a non-significant relationship (D).

242

243 **Systematic literature review**

244

245 SELECTION CRITERIA OF SYSTEMATIC REVIEW

246

247 We performed the literature search in the Web of Science (WoS; <http://apps.webofknowledge.com>)
248 using appropriate keywords related to our study topics. We used four kinds of keywords
249 simultaneously: 1) words that describe the trait-based, functional and phylogenetic diversity (funct*
250 OR trait* OR phylogen* OR “taxonomic distinctness”), 2) words related to freshwater habitats
251 (freshwater* OR lentic* OR lotic* OR lake* OR pond* OR stream* OR river* OR wetland* OR
252 spring*), 3) words that are related to diversity (divers* OR biodiv*), and 4) words that indicate
253 environmental change (environment* OR "climate change" OR eutrophication OR acidification OR
254 "habitat loss" OR "nutrient enrichment" OR "global change" OR “climate warming” OR invasive*
255 OR exotic* OR alien* OR urbanization OR pollution OR drought OR channelization). TITLE was
256 selected for the row describing trait-based, functional and phylogenetic diversity words, whereas
257 TOPIC was selected for all other rows. [Trait-based diversity and functional diversity do not mean the](#)
258 [same thing, as the former term is more inclusive than the latter, and the latter should only include](#)
259 [traits that really affect ecosystems functions](#). In practice, both terms have been extensively used in

260 the literature, often also interchangeably. The use of TITLE in other rows would have strongly
261 narrowed the number of potential articles in our search exercises that may have resulted to exclusion
262 of some matching papers. We did not have any temporal limitation in our search but all the possible
263 articles matching our criteria were selected. The main search for suitable articles was executed on 13
264 April 2017, followed by complementary searches done in 13 February 2018 and 21 September 2018
265 to account for all published articles in year 2017 and to include channelization as an additional
266 environmental change keyword, respectively. This extensive search protocol resulted in a total of
267 1475 results found. After the main WoS literature search, all authors were given an equal number of
268 articles to go through and select suitable articles matching our study scope. The first author selected
269 suitable articles from the complementary search effort of year 2017. The first and the last author
270 together double-checked all the selected articles to ensure uniformity and objectivity in the selection
271 process. We included articles that reported results for freshwater ecosystems and covered the effects
272 of environmental change on trait-based, functional and phylogenetic diversity of community-based
273 data through different indices. Instead, we excluded articles that used a space-for-time substitution to
274 illustrate, for example, the effects of global warming, articles that tested ecological theories only,
275 articles that did not have any clear stressors, purely predictive articles, review articles or conference
276 abstracts. These types of articles were common among the initial WoS search results, but they were
277 removed from the final selection. We stress that articles dealing with biological compositions
278 distinguished to functional groups or assemblages did not meet our criteria, because we focussed
279 purely on different *indices* used to characterize trait-based, functional and phylogenetic diversity of
280 freshwater organisms. Thus, articles dealing with grouping of species based on their traits or
281 functional properties (e.g., functional feeding groups of macroinvertebrates or growth forms of
282 macrophytes) and based often on ordination methods only did not pass our selection criteria. Articles
283 lacking clear statements of results were neither included in the final set of articles. All authors
284 collected information from articles that were likely suitable for comparative purposes (Table S1). The

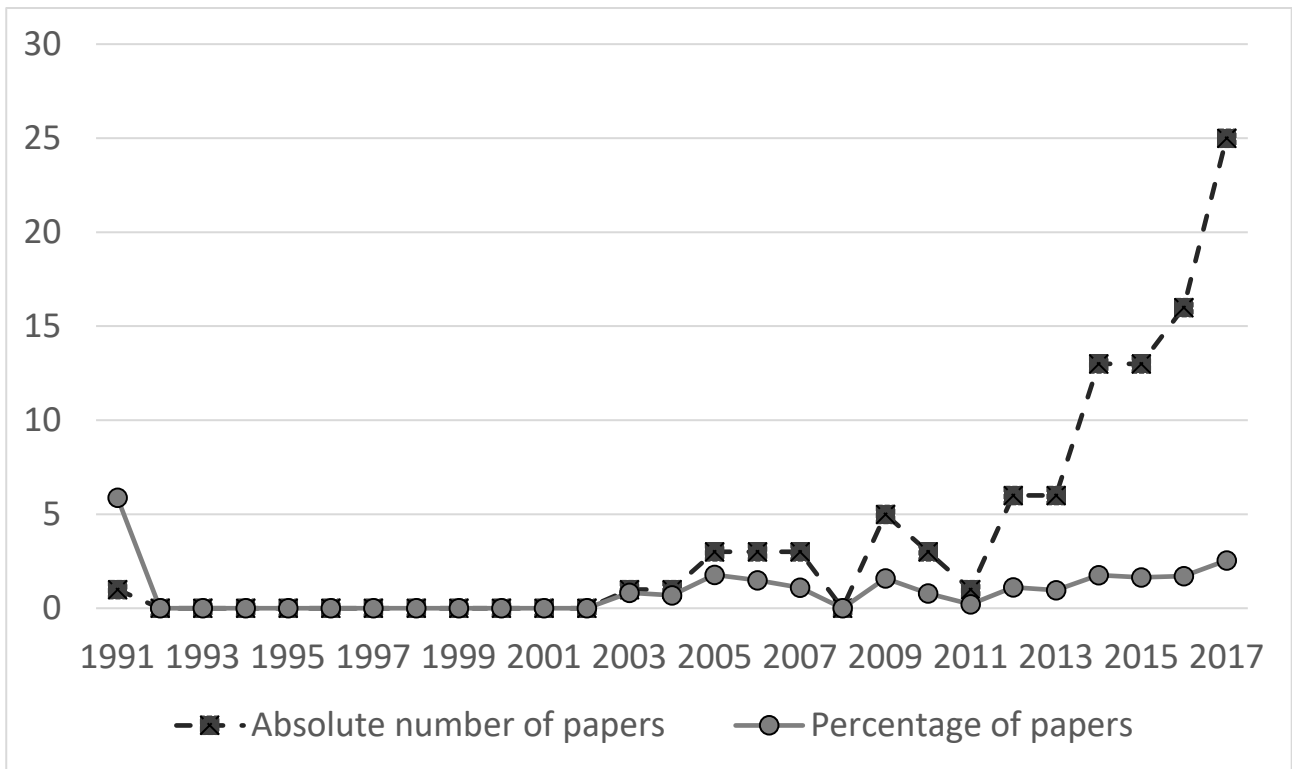
285 first author re-checked all the collected information to guarantee data quality, and we formed a
286 number of categories from the different variables. These included, for example, five groups of
287 organisms (i.e., macroinvertebrates, fish, macrophytes, bacteria, diatoms and other taxa; see Fig. 5),
288 four main stressors (i.e., eutrophication, physical habitat alteration, non-native species and climate
289 change, and their joint effects; see Fig. 6), and direction of stressor effect (i.e., no effect, increasing,
290 decreasing and multiple responses). Finally, the first author compiled a consistent dataset including
291 main information and variables from the final set of 100 selected articles matching our strict inclusion
292 criteria (Table S2).

293

294 MAIN FINDINGS FROM THE SYSTEMATIC REVIEW

295

296 Our systematic review on the trait-based, functional and phylogenetic diversity measures of
297 freshwater communities revealed that the first papers (beyond single ones) were published in 2003
298 (Fig. 4). Although a clear increase in the absolute numbers of papers was detected after 2011, there
299 was no increasing pattern in the proportion of papers in relation to similar studies executed in
300 terrestrial and marine systems (based on the similar WoS search but freshwater habitats as TOPIC
301 were excluded from the search). This suggests that findings on these community-based diversity
302 measures published in journals with general ecological foci have reached freshwater and
303 terrestrial/marine ecologists only relatively recently. Modern well-recognized papers on community-
304 based functional ecology were published in mid-2000s (e.g., McGill, Enquist, Weiher, & Westoby,
305 2006; Petchey, & Gaston, 2006; Vileger, Mason, & Mouillot, 2008), and freshwater ecologists have
306 found these measures relatively well.



307

308 Fig. 4. Absolute and percentage (absolute number of selected papers in relation to all papers dealing
 309 with environmental change and functional, trait-based and phylogenetic diversity in terrestrial and
 310 marine systems) changes in the number of articles published that focus on the relationship between
 311 environmental change and functional, trait-based and phylogenetic diversity in freshwaters over the
 312 years based on our selection criteria (see Selection criteria of systematic review).

313

314 The systematic review revealed that various different measures of trait-based, functional and
 315 phylogenetic diversity have been used in the freshwater research over the years. The most common
 316 measures were functional richness, functional evenness, functional divergence and taxonomic
 317 distinctness. Beside these indices, various other approaches were used including the following: trait
 318 diversity or number of trait combinations (e.g., through community-weighted mean), phylogenetic
 319 diversity, Rao's quadratic entropy and functional beta diversity. The majority of the rarely-used
 320 measures were used only in a single study.

321

322 Considering different organism groups, macroinvertebrates were the most studied group utilised in
323 half of the selected papers when investigating the relationship between functional, trait-based or
324 phylogenetic diversity and the environment (Fig. 5; Fig. S1). Functional diversity was the most
325 widely-used approach for macroinvertebrates (in 34 papers out of 47 macroinvertebrate papers),
326 followed by phylogenetic diversity studied in nine papers (Fig. S2). After the introduction of
327 taxonomic distinctness index as a proxy of phylogeny (Clarke and Warwick 2001), there were several
328 papers published where taxonomic distinctness of macroinvertebrates was correlated with
329 environmental variables (e.g., Abellan et al. 2006; Heino et al. 2007; Alahuhta et al. 2017a).
330 Macroinvertebrate studies were mostly done in lotic systems (33 out of 47) and were relatively
331 equally distributed among different years and continents where they had been investigated. Fish were
332 the second most studied organism group (20 out of 100) with 85% of the papers focussed on rivers
333 and streams. Similar to macroinvertebrates, functional diversity was the most studied index (16 out
334 of 20), and fish studies were found from different years and studied continents (e.g., Pool and Olden
335 2012; Matsuzaki et al. 2016; Sagouis et al. 2017). Bacteria, diatoms and macrophytes were each
336 investigated in ca. 10% of selected papers. For macrophytes and diatoms, functional diversity was
337 the most studied measure (six out of 10 and nine out of 13, respectively), whereas both functional and
338 phylogenetic diversity were solely used for bacteria. Compared to the other freshwater assemblages,
339 phylogenetic diversity studies on bacteria have been based on true phylogeny instead of proxy
340 measures (e.g., Barberan and Casamayor 2014). Bacteria, diatoms and macrophytes were mostly
341 investigated in lakes and ponds (six out of nine, 11 out of 13 and eight out of ten, respectively), but
342 also some river and stream studies have appeared. All of the three organism groups have been under
343 research mostly in North America, South America, Europe and China during the 2010s. Temporal
344 aspects were considered in ca. 30% of all selected papers, ranging from phylogenetic diversity of
345 stream macroinvertebrates in relation to damming (Campbell and Novelo-Gutierrez 2007) and

346 measuring the effects of climate change on functional resilience of multiple taxa in subarctic lakes
347 (Angeler et al. 2013) to temporal changes in nutrient enrichment on macroinvertebrate functional
348 diversity in boreal lakes (Nevalainen and Luoto 2017).

349

350 Biodiversity measures of different organism groups responded differently to environmental stressors
351 (Fig. S1). For macroinvertebrates (20 out of 47 studies), fish (11 out of 20), diatoms (nine out of 13)
352 and macrophytes (five out of 10), 'multiple effects' were the most common relationship between the
353 biodiversity measure and the stressor(s). On the other hand, all stressor types were equally common
354 in studies of bacterial biodiversity. Considering biodiversity measures across organism groups, the
355 typical relationship was that functional diversity showed multiple relationships with eutrophication
356 and physical habitat alteration (Fig. 6). These two stressor types were also the most studied both
357 separately and jointly. Instead, climate change and non-native species were studied only in less than
358 six percentage of the papers each. This is a rather alarming finding considering the multiple and
359 additive impacts climate change has been predicted to have on freshwater systems (Heino et al. 2009;
360 Moss et al. 2011). Climate change (two out of three), physical habitat alteration (22 out of 42) and
361 eutrophication (31 out of 52) most commonly showed multiple effects on biodiversity measures,
362 whereas only non-native species showed mainly negative influences on the biodiversity (four out of
363 seven). Physical habitat alteration quite often also decreased trait-based, functional and phylogenetic
364 diversity in the freshwater realm (11 out of 42). The effects of degradation of habitat conditions and
365 non-native species are often straightforward and direct in freshwater ecosystems that is why the
366 responses of biodiversity measures to these two environmental changes were negative more often
367 compared to other environmental change stressors (Campbell and Novelo-Gutierrez 2007; Liu et al.
368 2013; Matsuzaki et al. 2016). On the contrary, the influence of eutrophication and climate change on
369 ecosystem functioning is typically more multidimensional, having contradictory and often cumulative
370 effects on different organism groups and food chain levels (Leira et al. 2009; Angeler et al. 2013;

371 Boersma et al. 2016; Vilmi et al. 2016). In addition, functional diversity often consists of several
372 indices (e.g., functional richness, evenness and divergence) that show variable responses to the
373 environment (Petchey and Gaston 2006; Mouillot et al. 2012), resulting in the multiple effects
374 detected between biodiversity and environmental change. Interestingly, however, human-induced
375 stressors more often decreased biodiversity (18 out of 42), whereas natural stressors had frequently
376 various effects (i.e., multiple, increasing or no effect) on the studied biodiversity indices. In the
377 examples of decreased biodiversity due to global change, functional diversity was typically lower in
378 impacted sites than in reference water bodies or reduced over time (Liu et al. 2013; Matsuzaki et al.
379 2016).

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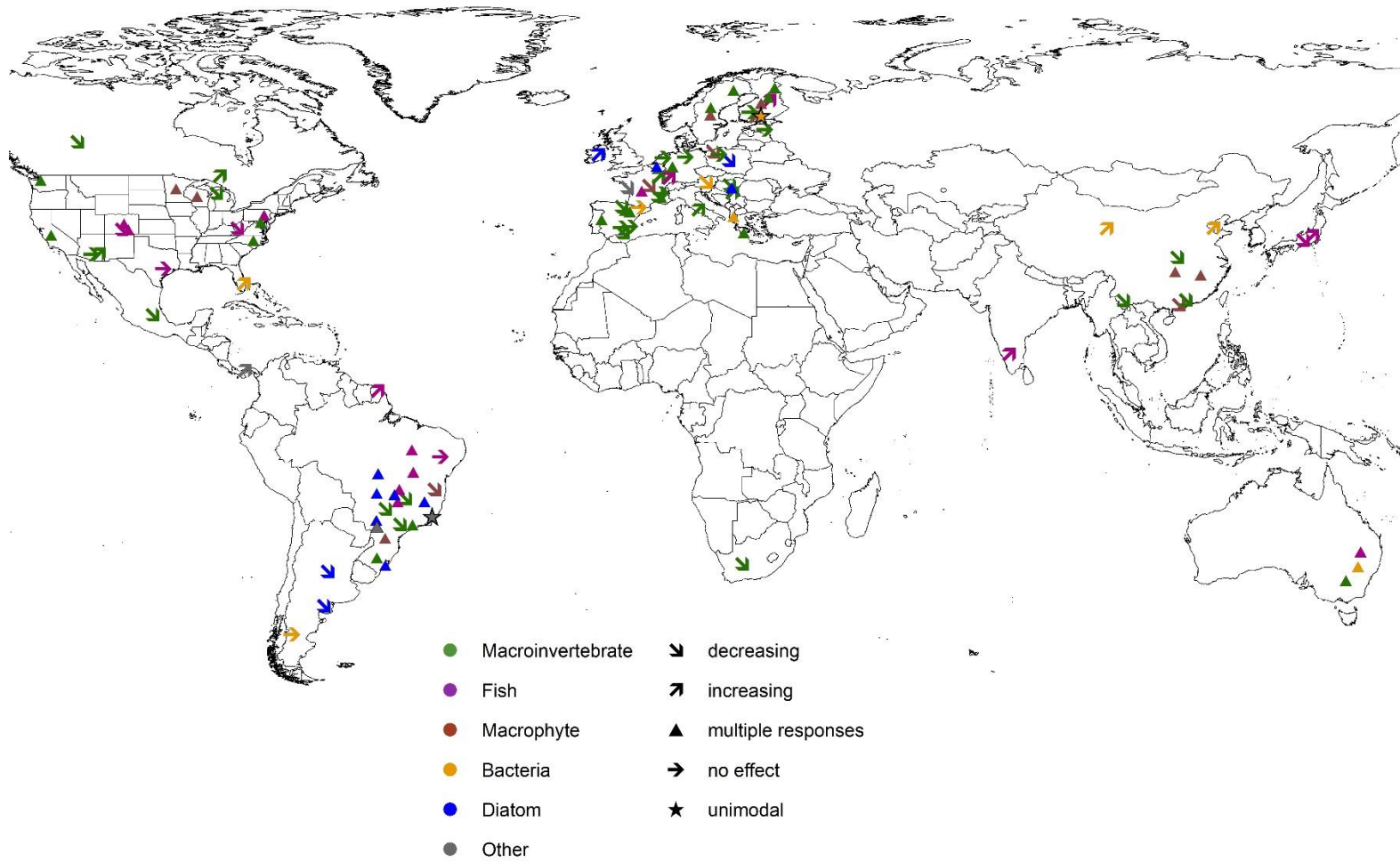
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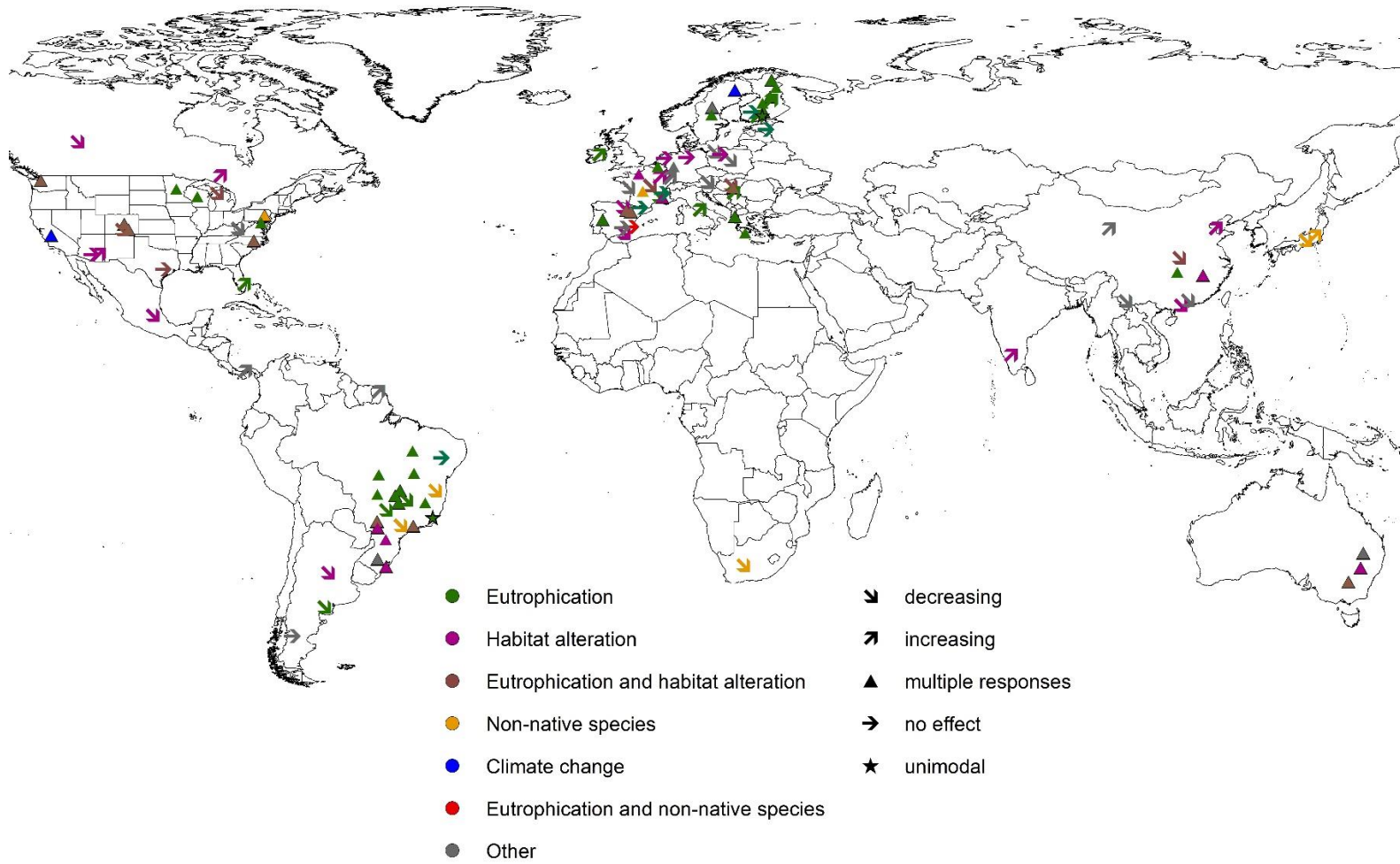
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386

387 Fig. 5. A map illustrating the biological groups used to study the relationship between trait-based, functional and phylogenetic diversity and the
 388 direction of effect caused by environmental change effects based on our systematic review in the freshwater realm (n=100).



389

390 Fig. 6. A map illustrating the relationship between specific environmental change stressor and the direction of effect found in the different
 391 articles (n=100) selected in the systematic review.

392 **Environmental change drives biodiversity patterns in freshwaters, but there is**
393 **no single type of biodiversity response to the change**

394

395 Environmental change, including both natural and human-induced environmental aspects, is driving
396 trait-based, functional and phylogenetic diversity in global freshwater ecosystems. However, it seems
397 that it is more difficult to find clear relationships between the biodiversity measures and the
398 environment when strong natural gradients are involved in a study. We found that biodiversity indices
399 often had multiple relationships with the environment, especially in cases when both natural and
400 anthropogenic characteristics were investigated in the same study or only natural environmental
401 change was under examination. For example, functional dispersion and functional evenness of fish
402 assemblages were driven by multiple environmental factors related to both natural and anthropogenic
403 gradients in Australian river basins (Stenberg et al. 2014). Similarly, two measures of taxonomic
404 distinctness of diatoms, macrophytes and macroinvertebrates showed opposite responses to total
405 phosphorus and nitrogen gradients in a large boreal lake (Vilmi et al. 2016). Previous exercises
406 regarding taxonomic distinctness have evidenced this situation for different freshwater organism
407 groups in various regions. For example, Bhat and Magurran (2006) first reported that the indices of
408 phylogenetic relatedness may be masked by influences of habitat variability on fish species
409 compositions in India. Subsequently, other studies have found that natural environmental
410 characteristics may overshadow the influences of anthropogenic pressures on taxonomic distinctness
411 (Heino et al. 2007; Alahuhta et al. 2017a). In addition, the performance and ability to detect human-
412 induced stress of taxonomic distinctness may depend on the phylogenetic structure of surveyed taxa
413 within a study region, as well as their evolutionary and ecological history (Abellan et al. 2006). These
414 findings are important because taxonomic distinctness measures should be independent of natural
415 environmental gradients and sampling effort (Clarke and Warwick 2001). Our systematic review
416 emphasises that biodiversity measures should be interpreted with caution in the situations where the

417 purpose is to quantify natural environmental changes (separately or together with anthropogenic
418 perturbations) in freshwater ecosystems.

419

420 Although the natural environmental characteristics create complexity to the freshwater ecosystems
421 and challenge ecologists in how to portray ecosystem functioning, we also found promising examples
422 of studies where diversity measures responded to anthropogenic disturbance in a predicted way (i.e.,
423 negatively; Arthaud et al. 2012; Liu et al. 2013; Matsuzaki et al. 2016). The relationship between
424 biodiversity and ecosystem functioning is assumed to be linearly positive, but global change effects
425 may disturb this relationship (Woodward et al. 2010; Cadotte et al. 2011). In the examples we found,
426 a single human-induced stressor was correlated with biodiversity, producing a decreasing trend. For
427 instance, increased water level led to decline in functional diversity of macrophytes in a subtropical
428 reservoir compared to that of adjacent wetlands (Liu et al. 2013), whereas urbanization reduced
429 functional diversity of aquatic insects in Neotropical streams (Gimenez and Higuete 2017). In the other
430 study, introduction of non-native fish species decreased functional diversity of native fish
431 assemblages over time (Matsuzaki et al. 2016). However, multiple global change effects can act
432 simultaneously in influencing ecosystem functioning, such as in the case of climate warming and
433 eutrophication in freshwaters. The joint effects of different global change factors are likely to decrease
434 strongly overall species richness and trait diversity by filtering out species not only located in different
435 parts of the functional space but also acting additively, or even acting in synergy, leading to rapid
436 extinctions when the effects of the stressors overlap in functional space (Mouillot et al. 2012). For
437 example, Olden et al. (2006) found that native fish communities experienced two shared pressures
438 mediated by functional traits: species were filtered out due to either vulnerable traits associated with
439 environmental changes or competition with exotic species sharing similar traits. This further
440 complicates our attempts to investigate how global change affects biodiversity and, subsequently,
441 ecosystem functioning.

442

443 **Research gaps and future study directions**

444

445 We have demonstrated a link between trait-based, functional and phylogenetic diversity and
 446 environmental change in freshwater ecosystems through the conceptual model and the systematic
 447 review. The latter also offered us details on the current research status and knowledge gaps. Next, we
 448 presented gaps in the knowledge of the relationship between freshwater biodiversity and
 449 environmental change, and suggested where the future research efforts should focus. The research
 450 gaps are related to a low number of biodiversity studies, species dispersal, lack of information on true
 451 phylogenies, niche conservatism of species traits, lack of data on species functional traits,
 452 understudied organism groups and global change stressors, geographical biases in research, [and lack
 453 of summarized information how restoration affects the relationships between trait-based, functional
 454 and phylogenetic diversity and environmental change](#) (Table 1).

455

456 Table 1. Summary of the known research gaps and suggestions for possible future research directions
 457 based on our systematic review on trait-based, functional and phylogenetic biodiversity of freshwater
 458 organism groups.

Research gap	Suggestion for future study direction
• Low number of community-based studies	→ More studies on the trait-based, functional and phylogenetic biodiversity as related to environmental change are required.
• Species dispersal	→ Alternative methods (e.g., dispersal proxies such as different distance metrics) to account for dispersal in multi-species communities is needed.
• Biotic interactions	→ Biotic interaction measures (e.g., Joint Species Distribution Models) should be included in future studies
• Lack of phylogenetic information	→ True phylogenies of freshwater organisms are desperately required and/or development of additional phylogeny proxies are needed.

<ul style="list-style-type: none"> • Conservatism of species traits • Lack of information on species functional traits • Understudied organism groups • Understudied global change stressors • Geographical bias in research • How restoration affects trait-based, functional and phylogenetic diversity 	<ul style="list-style-type: none"> → Conservatism of species traits needs to be evaluated for different organism groups before phylogeny can be used as a proxy for functional diversity → More research focus should be devoted to functional species traits and how they are actually related to freshwater ecosystem functioning. → More investigations especially on the biodiversity of macrophytes, diatoms, other algae and bacteria are needed → Studies are required on the effects of climate change and non-native species on different freshwater organism groups. → Additional studies from Africa, Southern Asia and Russia are needed. → Review whether restoration affects the relationships between trait-based, functional or phylogenetic diversity and environmental change
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459

460 We surprisingly found only 100 papers out of 1475 (7%) matching our selection criteria. The majority
461 of the papers in the initial selection phase concerned studies with space-for-time substitutions, testing
462 of ecological theories only, without any specific stressors, with purely predictive purposes, with
463 single species only and without original peer-reviewed contribution (i.e., review and conference
464 abstract). Fortunately, there has been a clear increase in the absolute number of published papers
465 during the past couple of years (Fig. 4), suggesting that community-based studies on the relationship
466 between biodiversity and environmental change are building up. This is an encouraging trend because
467 the species traits of biological community rather than that of, for example, a single species influence
468 the ecosystem functioning (Flynn et al. 2011; Mouillot et al. 2012).

469

470 One of the hot subjects in freshwater ecology is how dispersal may affect local communities (Heino
471 et al. 2015). The importance of dispersal is highlighted in the differently-connected freshwater
472 systems, including organisms with different dispersal abilities. Dispersal interacts with environmental

473 change so that anthropogenic disturbance affects poorly dispersing organisms more severely than
474 species with efficient dispersal traits, because poorly dispersing organism cannot track variation in
475 environmental changes as rapidly as strong dispersers. In addition, organisms in isolated freshwater
476 systems (e.g., springs, ponds, and lakes) are likely to be more strongly impacted by the joint effects
477 of limited dispersal and anthropogenic disturbance than those in more continuous ecosystems (e.g.,
478 streams and rivers) (Soininen 2014), but more research is needed to assess this idea further. Our
479 systematic review revealed that dispersal was rarely included, if at all, in the study of the biodiversity
480 measures considered. For example, in the partitioning of functional beta diversity, dispersal limitation
481 was the principal force structuring tropical fish assemblages due to low functional turnover (Cilleros
482 et al. 2016). Although passively moving organisms with small propagules (e.g. macrophytes, diatoms,
483 bacteria) could be expected to be less dispersal limited than actively dispersing large species (e.g.
484 macroinvertebrates and fish), increasing amount of evidence suggest a low level of congruence
485 among the findings of freshwater studies. However, conflicting results suggest (De Bie et al. 2012;
486 Soininen 2014) that freshwater organisms' dispersal depends on biological group, region and spatial
487 scale under study, as well as their combinations, and thus different ways to determine dispersal for
488 these case-specific situations are required (Heino et al. 2017).

489

490 Biotic interactions among species in a community can also strongly affect diversity measures. We
491 found that only in one study biotic interactions were accounted for in freshwaters though they were
492 not important predictors of functional diversity of stream fish in a semiarid region of Brazil
493 (Rodrigues-Filho et al. 2017). Recently emerged statistical tools of Joint Species Distribution
494 Modelling (JSDM) may offer valuable assistance in including species interactions to the models (e.g.,
495 Pollock et al. 2014). [At the moment, different JSDM methods are emerging, with the basic difference](#)
496 [whether direction of interaction is available or not.](#) Inclusion of biotic interactions to the diversity
497 models may also partly overcome low explained variations often found for freshwater communities.

498

499 In addition to the dispersal and biotic interaction proxies, comprehensive and true phylogenies rarely
500 exists for most of freshwater organism groups. The only biological group for which comprehensive
501 evolutionary history has often been revealed through DNA analysis is bacteria (Barberan and
502 Casamayor 2014). As demonstrated in our review, the majority of freshwater studies on PD has been
503 based on proxies for true phylogeny, such as taxonomic distinctness (Clarke and Warwick 2001).
504 However, these phylogeny proxies have not managed to quantify the relationship between
505 phylogenetic diversity and environmental change very well. Thus, we advise researchers to determine
506 the true phylogeny of freshwater assemblages, if possible, or develop alternative proxies for
507 phylogenetic diversity. These possible proxies should be able to function properly in complex
508 situations of natural and anthropogenic environmental effects on phylogenetic diversity, so that
509 different effects can be distinguished.

510

511 Phylogeny can be used as a proxy for functional diversity if the species traits considered are
512 phylogenetically conserved (Flynn et al. 2011). We found that the influence of niche conservatism
513 on the species traits was explicitly considered in two selected papers out of 27 studying phylogenetic
514 diversity. Carvajal-Castro and Vargas-Salinas (2016) assessed whether male body size and call
515 frequency of Neotropical anuran assemblages were conserved, and found a strong phylogenetic
516 signal. In another work, trait conservatism was evidenced only at short phylogenetic distances for
517 stream fungi (Mykrä et al. 2016). In the very few published papers of niche conservatism for
518 freshwater realm beyond our review, a significant phylogenetic signal was discovered for many of
519 the ecological optima of 217 diatom species (Keck et al. 2016), and thermal tolerances and
520 acclimation capacity of 82 fish species (Comte and Olden 2016). However, the strength of the signal
521 has varied or even lacked among the studied species and species traits (Litsios et al. 2012; Keck et al.
522 2016). Moreover, climate niches did but local niches did not suggest niche conservatism for lake

523 macrophytes in relation to their geographical distributions (Alahuhta et al. 2017b). These findings
524 indicate that niche conservatism in the freshwater realm should be more closely examined for species
525 traits before we can reliably use phylogeny as a proxy for trait-based or functional diversity for
526 freshwater organism groups.

527

528 Although other diversity measures (i.e., trait-based and functional diversity) were under intensive
529 research, the species traits used are not necessarily related to ecosystem functioning. Schmera et al.
530 (2017) reviewed functional diversity measures of macroinvertebrates and found that none of the
531 published papers actually quantified any ecosystem functioning. Instead, the reviewed publications
532 were focussed purely on perspectives of biodiversity that may affect ecosystem functions in general
533 (Schmera et al. 2017). Similar to their study, ecosystem functioning was investigated only in a
534 relatively few papers in our systematic review. For instance, the relationship between phylogeny of
535 methanogen bacteria and eutrophication were studied in the Florida Everglades (Castro et al. 2004).
536 In a second work on bacteria, ecologists investigated if an increase in water temperature would
537 influence heterotrophic metabolic activities of biofilms grown under light or dark conditions (Romani
538 et al. 2014). In a third example, linking primary producers to consumers, functional composition of
539 plant communities had a central role in structuring Collembola assemblages along a flood gradient
540 (Abgrall et al. 2017). Lack of species traits related to pure ecosystem functions may also be related
541 to a rather slow emergence of species trait databases including information on freshwater assemblages
542 especially for less-studied organism groups (see also Fig. 4). This general finding on the small number
543 of papers studying actual ecosystem functions emphasises that more efforts should be devoted to the
544 validation and development of freshwater species traits and investigations of true ecosystem
545 functions. In addition, state-of-art modelling tools (e.g., [gap filling of species trait database](#), Schrodte
546 et al. 2015) may offer help in building more comprehensive species trait databases for freshwater
547 assemblages, especially when studying broad-scale patterns.

548

549 Moreover, there is currently a consensus on which measures should be determined when ecosystem
550 functioning effects are assessed using functional diversity measures. Functional richness, evenness
551 and divergence have been identified as complementary indices to account for different aspects of
552 functional diversity affecting ecosystem functioning (Villegger et al. 2008; Mouchet et al. 2010). Our
553 systematic review revealed that these three functional diversity approaches have been the main foci
554 of freshwater ecologists only in the past couple of years. Although the use of several biodiversity
555 indices inevitably leads to increasing ‘multiple response effects’, we urge scientists for the sake of
556 comparability among different studies to continue to use at least these three elements of functional
557 diversity in the future studies on freshwater ecosystems.

558

559 Macroinvertebrates and fish were the biological groups investigated in most freshwater diversity
560 studies, covering 65% of all the selected studies. For the other biological groups, including
561 macrophytes, diatoms and bacteria, there were much fewer investigations. More research is needed
562 on these understudied biological assemblages to gain more profound understanding on the
563 relationship between biodiversity and environmental change.

564

565 To our surprise, climate change and non-native species were clearly less widely investigated than
566 other global change stressors. This is rather alarming considering that climate change likely severely
567 affects freshwater biodiversity and ecosystem functioning (Heino et al. 2009; Moss et al. 2011;
568 Jourdan et al. 2018). Moreover, the majority of climate change studies have focussed on individuals
569 or species populations, instead of entire communities and whole ecosystems (Woodward et al. 2010).

570

571 We also found a geographical bias in the published literature, as Europe, North America, South
572 America and China were the dominant study regions. Because the evidence seems to suggest that the
573 correlations between freshwater diversity and environmental change are dependent on a study region
574 and the background characteristics of those regions, more research is required from poorly studied
575 regions, such as Africa, Southern Asia and Russia. However, we acknowledge that freshwater
576 diversity in relation to the environmental change has been investigated especially in Russia but results
577 from these studies have not reached English-language dominated contemporary scientific literature.

578

579 Our review focussed on the relationships between trait-based, functional or phylogenetic diversity
580 and environmental change in freshwater ecosystems. Another important aspect would be to
581 investigate how restoration affects these relationships. Environmental change can be seen as a cause
582 of deterioration, whereas restoration is a desirable means, with which global change impacts on trait-
583 based, functional and phylogenetic biodiversity are repaired close to an original or a desirable state.
584 This topic is beyond our present review, but we urge other scientists to summarize how restoration
585 affects ecosystem functioning measured using these diversity indices as proxies (see e.g. Collier,
586 2017).

587

588 Finally, trait-based, functional and phylogenetic diversity measures not only provide basic scientific
589 knowledge on how environmental change affects freshwater biodiversity and ecosystem functioning,
590 but also act as early warning signals of the intensifying global change effects in the vulnerable
591 freshwater ecosystems. This is because they can possibly a priori be used to detect disturbance
592 impacts before species loss and extinctions actually take place (Mouillot et al. 2012). In addition,
593 freshwaters as vulnerable sentinel systems can provide early warnings of wider-scale environmental
594 change across different ecosystems (Woodward et al. 2010). Lastly, the biodiversity measures we

595 considered can help us 1) to detect which ecosystem functions should be monitored in freshwater
596 bioassessment, 2) whether the restoration of freshwater systems has actually revived valuable
597 ecosystem functions, and 3) whether protected areas are conserving different facets of biodiversity
598 and ecosystem functioning in addition to taxonomic diversity (e.g., Saito et al. 2015). We hope that
599 our current review will stimulate more research on the less well-known facets and topics of
600 biodiversity in highly vulnerable freshwater ecosystems.

601

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609

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873 **Supporting Information**

874 Fig. S1.

875 Fig. S2.

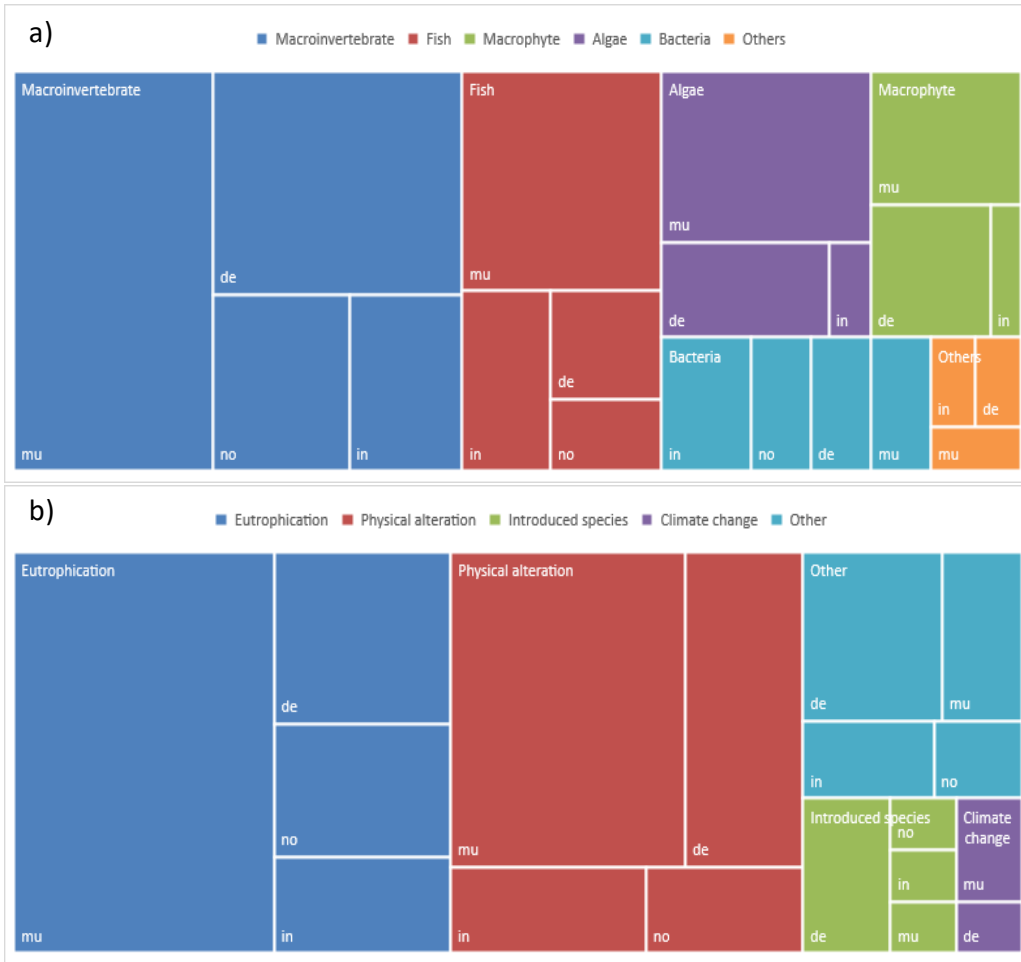
876 Table S1.

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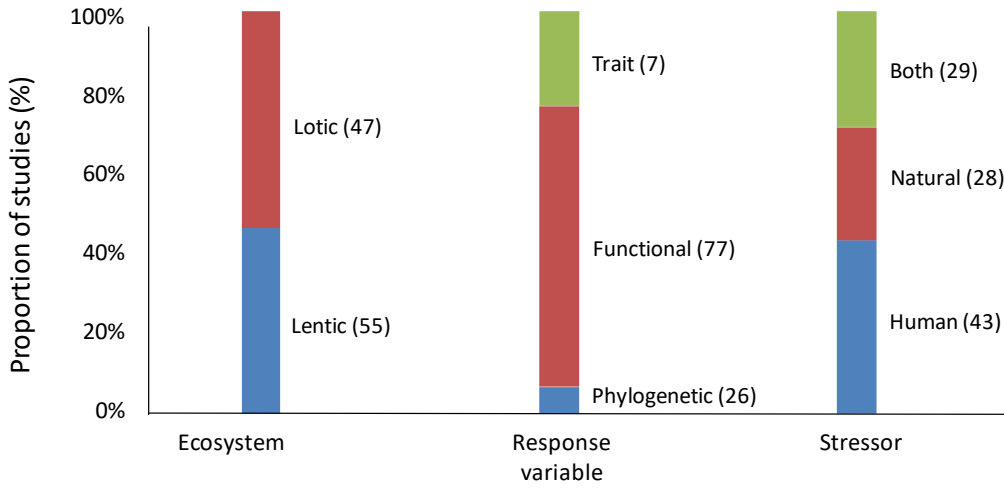
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882 Fig. S1. Tree diagrams showing (a) whether environmental change has had an increasing (in),
883 decreasing (de), multiple (mu) or no (no) effect on different freshwater organism groups, and (b)
884 whether different environmental change effects had increased, decreased, multiple or no effect on
885 freshwater trait-based, functional and phylogenetic diversity. The size of a rectangle is proportional
886 to the number of studies considered in the systematic review. Organism groups in the “Other”
887 comprise of birds, frogs and anuran assemblages.

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892 Fig. S2. Proportion of studies in different ecosystems (a), based on different response variables (b),
893 and focusing on different stressors (c). The numbers within the bars refer to the number of studies.

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- authors
 - title
 - journal
 - publ. year
 - country/place
 - latitude
 - longitude
 - spatial scale
 - temporal focus (contemporary/historical/paleo)
 - observational/experimental
 - data collected (year(s))
 - ecosystem (lotic vs. lentic)

 - pristine (yes/no)
 - number of sites
 - taxonomic group(s)
 - tax group
 - response variable(s)
 - stressor (natural/human)
 - specific stressor
 - statistical methods

 - number of species/taxa/OTUs
 - effect (increasing/decreasing/U-shaped/hump-shaped/no effect/not applicable/multiple responses)
 - main findings
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- temporal variation (yes/no)
 - spatial variation (yes/no)
 - descriptive/predictive/both
 - extra information
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915 Table S2. List of included final articles based on our selection criteria with each article's author(s), title, journal and publication year given.

Authors	Title	Journal	Publication year
Azzanti, M	Sandy bottom macroinvertebrates in two moderately polluted stations of the River Treia (Central Italy): structural and functional organization	ANNALES DE LIMNOLOGIE-INTERNATIONAL JOURNAL OF LIMNOLOGY	1991
Ross, RM; Bennett, RM; Snyder, CD; Young, JA; Smith, DR; Lemarie, DP	Influence of eastern hemlock (<i>Tsuga canadensis</i> L.) on fish community structure and function in headwater streams of the Delaware River basin	ECOLOGY OF FRESHWATER FISH	2003
Castro, H; Ogram, A; Reddy, KR	Phylogenetic characterization of methanogenic assemblages in eutrophic and oligotrophic areas of the Florida Everglades	APPLIED AND ENVIRONMENTAL MICROBIOLOGY	2004
Devin, S; Beisel, JN; Usseglio- Polatera, P; Moreteau, JC	Changes in functional biodiversity in an invaded freshwater ecosystem: the Moselle River	HYDROBIOLOGIA	2005
Heino, J	Functional biodiversity of macroinvertebrate assemblages along major ecological gradients of boreal headwater streams	FRESHWATER BIOLOGY	2005
Heino, J; Soininen, J; Lappalainen, J; Virtanen, R	The relationship between species richness and taxonomic distinctness in freshwater organisms	LIMNOLOGY AND OCEANOGRAPHY	2005
Abellan, P.; Bilton, D. T.; Millan, A.; Sanchez- Fernandez, D.; Ramsay, P. M.	Can taxonomic distinctness assess anthropogenic impacts in inland waters? A case study from a Mediterranean river basin	FRESHWATER BIOLOGY	2006
Bhat, A; Magurran, AE	Taxonomic distinctness in a linear system: a test using a tropical freshwater fish assemblage	ECOGRAPHY	2006
Salas, F; Patricio, J; Marcos, C; Pardal, MA; Perez-Ruzafa, A; Marques, JC	Are taxonomic distinctness measures compliant to other ecological indicators in assessing ecological status?	MARINE POLLUTION BULLETIN	2006
Campbell, WB; Novelo- Gutierrez, R	Reduction in odonate phylogenetic diversity associated with dam impoundment is revealed using taxonomic distinctness	FUNDAMENTAL AND APPLIED LIMNOLOGY	2007
Heino, Jani; Mykra, Heikki; Hamalainen, Heikki; Aroviita, Jukka; Muotka, Timo	Responses of taxonomic distinctness and species diversity indices to anthropogenic impacts and natural environmental gradients in stream macroinvertebrates	FRESHWATER BIOLOGY	2007
Marchant, Richard	The use of taxonomic distinctness to assess environmental disturbance of insect communities from running water	FRESHWATER BIOLOGY	2007

Eros, T; Heino, J; Schmera, D; Rask, M	Characterising functional trait diversity and trait-environment relationships in fish assemblages of boreal lakes	FRESHWATER BIOLOGY	2009
Gallardo, B; Gascon, S; Cabezas, A; Gonzalez, M; Garcia, M; Comin, FA	Relationship between invertebrate traits and lateral environmental gradients in a Mediterranean river-floodplain	FUNDAMENTAL AND APPLIED LIMNOLOGY	2009
Gallardo, B; Gascon, S; Garcia, M; Comin, FA	Testing the response of macroinvertebrate functional structure and biodiversity to flooding and confinement	JOURNAL OF LIMNOLOGY	2009
Leira, M.; Chen, G.; Dalton, C.; Irvine, K.; Taylor, D.	Patterns in freshwater diatom taxonomic distinctness along an eutrophication gradient	FRESHWATER BIOLOGY	2009
Tullos, DD; Penrose, DL; Jennings, GD; Cope, WG	Analysis of functional traits in reconfigured channels: implications for the bioassessment and disturbance of river restoration	JOURNAL OF THE NORTH AMERICAN BENTHOLOGICAL SOCIETY	2009
Michelan, TS; Thomaz, SM; Mormul, RP; Carvalho, P	Effects of an exotic invasive macrophyte (tropical signalgrass) on native plant community composition, species richness and functional diversity	FRESHWATER BIOLOGY	2010
Peru, N; Doledec, S	From compositional to functional biodiversity metrics in bioassessment: A case study using stream macroinvertebrate communities	ECOLOGICAL INDICATORS	2010
Pool, TK; Olden, JD; Whittier, JB; Paukert, CP	Environmental drivers of fish functional diversity and composition in the Lower Colorado River Basin	CANADIAN JOURNAL OF FISHERIES AND AQUATIC SCIENCES	2010
Strecker, AL; Olden, JD; Whittier, JB; Paukert, CP	Defining conservation priorities for freshwater fishes according to taxonomic, functional, and phylogenetic diversity	ECOLOGICAL APPLICATIONS	2011
Arthaud, F; Vallod, D; Robin, J; Bornette, G	Eutrophication and drought disturbance shape functional diversity and life-history traits of aquatic plants in shallow lakes	AQUATIC SCIENCES	2012
Milosevic, Djuradj; Simic, Vladica; Stojkovic, Milica; Zivic, Ivana	Chironomid faunal composition represented by taxonomic distinctness index reveals environmental change in a lotic system over three decades	HYDROBIOLOGIA	2012
Pool, TK; Olden, JD	Taxonomic and functional homogenization of an endemic desert fish fauna	DIVERSITY AND DISTRIBUTIONS	2012
Schmera, D; Baur, B; Eros, T	Does functional redundancy of communities provide insurance against human disturbances? An analysis using regional-scale stream invertebrate data	HYDROBIOLOGIA	2012

Silver, CA; Vamosi, SM; Bayley, SE	Temporary and permanent wetland macroinvertebrate communities: Phylogenetic structure through time	ACTA OECOLOGIA	2012
Teresa, FB; Casatti, L	Influence of forest cover and mesohabitat types on functional and taxonomic diversity of fish communities in Neotropical lowland streams	ECOLOGY OF FRESHWATER FISH	2012
Angeler, DG; Allen, CR; Johnson, RK	Measuring the relative resilience of subarctic lakes to global change: redundancies of functions within and across temporal scales	JOURNAL OF APPLIED ECOLOGY	2013
Colzani, E; Siqueira, T; Suriano, MT; Roque, FO	Responses of Aquatic Insect Functional Diversity to Landscape Changes in Atlantic Forest	BIOTROPICA	2013
Liu, WZ; Liu, GH; Liu, H; Song, Y; Zhang, QF	Subtropical reservoir shorelines have reduced plant species and functional richness compared with adjacent riparian wetlands	ENVIRONMENTAL RESEARCH LETTERS	2013
Martinez, A; Larranaga, A; Basaguren, A; Perez, J; Mendoza-Lera, C; Pozo, J	Stream regulation by small dams affects benthic macroinvertebrate communities: from structural changes to functional implications	HYDROBIOLOGIA	2013
Matsuzaki, SS; Sasaki, T; Akasaka, M	Consequences of the introduction of exotic and translocated species and future extirpations on the functional diversity of freshwater fish assemblages	GLOBAL ECOLOGY AND BIOGEOGRAPHY	2013
Paillex, A; Doledec, S; Castella, E; Merigoux, S; Aldridge, DC	Functional diversity in a large river floodplain: anticipating the response of native and alien macroinvertebrates to the restoration of hydrological connectivity	JOURNAL OF APPLIED ECOLOGY	2013
Arce, E; Archaimbault, V; Mondy, CP; Usseglio-Polatera, P	Recovery dynamics in invertebrate communities following water-quality improvement: taxonomy- vs trait-based assessment	FRESHWATER SCIENCE	2014
Barberan, A; Casamayor, EO	A phylogenetic perspective on species diversity, beta-diversity and biogeography for the microbial world	MOLECULAR ECOLOGY	2014
Boersma, KS; Bogan, MT; Henrichs, BA; Lytle, DA	Invertebrate assemblages of pools in arid-land streams have high functional redundancy and are resistant to severe drying	FRESHWATER BIOLOGY	2014
Feld, CK; de Bello, F; Doledec, S	Biodiversity of traits and species both show weak responses to hydromorphological alteration in lowland river macroinvertebrates	FRESHWATER BIOLOGY	2014

Fernandez, C; Caceres, EJ; Parodi, ER	Phytoplankton Development in a Highly Eutrophic man-made Lake From the Pampa plain of Argentina-a functional Approach	INTERNATIONAL JOURNAL OF ENVIRONMENTAL RESEARCH	2014
Hitt, NP; Chambers, DB	Temporal changes in taxonomic and functional diversity of fish assemblages downstream from mountaintop mining	FRESHWATER SCIENCE	2014
Huang, QY; Briggs, BR; Dong, HL; Jiang, HC; Wu, G; Edwardson, C; De Vlaminck, I; Quake, S	Taxonomic and Functional Diversity Provides Insight into Microbial Pathways and Stress Responses in the Saline Qinghai Lake, China	PLOS ONE	2014
Jiang, Xiaoming; Song, Zhuoyan; Xiong, Jing; Xie, Zhicai	Can excluding non-insect taxa from stream macroinvertebrate surveys enhance the sensitivity of taxonomic distinctness indices to human disturbance?	ECOLOGICAL INDICATORS	2014
Kovalenko, KE; Brady, VJ; Ciborowski, JJH; Ilyushkin, S; Johnson, LB	Functional Changes in Littoral Macroinvertebrate Communities in Response to Watershed-Level Anthropogenic Stress	PLOS ONE	2014
Navarro, MB; Balseiro, E; Modenutti, B	Bacterial Community Structure in Patagonian Andean Lakes Above and Below Timberline: From Community Composition to Community Function	MICROBIAL ECOLOGY	2014
Romani, AM; Borrego, CM; Diaz-Villanueva, V; Freixa, A; Gich, F; Ylla, I	Shifts in microbial community structure and function in light- and dark-grown biofilms driven by warming	ENVIRONMENTAL MICROBIOLOGY	2014
Sternberg, D; Kennard, MJ; Balcombe, SR	Biogeographic determinants of Australian freshwater fish life-history indices assessed within a spatio-phylogenetic framework	GLOBAL ECOLOGY AND BIOGEOGRAPHY	2014
Timoner, X; Acuna, V; Frampton, L; Pollard, P; Sabater, S; Bunn, SE	Biofilm functional responses to the rehydration of a dry intermittent stream	HYDROBIOLOGIA	2014
Angeler, DG; Allen, CR; Uden, DR; Johnson, RK	Spatial Patterns and Functional Redundancies in a Changing Boreal Lake Landscape	ECOSYSTEMS	2015
Arrieira, RL; Schwind, LTF; Bonecker, CC; Lansac-Toha, FA	Use of functional diversity to assess determinant assembly processes of testate amoebae community	AQUATIC ECOLOGY	2015

Carvalho, RA; Tejerina-Garro, FL	Environmental and spatial processes: what controls the functional structure of fish assemblages in tropical rivers and headwater streams?	ECOLOGY OF FRESHWATER FISH	2015
Casatti, L; Teresa, FB; Zeni, JD; Ribeiro, MD; Brejao, GL; Ceneviva-Bastos, M	More of the Same: High Functional Redundancy in Stream Fish Assemblages from Tropical Agroecosystems	ENVIRONMENTAL MANAGEMENT	2015
Chmara, R; Banas, K; Szmeja, J	Changes in the structural and functional diversity of macrophyte communities along an acidity gradient in softwater lakes	FLORA	2015
Cibils, L; Principe, R; Marquez, J; Gari, N; Albarino, R	Functional diversity of algal communities from headwater grassland streams: How does it change following afforestation?	AQUATIC ECOLOGY	2015
Espanol, C; Gallardo, B; Comin, FA; Pino, MR	Constructed wetlands increase the taxonomic and functional diversity of a degraded floodplain	AQUATIC SCIENCES	2015
Fu, H; Zhong, JY; Yuan, GX; Guo, CJ; Ding, HJ; Feng, Q; Fu, Q	A functional-trait approach reveals community diversity and assembly processes responses to flood disturbance in a subtropical wetland	ECOLOGICAL RESEARCH	2015
He, FZ; Jiang, WX; Tang, T; Cai, QH	Assessing impact of acid mine drainage on benthic macroinvertebrates: can functional diversity metrics be used as indicators?	JOURNAL OF FRESHWATER ECOLOGY	2015
Pease, AA; Taylor, JM; Winemiller, KO; King, RS	Ecoregional, catchment, and reach-scale environmental factors shape functional-trait structure of stream fish assemblages	HYDROBIOLOGIA	2015
Queiroz, CD; da Silva, FR; Rossa-Feres, DD	The relationship between pond habitat depth and functional tadpole diversity in an agricultural landscape	ROYAL SOCIETY OPEN SCIENCE	2015
Saito, VS; Siqueira, T; Fonseca-Gessner, AA	Should phylogenetic and functional diversity metrics compose macroinvertebrate multimetric indices for stream biomonitoring?	HYDROBIOLOGIA	2015
Santos, AMC; Carneiro, FM; Cianciaruso, MV	Predicting productivity in tropical reservoirs: The roles of phytoplankton taxonomic and functional diversity	ECOLOGICAL INDICATORS	2015
Boersma, KS; Nickerson, A; Francis, CD; Siepielski, AM	Climate extremes are associated with invertebrate taxonomic and functional composition in mountain lakes	ECOLOGY AND EVOLUTION	2016
Carvajal-Castro, JD; Vargas-Salinas, F	Stream noise, habitat filtering, and the phenotypic and phylogenetic structure of Neotropical anuran assemblages	EVOLUTIONARY ECOLOGY	2016

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Godet, L; Devictor, V; Burel, F; Robin, JG; Menanteau, L; Fournier, J	Extreme landscapes decrease taxonomic and functional bird diversity but promote the presence of rare species	ACTA ORNITHOLOGICA	2016
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Ding, N; Yang, WF; Zhou, YL; Gonzalez-Bergonzoni, I; Zhang, J; Chen, K; Vidal, N; Jeppesen, E; Liu, ZW; Wang, BX	Different responses of functional traits and diversity of stream macroinvertebrates to environmental and spatial factors in the Xishuangbanna watershed of the upper Mekong River Basin, China	SCIENCE OF THE TOTAL ENVIRONMENT	2017
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Saulino, HHL; Trivinho- Strixino, S	The invasive white ginger lily (<i>Hedichium coronarium</i>) simplifies the trait composition of an insect assemblage in the littoral zone of a Savanna reservoir	REVISTA BRASILEIRA DE ENTOMOLOGIA	2017
Suarez, ML; Sanchez- Montoya, MM; Gomez, R; Arce, MI; del Campo, R; Vidal- Abarca, MR	Functional response of aquatic invertebrate communities along two natural stress gradients (water salinity and flow intermittence) in Mediterranean streams	AQUATIC SCIENCES	2017
Gimenez, Barbara C. G.; Higuti, Janet	Land use effects on the functional structure of aquatic insect communities in Neotropical streams	INLAND WATERS	2017
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Weithoff, Guntram; Gaedke, Ursula	Mean functional traits of lake phytoplankton reflect seasonal and inter-annual changes in nutrients, climate and herbivory	JOURNAL OF PLANKTON RESEARCH	2017
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Sousa Rodrigues-Filho, Carlos Alberto; Gurgel-LourenOo, Ronaldo Cesar; Queiroz Lima, Sergio Maia; de Oliveira, Edson Fontes; Sanchez-Botero, Jorge Ivan	What governs the functional diversity patterns of fishes in the headwater streams of the humid forest enclaves: environmental conditions, taxonomic diversity or biotic interactions?	ENVIRONMENTAL BIOLOGY OF FISHES	2017
Sagouis, Alban; Jabot, Franck; Argillier, Christine	Taxonomic versus functional diversity metrics: how do fish communities respond to anthropogenic stressors in reservoirs?	ECOLOGY OF FRESHWATER FISH	2017
Machado, Karine Borges; Teresa, Fabricio Barreto; Nabout, Joao Carlos	Assessing the spatial variation of functional diversity estimates based on dendrograms in phytoplankton communities	ACTA BOTANICA BRASILICA	2017
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