

# Unexpected genetic structure of mussel populations in South Africa: indigenous *Perna perna* and invasive *Mytilus galloprovincialis*

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**ABSTRACT:** Genetic structure of sedentary marine organisms with planktonic larvae can be influenced by oceanographic transport, larval behaviour and local selection. We analysed the population genetic structure (based on mtDNA) of the invasive mussel *Mytilus galloprovincialis* and the indigenous mussel *Perna perna* along the southern African coastline. Low genetic divergence of *M. galloprovincialis* confirms its recent arrival in South Africa. In contrast, the genetic structure of *P. perna* revealed strong divergence on the south-east coast, forming a western and an eastern lineage. The distribution of the 2 lineages is extraordinary. They overlap for ca. 200 km on the south-east coast, and the western lineage includes animals occurring on either side of a 1000 km break in distribution across the Benguela upwelling system. In cluster analyses, animals on the south coast grouped with others 1000s of km to the west, rather than with those only 200 km to the east. This genetic disjunction may be caused by the south-flowing Agulhas Current preventing larval dispersal, or by different selective forces acting on local populations. *M. galloprovincialis* spread eastward along the south coast for 15 yr, but its range extension has virtually ceased in the region of genetic disjunction in *P. perna*, again indicating an oceanographic barrier to larval dispersal or selection driven by sharp gradients in environmental conditions. The results suggest that local selection can produce genetic structure opposite to that predicted by oceanographic data and that determining the population structure of indigenous species with similar larval dispersal can help us understand domain expansion of invading species.

**KEY WORDS:** *Mytilus galloprovincialis* · *Perna perna* · Invasion · Biogeographic region · Currents · mtDNA

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## INTRODUCTION

An important factor influencing the dynamics of a biological invasion is the ability of the invader to disperse, thus spreading from the founder population and extending its domain. The dispersal potential of seeds, spores influences large-scale patterns of distribution and geographic ranges of sedentary organisms (Perron & Kohn 1985, Richmond 1987, Scheltema 1989, Gaines & Bertness 1992, Emler 1995). Marine species possessing pelagic larvae have the potential for a high degree of connectivity among local popula-

tions. This connectivity is potentially of considerable importance and reflects the magnitudes of immigration and emigration (Sale 1991, Caley et al. 1996). Gene flow (via larval dispersal) can provide information on connectivity among populations (Avice et al. 1987, Burton 1998). In this study, we explored the population genetic structure of the 2 dominant intertidal mussel species in South Africa, the invasive *Mytilus galloprovincialis* and the indigenous *Perna perna*, over the entire South African and southern Namibian coastlines, and we found unexpected population structure.

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*Mytilus galloprovincialis* is invasive in many parts of the world and is the most successful marine invader in South Africa (Grant & Cherry 1985, Branch & Steffani 2004, Robinson et al. 2005). This European species arrived in South Africa at Saldanha Bay (Fig. 1; Site 14, 150 km north of the Cape of Good Hope) in the early 1970s, probably with shipping (Grant & Cherry 1985). Soon after its first detection, it spread to the north at an average rate of  $115 \text{ km yr}^{-1}$  and to the south at about  $25 \text{ km yr}^{-1}$  (Hockey & van Erkom Schurink 1992). It is now the dominant intertidal mussel from the Cape of Good Hope in South Africa to Lüderitz in southern Namibia (Fig. 1; Griffiths et al. 1992, Hockey & van Erkom Schurink 1992, Branch & Steffani 2004). The European mussel stopped its progression in central Namibia where, presumably, the subtropical conditions characterising northern Namibia (Shannon 1985) represent the limits of its tolerance. *M. galloprovincialis* has spread from the Cape of Good Hope eastwards to within 20 km west of East London (Fig. 1; McQuaid & Phillips 2000, Robinson et al. 2005). In the last 10 yr, its expansion on the south coast has virtually ceased, and it may have reached its biogeographic limit (Robinson et al. 2005). As on the Namibian coast,

this limit could be explained by the subtropical conditions found north of East London, indicating that *M. galloprovincialis* invasion in South Africa conforms to the antitropical distribution pattern typical of *Mytilus* spp. (Hilbish et al. 2000).

*Perna perna* is abundant along the entire east coast of South Africa, and on the south coast of South Africa it shows partial habitat segregation with *Mytilus galloprovincialis* (Bownes & McQuaid 2006), which it competitively excludes from the low shore (Rius & McQuaid 2006). The cold waters of the Benguela upwelling system limit *P. perna* distribution north of the Cape of Good Hope, and it is absent from there to Lüderitz in southern Namibia (Fig. 1), a distance of approximately 1000 km. *P. perna* re-appears in Namibia and extends along the west coast of Africa to the Mediterranean Sea as far as the Gulf of Tunis (van Erkom Schurink & Griffiths 1990).

Both species have planktotrophic larvae that disperse in the water column for a period of feeding and growth, before settling to the substratum and being recruited into an adult population. Given the relatively long period of larval duration, both species have the potential for high dispersal rates. However, several studies have shown discrepancies between the potential and the realised dispersal of larvae (Palumbi 1994, Todd 1998), and on the South African coast, dispersal scales of *Mytilus galloprovincialis* larvae are limited to 10s of km (McQuaid & Phillips 2000). Such discrepancies are mainly due to the influence of oceanographic conditions on larval behaviour, to reproductive timing and to the availability of suitable habitat in terms of both biotic and abiotic factors (Shanks 1995, Todd 1998, Shanks & Brink 2005). Consequently, we would expect the genetic structure of populations of species with planktonic larvae to be strongly shaped by prevailing oceanographic conditions.

The primary oceanographic influence on the east and south coasts of South Africa is the Agulhas Current. This warm current is about 60 to 100 km wide and flows to the southwest along the eastern seaboard of South Africa (from  $27^{\circ}\text{S}$  to  $40^{\circ}\text{S}$ ) at rates of 10 to  $20 \text{ km d}^{-1}$ , following the 200 m isobath of the continental shelf from Maputo in Mozambique to the tip of the Agulhas Bank in South Africa (Fig. 1; Lutjeharms 1998, 2004). The inshore thermal front of this current varies geographically and in time, and may alter patterns of along-shelf larval dispersal. It usually lies 14 to 38 km offshore, but it can flow onto the coast at 0 to 1 km offshore (Goschen & Schumann 1990). In contrast, the coastal environment of western South Africa, Namibia and southern Angola is profoundly influenced by the Benguela Current, which flows from the Cape of Good Hope northwards to Lüderitz, where the main flow is deflected away from the coast to the northwest (Peter-

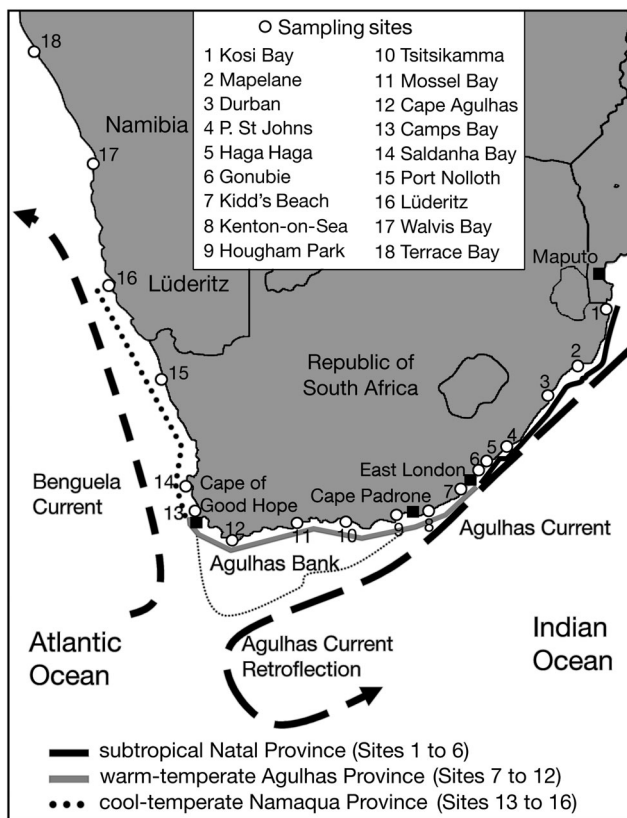


Fig. 1. Sampling locations for *Mytilus galloprovincialis* and *Perna perna*, and location of Agulhas and Benguela Currents. Biogeographical provinces according to Emanuel et al. (1992)

son & Stramma 1991, Wedepohl et al. 2000; Fig. 1). The Benguela Current is characterised by Ekman-driven coastal upwelling. Intense and consistent upwelling off Lüderitz (27–28°S) separates the Northern Benguela from the Southern Benguela and creates a semi-permanent environmental barrier (Boyd & Cruickshank 1983, Agenbag & Shannon 1988). *Perna perna* is missing from the entire Benguela region along 1000 km of coast.

Genetic differentiation can occur among populations of marine organisms with high dispersal potential due to local differences in selection (Johnson & Black 1984, Hedgecock 1986, Bertness & Gaines 1993). The South African coastline covers a wide range of climatic and oceanic conditions and can be divided into biogeographic regions that support a great diversity of algae and animals. Based on an analysis of rocky shore invertebrates, Emanuel et al. (1992) divided the South African coast into 3 zoogeographic regions (Fig. 1): from Lüderitz in Namibia to the Cape of Good Hope (cool-temperate Namaqua Province); from the Cape of Good Hope to East London (warm-temperate Agulhas Province); from East London to Mozambique (sub-tropical Natal Province).

In this study, we used mitochondrial DNA (mtDNA) to investigate the genetic structure of *Mytilus galloprovincialis* and *Perna perna* populations along the South African and Namibian coastline. Specifically, we wished to compare an invasive and an indigenous species and to examine the effects on their genetics of both biogeography and the oceanographic conditions dominating the nearshore environment.

## MATERIALS AND METHODS

**Sampling, DNA extraction, amplification and sequencing.** Populations of *Mytilus galloprovincialis* and *Perna perna* were sampled at 11 and 14 sites, respectively, along the Namibian and South African coasts (10 individuals from each population; Fig. 1). Mussels were opened in the laboratory, and a piece of gonad tissue was examined under a microscope to determine the sex of the animal by the presence of eggs or sperm. Because of doubly uniparental inheritance, only female individuals were used in this study. Whole genomic DNA was extracted from approximately 1 mm<sup>3</sup> of gonad tissue (attached to the mantle) using a standard phenol-chloroform extraction method, and samples were then re-dissolved in 50 µl water. The primers LCOI 1490 (5'-GGT CAA CAA ATC ATA AAG ATA TTG-3') and HCO 2198 (5'-TAA ACT TCA GGG TGA CCA AAA AAT CA-3') (Folmer et al. 1994) were used in a polymerase chain reaction (PCR) to amplify a portion of the mitochondrial

cytochrome oxidase subunit I gene (mtDNA COI). Amplifications were performed in a 100 µl solution containing 10 to 100 µg of DNA, 0.4 µM of each primer, 5 µl of Qiagen PCR buffer, 200 µM of each dNTP, and 2.5 U of Taq DNA polymerase (Qiagen). The PCR cycling profile comprised an initial denaturation step at 94°C for 2 min, 35 cycles of denaturation at 94°C for 60 s, annealing at 54°C for 60 s, extension at 72°C for 90 s, and a final extension at 72°C for 5 min. PCR products from each individual were purified with a Qiaquick gel extraction kit (Qiagen) and cycle-sequenced in both forward and reverse direction with the same primers used in the amplification, using a BigDye<sup>®</sup> Terminator v3.1 Cycle Sequencing kit (Applied Biosystems) and sequenced on an ABI 3100 genetic analyser.

**Data analysis.** DNA sequences were translated into amino acid data using the standard invertebrate mitochondrial genetic code to verify their mitochondrial origin. We are confident of the origin because: (1) there was at least one reading frame with no stop codon, (2) all translated sequences matched with published records of COI sequences in GenBank, and (3) no variable nucleotide positions consistently showing double peaks were encountered in chromatograms, as might be expected in the case of nuclear pseudogenes.

PAUP\* version 4.0b10 (Swofford 2002) was used to construct neighbour-joining trees using the most suitable model of sequence evolution for each species, as suggested by the Akaike Information Criterion (AIC), as implemented in the program MODELTEST 3.06 (Posada & Crandall 1998). In addition, parsimony analyses were done using PAUP\* version 4.0b10 (Swofford 2002). Heuristic tree searches were performed using 1000 random sequence additions and tree bisection-reconnection (TBR) branch swapping. Non-parametric bootstrap values (Felsenstein 1985) were calculated using 1000 replicates and 10 random taxon additions.

A triangular matrix of p-distances among *Perna perna* individuals was generated using PAUP\* version 4.0b10 (Swofford 2002) and imported into Excel to calculate mean intra- and inter-lineage average sequence divergence.

**Isolation by distance.** Mitochondrial DNA lineages identified using genealogical reconstructions were examined for evidence of isolation by distance by testing for correlation between genetic and geographic distance (Slatkin 1993). ARLEQUIN version 2.001 (Schneider et al. 2000) was used to calculate  $\phi_{ST}$  values (as a measure of sequence divergence among haplotypes; Excoffier et al. 1992) among populations associated with specific lineages. The geographic distance between populations was measured as the shortest continuous water-surface distance. The relationship

between genetic differentiation and geographical distance was assessed by performing Mantel tests using the program MANTEL for Windows version 1.16 (Cavalcanti 2005). The significance of the Mantel statistic  $Z$  is tested by a permutation procedure in which values in 1 data matrix are randomly reshuffled (20 000 permuted data-sets were created).

## RESULTS

### Sequence characteristics

Sequence data (400 bp) from 110 and 140 individuals of *Mytilus galloprovincialis* and *Perna perna*, respectively, were analysed. In the case of *M. galloprovincialis*, 29 nucleotide sites were polymorphic and 18 were parsimony informative; a total of 21 haplotypes was identified. *P. perna* had 50 unique haplotypes and 58 polymorphic sites, 27 of which were parsimony informative. Haplotypes of both species have been deposited in GenBank (*P. perna*: DQ351427–DQ351476, *M. galloprovincialis*: DQ351477–DQ351497).

### Phylogeographical patterns

The neighbour-joining tree of *Perna perna* recovered 2 distinct lineages (Fig. 2). One lineage included samples from Terrace Bay to Haga Haga (i.e. from the Namibian coast to the southeast South African coast). Samples of the other lineage originated from Kosi Bay to Kenton-on-Sea (i.e. east coast to southeast coast). The distributions of the 2 *P. perna* lineages thus overlap for about 200 km between Haga Haga and Kenton-on-Sea on the southeast coast (Fig. 2). Average within-lineage sequence divergences were 1.1 and 1.0% for the eastern and western lineage, respectively. Average sequence divergence between the 2 lineages was 2.9%.

*Perna perna* showed greater nucleotide diversity than *Mytilus galloprovincialis*. Haplotypes 1 and 2 were the most common in populations from the eastern lineage

(east coast) and the western lineage (south and west coasts), respectively (Fig. 3). Haplotype 1 was never found in the Namibian populations, and west of Gonubie it was found in only 2 individuals (1 each from Kenton-on-Sea and Kidd's Beach; Fig. 3). Haplotype 2 was the most common haplotype on the south coast and was also sampled once in Walvis Bay. It was never found farther east than Kidd's Beach (Fig. 3). Haplotype 3 was most common on the south coast, but was also sampled at Terrace Bay in Namibia and Haga Haga. Haplotypes restricted to a single site (private haplotypes) were found at all locations, and were more frequent in the western lineage than in the eastern lineage. Note that a greater sample size could have changed the proportion of

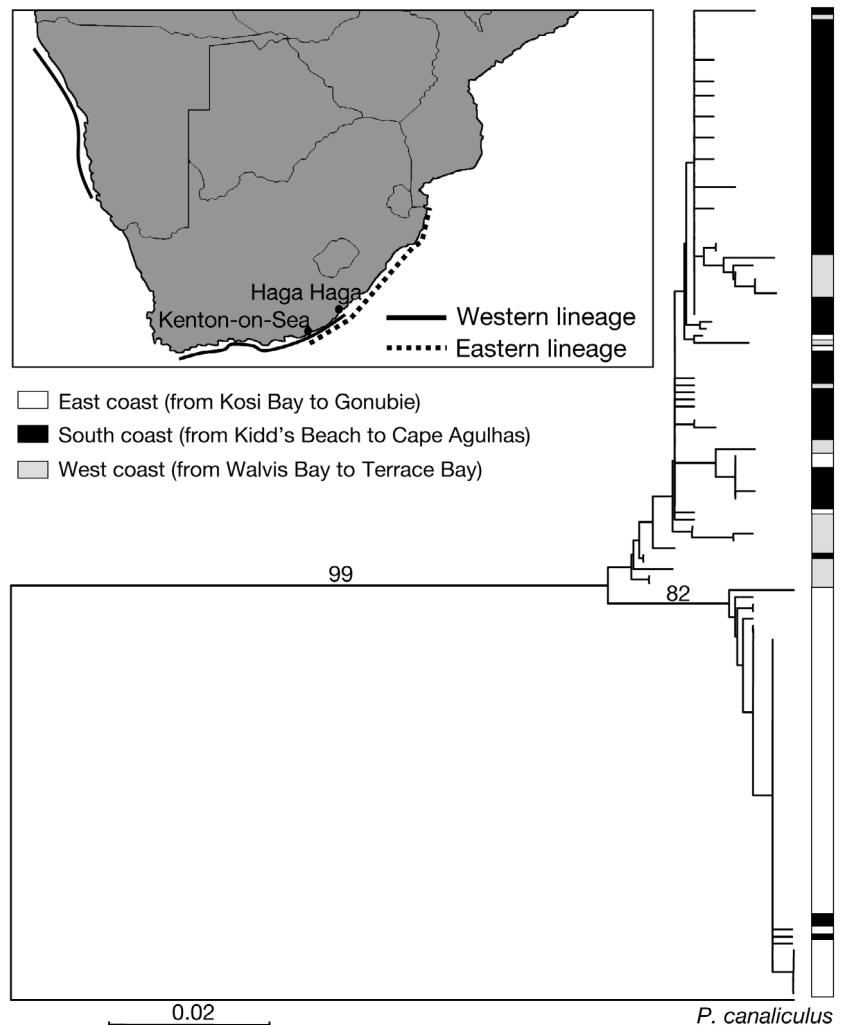


Fig. 2. *Perna perna*. Geographic distribution pattern of the mtDNA lineages and neighbour-joining phylogram based on sequences from cytochrome oxidase I (COI). The tree was rooted using a female *P. canaliculus* as an outgroup. Parsimony bootstrap values from 1000 bootstrap replications were added to some branches. Model selected: TrN (Tamura & Nei 1993). Gamma distribution parameter ( $\gamma$ ) estimated in MODELTEST was 0.22. See Fig. 1 for location codes

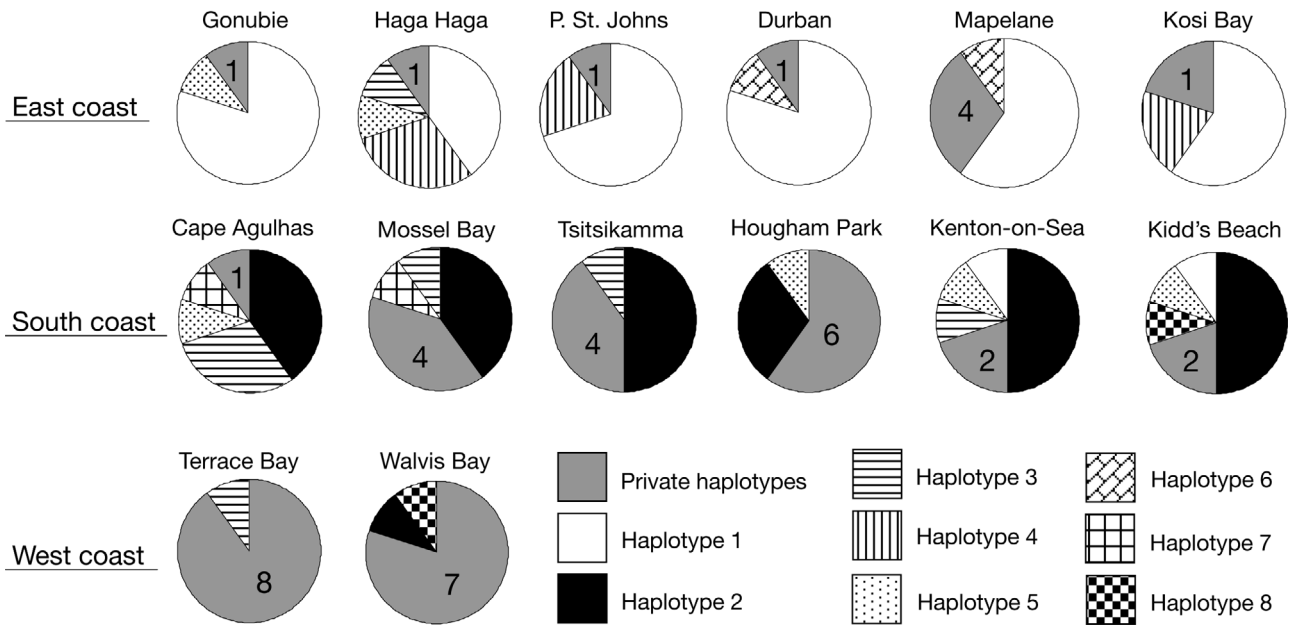


Fig. 3. *Perna perna*. Haplotype frequencies at each sampling site. For private haplotypes, the number of haplotypes is given

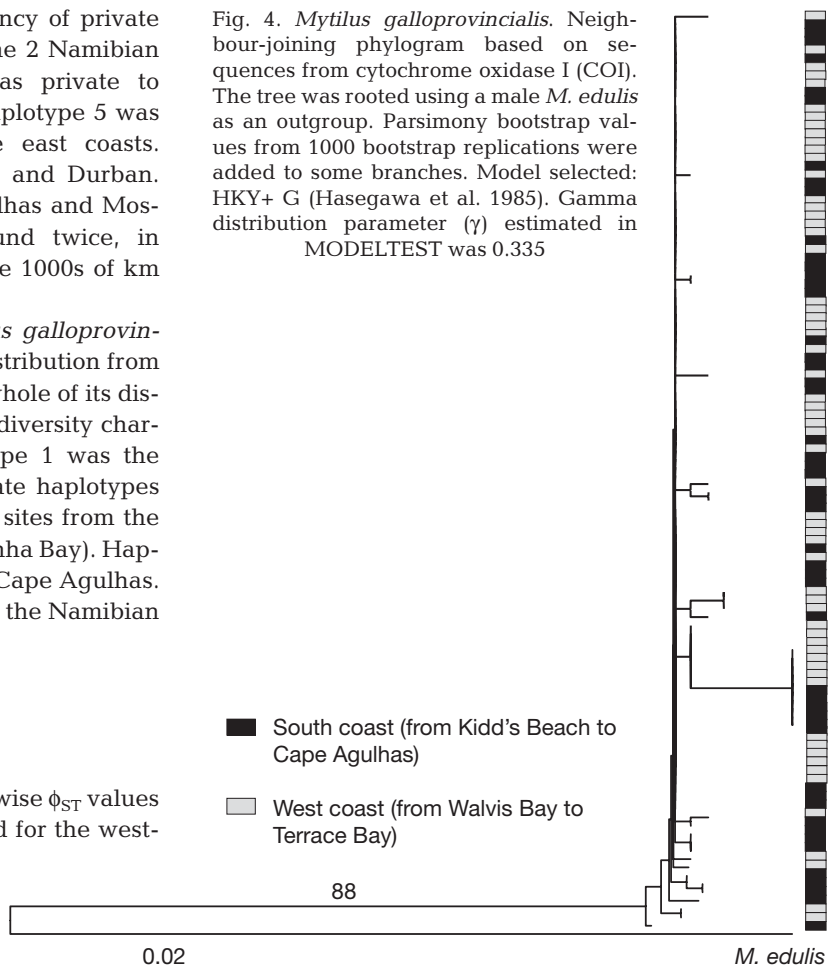
private haplotypes. The highest frequency of private haplotypes (80 to 90%) was found in the 2 Namibian populations (Fig. 3). Haplotype 4 was private to the east coast. As with Haplotype 1, Haplotype 5 was sampled on both the south and the east coasts. Haplotype 6 was private to Mapelane and Durban. Haplotype 10 was private to Cape Agulhas and Mossel Bay. Haplotype 11 was only found twice, in Walvis Bay and Kidd's Beach, which are 1000s of km apart (Fig. 1).

The neighbour-joining tree of *Mytilus galloprovincialis* recovered only 1 lineage with a distribution from Terrace Bay to Kenton-on-Sea, i.e. the whole of its distribution range (Fig. 4). Low haplotype diversity characterised *M. galloprovincialis*. Haplotype 1 was the most common at all sites (Fig. 5). Private haplotypes were found at all locations, except at 3 sites from the west coast (Terrace Bay, Lüderitz, Saldanha Bay). Haplotype 3 was private to Mossel Bay and Cape Agulhas. Haplotypes 4, 5 and 6 were all private to the Namibian and South African west coasts.

#### Isolation by distance

A significant correlation between pairwise  $\phi_{ST}$  values and geographic distance was only found for the western lineage of *Perna perna* (Table 1). However, when the 2 Namibian populations were excluded, the relationship was no longer significant.

Fig. 4. *Mytilus galloprovincialis*. Neighbour-joining phylogram based on sequences from cytochrome oxidase I (COI). The tree was rooted using a male *M. edulis* as an outgroup. Parsimony bootstrap values from 1000 bootstrap replications were added to some branches. Model selected: HKY+ G (Hasegawa et al. 1985). Gamma distribution parameter ( $\gamma$ ) estimated in MODELTEST was 0.335





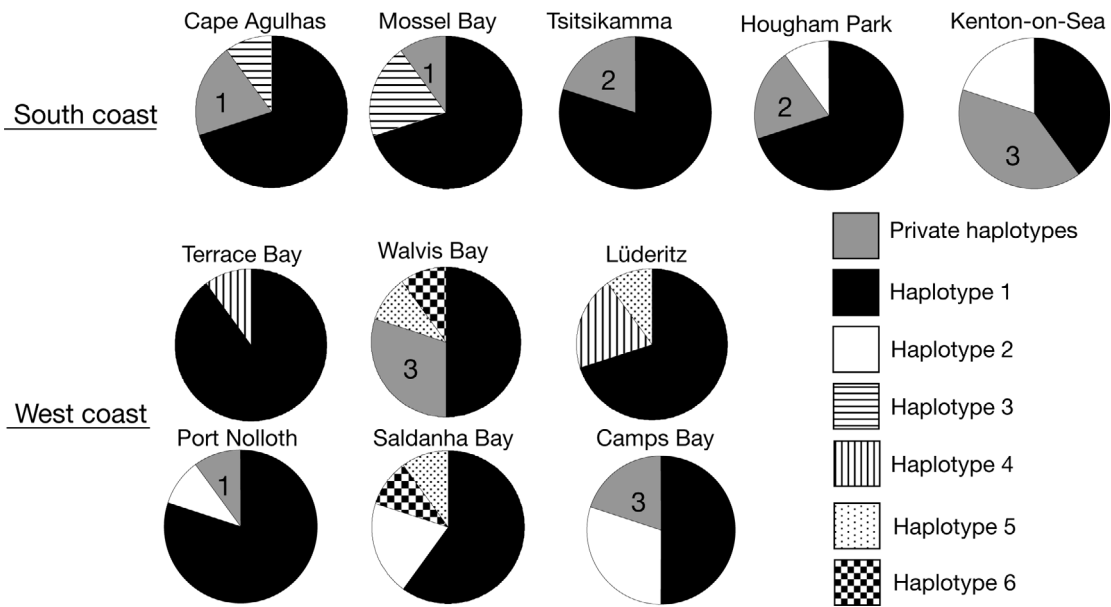


Fig. 5. *Mytilus galloprovincialis*. Haplotype frequencies at each sampling site. For private haplotypes, the number of haplotypes is given

**DISCUSSION**

Mitochondrial DNA sequences of *Perna perna* indicate a strong phylogeographic break on the east coast between Kidd’s Beach and Gonubie, a distance of about 40 km. A previous study showed that there was essentially no allozyme-frequency differentiation in *P. perna* along the whole South African coastline (Grant et al. 1992). The only difference in allozyme frequencies was found at the western edge of the geographic distribution of *P. perna* on the southwest coast. We could not find *P. perna* on the western shore of the

Cape Peninsula, which seems to confirm the suggestion of Grant et al. (1992) that populations are not self-sustaining, but originate by chance recruitment from more eastern subpopulations and are subject to dramatic changes in population structure. Grant et al. (1992) collected an additional sample at Swakopmund (Namibia), showing the absence of alleles unique to the Namibian samples, and indicating close genetic affinity to South African populations.

Our study shows that some haplotypes occurring in the South African samples are also present in the Namibian samples, suggesting a certain level of gene flow between the 2 regions. It is unlikely that *Perna perna* larvae are now able to connect Namibian and South African populations by dispersing through the cold waters of the Benguela upwelling system. However, the Pleistocene or Quaternary oceanic warming (Shannon 1985, Thackeray & Herbert 1991) could have allowed colonisation of this stretch of coast and dispersal through the present distribution gap. Mantel test results reported here indicate that the gap of about 1000 km in distribution of *P. perna* between Cape Agulhas and Walvis Bay on the west coast results in significant isolation by distance. Nevertheless, *P. perna* from Kenton-on-Sea group with those from Namibia, 2000 km away, rather than with those from Haga Haga, 200 km to the east. Several competing hypotheses could explain the separation of the 2 lineages in *Perna perna*:

(1) Physical isolation, resulting in reduced gene flow. The Agulhas Current lies relatively close inshore in the overlap region and could act as a barrier to the disper-

Table 1. Results of Mantel tests on matrices of genetic differentiation among populations (pairwise  $\Phi_{ST}$  using p-distances) and geographic distance of the *Mytilus galloprovincialis* mtDNA lineage, and of the *Perna perna* Western lineage, Eastern lineage and Western lineage excluding the 2 Namibian populations.  $r$  = Pearson product-moment correlation (shown instead of the Mantel  $Z$  statistic for simplicity of interpretation),  $t$  = approximate Mantel  $t$ -test statistic,  $p$  = probability that random  $Z < \text{observed } Z$ . Significant value in **bold**

	$r$	$t$	$P$
<i>Perna perna</i>			
Western	-0.358	-2.252	<b>0.0122</b>
Western excluding Terrace Bay and Walvis Bay	0.202	-0.966	0.833
Eastern	0.192	0.850	0.8
Total (Western + Eastern)	-0.141	-1.354	0.09
<i>Mytilus galloprovincialis</i>			
Total	-0.052	-0.374	0.35

sal of larvae from west to east, but this does not explain the lack of dispersal in the opposite direction. Nor does it accord with the fact that *P. perna* on either side of the Benguela System, which forms a formidable barrier to dispersal, share a dominant haplotype absent from the eastern lineage.

(2) Recent secondary contact between 2 lineages that have been separated for a long period, though again it is difficult to identify possible mechanisms of isolation.

(3) Differential selection. Many studies have investigated the limits of the marine biogeographic regions of South Africa using different organisms (Stephenson & Stephenson 1972, Day 1981, Potter et al. 1990, Prochazka 1994, Turpie et al. 2000). The general agreement is that the boundary between the warm temperate region and the subtropical region lies on the southeast coast (Harrison 2000). Environmental factors characterising these areas may subject *P. perna* populations to different selective forces leading to genetic divergence. To produce such a sharp discontinuity, such selective forces would have to be very powerful.

Grant & Cherry (1985) examined shells taken from prehistoric shell middens and the shell collection at the Zoology Department of the University of Cape Town and concluded that the introduction of *Mytilus galloprovincialis* to southern Africa took place within the previous 20 yr. The low haplotype diversity over the total geographic range of *M. galloprovincialis* in South Africa confirms its recent arrival. After spreading along the south coast, it seems that *M. galloprovincialis* has reached its biogeographic limit near Kidd's Beach in the East London area (McQuaid & Philips 2000, Robinson et al. 2005), that is, in the area of phylogenetic discontinuity in *Perna perna*. *M. galloprovincialis* is highly invasive in many parts of the world and is well adapted to a wide range of environmental factors, including temperature (Branch & Steffani 2004). Invasions frequently constitute rapid evolutionary events (Reznick & Ghilambor 2001), resulting in populations that are genetically dynamic over both space and time. Invasive species are often assumed to penetrate habitat boundaries through their broad tolerance (eurytolerance) or phenotypic plasticity (Ricciardi & MacIsaac 2000, Wolff 2000, Reid & Orlova 2002). However, certain populations cannot tolerate or acclimatise to the full range of environmental factors occupied by the species complex, but instead experience strong selection when invading new habitats (Lee 1999, Lee & Petersen 2003). Consequently, the environmental conditions prevailing in the different biogeographic regions of South Africa could exert strong selection on the physiological tolerance and performance of *M. galloprovincialis* during habitat invasion. Another possible explanation for the apparent drop in rate of

spread near East London is that *M. galloprovincialis*, under east coast environmental conditions, is a weaker competitor in the interaction with the eastern lineage of *P. perna*, which excludes *M. galloprovincialis* from the low shore in the Kenton-on-Sea area (Rius & McQuaid 2006), and it is possible that farther east, weaker performance of *M. galloprovincialis*, coupled with stronger performance of a different genetic lineage of *P. perna*, excludes the invasive species not only from the lower shore but from the entire intertidal habitat.

The west coast of South Africa is influenced by the upwelling of cool, nutrient-rich water, and *Perna perna* is presumably excluded from this region by low water temperatures. Strong perennial upwelling off Lüderitz effectively separates the Northern Benguela from the Southern Benguela. A northwesterly moving tongue of upwelled, turbulent water acts as a semi-permanent environmental barrier to the longshore transport of pelagic fish eggs and larvae (O'Toole 1977, Agenbag 1980, Boyd & Cruickshank 1983, Agenbag & Shannon 1988). However, this oceanographic barrier did not prevent the expansion of *Mytilus galloprovincialis* from Saldana Bay, and it was able to rapidly colonise this stretch of coast as far as central Namibia (Branch & Steffani 2004).

High levels of genetic divergence, either within or among populations of marine taxa, are not uncommon (e.g. Quattro et al. 2001, Tarjuelo et al. 2001, Baker et al. 2003, Papakostas et al. 2005, Remerie et al. 2006), and they have often been related to the presence of cryptic species. *Perna perna* mtDNA sequence data indicate a continuous coastal distribution with cryptic phylogeny, with a 2.9% sequence divergence between the eastern and western lineages. Past work has provided conflicting perspectives on the likely efficacy of mtDNA markers in delineating species boundaries. Some studies, including extensive analyses of GenBank data, have indicated that even closely related species ordinarily show marked mitochondrial divergence (Avice & Walker 1999, Hebert et al. 2003). However, others suggest that mtDNA markers will often encounter problems in species resolution (Funk & Omland 2003, Lipscomb et al. 2003, Mallet & Willmott 2003). For example, Will & Rubinoff (2004) concluded that nearly one-fourth of all animal species fail the test of mitochondrial monophyly. Future studies could be important in determining whether there are previously unnoticed phenotypic differences between the 2 *P. perna* mtDNA lineages, but only breeding experiments and the use of nuclear gene markers would show if the 2 lineages are reproductively isolated. Transplant experiments, i.e. moving different haplotypes between regions, would help us to understand whether adaptation to different environmental condi-

tions is evident. In addition, mtDNA sequences of *P. perna* larvae sampled in the south, warm-temperate and eastern subtropical biogeographic regions would help to determine whether the Agulhas Current operates as an oceanographic barrier, physically limiting dispersal, or if the genetic divergence results from the selective action of different environments.

The invasion of South Africa by *Mytilus galloprovincialis* has had major ecological consequences for the structure of intertidal communities, including the replacement of indigenous species and effects at higher trophic levels. After initially spreading rapidly, its rate of expansion has dramatically decreased. It now has an antitropical distributional range typical of the *M. edulis* group (i.e. *M. edulis*, *M. galloprovincialis*, *M. trossulus*). This supports the prediction that this invasive species has reached its limits in southern Africa and that it will not expand further.

Hydrography is often critical to gene flow among populations, but the patterns identified here are not those that would be predicted by consideration of hydrodynamics alone. The Agulhas Current flows very close inshore in the region of overlap between our lineages, and this could plausibly minimise gene flow by advecting larvae away from the coast. However, this effect would have to be both powerful and continuous, as mussels do not breed in the same month each year, and we must reconcile this with the fact that *Perna perna* on either side of a semi-permanent upwelling cell near Lüderitz in Namibia belong to the same lineage. Indeed, the distribution of haplotypes is antithetical to that expected on the basis of oceanographic data alone. Whatever the mechanism producing the patterns seen in *P. perna* genetics, our results suggest that the factors responsible for the phylogeographic break in *P. perna* populations also limit the distribution of *Mytilus galloprovincialis*, so that defining the population genetic structure of an indigenous species can be important in determining the presence of environmental barriers that could limit the demographic expansion of an invasive species with similar potential dispersal.

**Acknowledgements.** This study was supported by the National Research Foundation of South Africa. We are grateful to K. R. Nicastro, K. Sink, J. Basson, T. Robinson, J. Harris and B. Currie for helping with the sampling.

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*Editorial responsibility: Otto Kinne (Editor-in-Chief), Oldendorf/Luhe, Germany*

*Submitted: June 7, 2006; Accepted: October 31, 2006  
Proofs received from author(s): April 27, 2007*