

Canadian Science Publishing Canadian Journal of Fisheries and Aquatic Sciences

Variation in inhibitor effects on qPCR assays and implications for eDNA surveys

Journal:	Canadian Journal of Fisheries and Aquatic Sciences
Manuscript ID	cjfas-2018-0263.R2
Manuscript Type:	Article
Date Submitted by the Author:	21-Mar-2019
Complete List of Authors:	Lance, Richard; US Army Engineer Research and Development Center, Environmental Laboratory Guan, Xin; Bennett Aerospace, Inc.
Keyword:	PCR inhibition, quantitative real-time PCR, <i>Hypophthalmichthys </i> , aquatic eDNA, environmental DNA
Is the invited manuscript for consideration in a Special Issue? :	Not applicable (regular submission)
	-



1 Variation in inhibitor effects on qPCR assays and implications for eDNA surveys

2

3 Richard F. Lance and Xin Guan

4

- 5 RF Lance: Environmental Laboratory, US Army Engineer Research and Development Center, 3909
- 6 Halls Ferry Road, Vicksburg, MS 39180, USA (email: richard.f.lance@erdc.dren.mil)
- 7 X Guan: Bennett Aerospace, 3909 Halls Ferry Road, Vicksburg, MS 39180, USA (email:

8 xin.guan@erdc.dren.mil)

- 9
- 10 Corresponding Author: Richard F. Lance, CEERD-EPP, 3909 Halls Ferry Road, Vicksburg, MS
- 11 39180; 601-634-3971 (tel); 601-634-4017 (fax); richard.f.lance@erdc.dren.mil (email)

Abstract: Aquatic environmental DNA (eDNA) surveys are sometimes impacted by polymerase chain 13 reaction (PCR) inhibitors. We tested varying concentrations of different inhibitors (humic, phytic, and 14 tannic acids; crude leaf extracts) for impacts on quantitative PCR (qPCR) assays designed for eDNA 15 surveys of bigheaded carp (*Hypophthalmichthys nobilis* and *H. molitrix*). We also tested for inhibition 16 by high concentrations of exogenous DNA, hypothesizing that DNA from increasingly closely-related 17 species would be increasingly inhibitory. All tested inhibitors impacted qPCR, though only at very high 18 concentrations—likely a function, in part, of having used an inhibitor-resistant qPCR solution. Closer 19 phylogenetic relatedness resulted in inhibition at lower exogenous DNA concentrations, but not at 20 relatively close phylogenetic scales. Inhibition was also influenced by the qPCR reporter dye used. 21 Importantly, different qPCR assays responded differently to the same inhibitor concentrations. 22 Implications of these results are that the inclusion of more than one assay for the same target taxa in an 23 eDNA survey may be an important counter-measure against false negatives and that internal positive 24 controls (IPCs) may not, in the absence of efforts to maximize inhibition compatibility, provide useful 25 information about the inhibition of an eDNA assay. 26 27 28 29

- 30
- 31

32 Introduction

Over the last several years there has been a growing interest in the detection of plant and animal 33 DNA in water samples (Ficetola et al. 2008; Dejean et al. 2011; Goldberg et al. 2011; Jerde et a. 2011, 34 Fujiwara et al. 2016). The DNA obtained from these samples is generally termed "environmental 35 DNA" or "eDNA." With typical eDNA sampling, free DNA, DNA-bearing particulates (e.g. clay 36 particles; Goring and Bartholomew 1952) or DNA containing matter (e.g. intact mitochondria) and 37 other undissolved matter are isolated from water samples by passing water through submicron or 38 micron-scale filters, or by centrifuging the material and decanting water away from the resulting pellet 39 (Ficetola et al. 2008, Goldberg et al. 2011). Next, eDNA samples are processed through the following 40 steps: 1) using one of many possible DNA extraction protocols to isolate the DNA into an aqueous 41 solution (or elution), 2) applying an aliquot of the elution to conventional end-point polymerase chain 42 reaction (PCR) or quantitative real-time PCR (qPCR) with taxa-specific primers (and hydrolysis probes 43 for some variants of qPCR; Ficetola et al. 2008; Thomsen et al. 2012), 3) viewing the PCR or qPCR 44 results to see if expected DNA amplicons are apparent and, sometimes, 4) sequencing the PCR 45 amplicon to verify that it corresponds to the targeted species' DNA. In some cases, PCR is used to 46 enrich the eDNA sample for a subsequent next-generation sequencing assay (i.e., DNA metabarcoding; 47 Taberlet et al. 2012) 48

One of the most commonly cited confounding factors in eDNA results is potential PCR inhibition (Goldberg et al. 2011; Jane et al. 2015; McKee et al. 2015*b*). PCR inhibition occurs when a constituent in a PCR reaction – usually an unwanted compound that persists in solution through DNA extraction and isolation – interferes with PCR at a molecular level and either completely prevents the production of the target amplicon or significantly reduces the amount of amplicon that is produced (Opel et al. 2012). Such an effect is particularly worrisome for eDNA surveys, as intact target DNA is typically a

very minor, partly degraded constituent of the overall environmental DNA sample. Because PCR 55 inhibitors can be difficult to completely eliminate or ameliorate, many eDNA survey protocols call for 56 internal positive PCR controls (IPCs). IPCs are comprised of a DNA template and PCR primers for 57 amplifying that template. In the case of hydrolysis probe qPCRs (e.g., TaqMan qPCR), the IPC will 58 also include a short dual-labeled oligonucleotide (i.e., the probe) complimentary to a DNA sequence 59 within the IPC amplicon. IPCs can be run in multiplex PCRs (or qPCRs) with the actual eDNA assay 60 (Schmidt et al. 2013; Turner et al. 2014; Wilson et al. 2014) or as separate IPC-only PCRs (or qPCRs). 61 A reaction is considered inhibited if there is a failure to either amplify the IPC or if an increase in the 62 qPCR Ct (cycle number at which amplification threshold surpasses the background fluorescence 63 threshold) relative to that observed in known clean solutions (e.g., positive controls) is observed. This 64 approach, however, assumes that PCR of the IPC will be impacted by inhibitors in the same manner as 65 the eDNA assay. Other than exploring remedies to PCR inhibition (McKee et al. 2015*a*, Williams et al. 66 2017), and the somewhat standard employment of IPCs, there has been very little effort among eDNA 67 practitioners to better understand the complexities and implications that different inhibitors bring to 68 bear on eDNA practice (but see Stoeckle et al. 2017). 69

Fortunately, some information on PCR inhibition can be obtained from studies of forensic DNA, 70 ancient DNA, water quality science, soil metagenomics, etc. (Wilson 1997; Sørensen et al. 2003; 71 Albers et al. 2013). It is known that PCR inhibitors usually affect PCR through direct interaction with 72 DNA or interference with the DNA polymerase and/or other PCR reagents. There is a large variety of 73 74 PCR inhibitory compounds which can be co-extracted with DNA. For example, tannic acid may bind magnesium, which is a co-factor affecting DNA polymerases, and thus inhibit polymerase activity 75 (Opel et al. 2010). Humic acids interact with the template DNA and the polymerase, subsequently 76 preventing the enzymatic reaction even at low concentrations (Sutlovic et al. 2005, 2008; Opel et al. 77

2008). Recently, Stoeckle et al. (2017) identified humic substances as particularly problematic 78 inhibitors for eDNA surveys. For qPCR methods utilizing double-stranded DNA-binding dyes (e.g., 79 SYBR Green I), humic substances may quench dye fluorescence (Sidstedt et al. 2015). In the same 80 study, hydrolysis probe gPCR (e.g. TagMan gPCR) fluorescence was not guenched by humic 81 substances. In both types of qPCR (DNA-binding dye and hydrolysis probe), assay results may be 82 impacted when humic substances inhibit PCR efficiency. Schrader et al. (2012) list exogenous DNA as 83 a potential qPCR inhibitor, though very high concentrations are required to achieve inhibition (Kainz et 84 al. 2000). Such a situation might arise in an eDNA survey when exogenous eDNA from one or more 85 relatively abundant and closely-related taxa is amplified due to sequence complementarity with assay 86 primers, but not detected due to mismatches with the hydrolysis probe sequence (Kainz et al. 2000). 87 Improved understanding of the fundamental, functional, and quantitative effects of PCR inhibition 88 on eDNA surveys has the potential to significantly improve survey planning, data interpretation, and 89 model generation. Our aim is to contribute to this understanding by experimentally characterizing the 90 interplay between several known or suspected inhibitors with a suite of different qPCR assays designed 91 to detect macrobial aquatic species. In this case we focused on several different inhibitors that might be 92 expected to occur in environmental water samples and on qPCR assays designed to detect silver and 93 bighead carp (Hypophthalmichthys molitrix and H. nobilis). These two species have established 94 invasive populations in many areas of the world and have been the subject of considerable eDNA 95 development and study (e.g., Jerde et al. 2011; Turner et al. 2014; Klymus et al. 2015). 96

97

98 Materials and methods

99 Inhibitors and inhibitor preparations

The effects on qPCR of two different general classes of PCR inhibitors that may be co-extracted in 100 aquatic eDNA samples were investigated: 1) compounds derived from plants and 2) high 101 concentrations of exogenous DNA. In the case of compounds derived from plants, we initially tested 102 103 the inhibitory effects of discrete compounds that may be released from decaying organic matter, particularly plant material, in water bodies, namely humic acid, phytic acid, and tannic acid. We further 104 tested the inhibition of qPCR by crude extracts from red maple (Acer rubrum) and black willow (Salix 105 *nigra*) leaves. We selected these two species because they are local (Vicksburg, MS, USA) 106 representatives of tree genera that are typically abundant along river bottomlands where invasive 107 populations of *Hypophthalmichthys* are found in North America, and are thus representative of species 108 that likely contribute decaying leaf biomass, and associated inhibitory substances, into those aquatic 109 systems. In the case of high-concentration DNA we further explore the influence of DNA sequence 110 similarity on PCR inhibition by investigating DNA from three sources of differing evolutionary 111 distance from *Hypophthalmichthys* (closest to most distant; Table S1): golden shiner (*Notemigonus* 112 crysoleucas), salmon (Salmonidae), and the bacterium *Escherichia coli*. Golden shiners are confamilial 113 (Cyprinidae) with Hypophthalmichthys. The effects of increasing concentrations of inhibitors on qPCR 114 were measured as declines in the estimated copy number of template DNA (concentration remained 115 constant), which we term, for the purposes of this study, as "declines in sensitivity." This approach to 116 quantifying the effects of inhibitors on qPCR assays of eDNA samples differs somewhat from the 117 common usage of ΔCq (i.e., change in quantification cycle) as a measure of inhibition (Jane et al. 2015, 118 Goldberg et al. 2016). Nor do we use the common criterion of $\Delta Cq = 3$ as a threshold for considering a 119 sample as "inhibited" (Turner et al. 2015; Goldberg et al. 2016). There is currently no standard 120 criterion among eDNA studies for describing levels of inhibition and we believed that expressing 121 inhibition as changes in estimated copy number would be both sufficient for those readers with qPCR 122 expertise and more meaningful to those readers lacking such expertise. 123

Humic acid sodium salt (Sigma-Aldrich, USA; product # H16752), phytic acid sodium salt hydrate 124 (Sigma-Aldrich; product # P8810; C₆H₁₈O₂₄P₆), and tannic acid powder (Sigma-Aldrich; product # 125 403040; C₇₆H₅₂O₄₆) were each serially diluted and incorporated into 20 µl gPCR solutions in order to 126 test the effects of changing inhibitor concentrations on different qPCR assays. Fresh maple and willow 127 leaves were collected locally and air-dried for two weeks at room temperature, after which 10 g of 128 dried leaf material were boiled in 2 L of pure water for one hour with agitation. A colored aqueous 129 solution resulted — presumably containing qPCR inhibitors — and was then serially diluted as 130 described in Table 1, and then included as 1 µl aliquots in 20 µl qPCRs in order to test the effects of 131 changes in these inhibitor concentrations on the different qPCR assays. An additional test of even more 132 concentrated leaf solutions was conducted by drying down 1 µl, 5 µl, 25 µl and 50 µl of original 133 undiluted solutions in the wells of 384-well plates then running qPCR assays in those wells (0.05X-134 2.5X-fold dilutions; Table 1). 135

Total genomic DNA (gDNA) was extracted from golden shiner tissues using the DNeasy Blood & 136 Tissue Kit (Qiagen, USA) using a NanoDrop 1000 spectrophotometer (Thermo Fisher Scientific, 137 USA). A dilution series (Table 1) with nine concentration classes was then generated. A 10 mg/ml pure 138 stock of salmon sperm DNA was purchased from Life Technologies (Life Technologies; product # 139 15632011) and a dilution series (Table 1) with eight concentration classes was generated. Whole 140 gDNA from overnight culture of E. coli strain K12 ER2738 (New England Biolabs[®], Inc., Ipswich, 141 MA, US) was extracted using the DNeasy Blood and Tissue Kit and quantified using the Nanodrop 142 1000. A dilution series of the E. coli DNA (Table 1) with eight concentration classes was then 143 generated. 144

All dilutions of the inhibitors were carried out using AmbionTM nuclease-free purified water
(Thermo Fischer Scientific, USA). All inhibitor dilution series used in these trials were based on

preliminary range-finding tests that identified the minimum compound concentrations at which qPCR 147 assays were completely inhibited (data not shown). The tested levels of humic, phytic, and tannic acids, 148 which are all classes of dissolved organic carbon (DOC), are very high (about $10^{0}-10^{3}$ X) compared to 149 some of the highest published DOC concentrations (Sobek et al. 2007; Osburn et al. 2011; Mostofa et 150 al. 2013). We assume that these high concentrations were, in large part, required due to our use of 151 Environmental Master Mix 2.0 (Life Technologies) for all assays. Environmental Master Mix 2.0 is 152 well known to provide robust qPCR results in the presence of inhibitors (Jane et al. 2015; Verhaegen et 153 al. 2016) and is likely the most commonly used component in eDNA assays that utilize hydrolysis 154 probe qPCR (e.g., TaqMan qPCR; Turner et al. 2014; Jane et al. 2015; Sigsgaard et al. 2015). As 155 hydrolysis probe qPCR is the recommended (Goldberg et al. 2016) and, likely, most common PCR 156 method utilized for eDNA surveys, and as even qPCRs utilizing Environmental Master Mix 2.0 can be 157 inhibited (Doi et al. 2015; Sigsgaard et al. 2015; Turner et al. 2015), our results are particularly relevant 158 to current eDNA practice. We would also surmise that inhibition may be a function of multiple co-159 acting substances, which may each individually contribute to inhibition while occurring at lower 160 concentrations than we had to use for our trials. We also note that the concentrations of exogenous 161 DNA used in our inhibition trials are very high (about 10^{1} – 10^{4} X) compared to DNA concentrations 162 typically found in total DNA extracts from water samples (Deiner et al. 2015; Djurhuus et al. 2017; 163 Shahraki et al. in press), and even about 3-10X the concentrations that might be expected from a typical 164 PCR yield. In any case, a primary focus of this study was to explore a general trend in how different 165 qPCR assays interact with the same inhibitors, not to detail precise quantitative relationships between 166 particular assays and different inhibitory substances. A key objective then was to simply achieve qPCR 167 inhibition, regardless of the concentrations of inhibitors required to achieve it. 168

170 Hypophthalmichthys eDNA assays and qPCRs

Several published qPCR assays designed for detecting *Hypophthalmichthys* carp were used to test 171 how the effects of inhibitors might vary among different assays (Farrington et al. 2015). Descriptive 172 metrics (Table 2: Table S1 (Supplementary Material)) for each assay and its associated amplicon were 173 determined using Geneious R8 (Biomatters, NZ). Commercial synthesized G-block fragments (IDT, 174 USA) matching the qPCR amplicon sequences were used as DNA templates. A ViiA 7 Real-Time PCR 175 System (Thermo Fisher Scientific) was used for all qPCR trials. The qPCR assays were multiplexed in 176 the following pairs: BH-TM1/BH-TM2, SC-TM4/SC-TM5, and AC-TM1/AC-TM3 (Table 2). Within 177 each multiplex, one primer set would bear the fluorescent reporter dye 6FAMTM and the other would 178 bear the fluorescent reporter dye VIC[®] (Life Technologies). Farrington et al. (2015) found no 179 substantial decline in assay sensitivity when these assays were used in multiplex. Additionally, these 180 multiplex qPCRs provide insight into eDNA assays that are multiplexed with IPCs. All qPCR reactions 181 contained 10 µl of 2X TagMan[®] Environmental Master Mix 2.0, 0.5 µM of each primer, 0.125 µM of 182 probe, 1 μ L of DNA template (1000 copies/ μ l), 1 μ l of select inhibitor solution (except trials with 183 dried-down leaf extract already in PCR plate wells), and ultrapure water to a final volume of 20 µl. 184 DNA template concentrations were calculated based on manufacturer-provided stock concentrations 185 and subsequent dilution factors.. Thermal-cycling profiles for the qPCR tests were as follows: 186 temperature cycling began with an initial denaturation step at 95 °C for 10 min, followed by 40 cycles 187 of 95 °C for 15 s and 60 °C for 1 min. Four replicate qPCRs were run for each combination of assay, 188 inhibitor, and inhibitor concentration. A minimum of four negative controls (water blanks) and at least 189 eight positive controls (no added inhibitor) were run for each assay. Positive controls were run in order 190 to provide a robust baseline for expected copy number estimates in uninhibited qPCRs. Six-fold 5X 191 dilution series (10-31250 copies/µl) of each G-block DNA template were used as quantitation 192

standards. Finally, four no-template qPCRs containing inhibitors, at all tested concentrations, were run
for each type of inhibitor in order to determine if inhibitory substances caused any changes to
background fluorescence.

196

197 Statistical Analysis

For quality control purposes, any qPCR demonstrating a threshold cycle (Ct) outside of the first 198 standard deviation of the mean Ct for the four replicate reactions for each trial (inhibitor concentration 199 + assay) was eliminated from the data set. For each type of inhibitor, the distribution of qPCR estimates 200 of DNA copy number (CN) were plotted against increasing inhibitor concentrations. The differences in 201 how assays responded to various concentrations of inhibitor were assessed using an analysis of 202 variance (ANOVA) model with log(CN) as a function of inhibitor concentration and assay, and 203 interaction effects between inhibitor concentration and assay included in the model. CN values of 0 204 were adjusted to minimal non-zero values (e.g. 0.0001) in order to linearize data using a log(10) 205 transformation. The potential influence of the fluorescent gPCR reporter dyes (6FamTM, Vic[®]; Thermo 206 Fisher Scientific; Table S1) as co-factors in our results were likewise assessed using ANOVA, with 207 log(CN) as a function of inhibitor concentration and dye, with interaction effects between inhibitor 208 concentration and dye included in the model. For each inhibitor we explored the influence of primer-209 probe attributes (%GC content, melting temperature; Table 2) on inhibition using analyses of 210 covariance (ANCOVAs), where the regression coefficient determined earlier for each assay in the 211 preceding ANOVAs comprised the dependent variable. The same type of analyses were conducted for 212 amplicon attributes (length, %GC content, melting temperature; Table 2), as well. All statistical 213 analyses were conducted using R version 3.0.3. 214

216 **Results**

Inhibitors by themselves did not induce any detectable change in background fluorescence levels in anyqPCR assay conducted during the study.

219

220 Humic, phytic, and tannic acids

Humic acid, phytic acid, and tannic acid all inhibited qPCR at some concentration (Fig. 1-3), but 221 patterns of inhibition differed considerably among inhibitors, and to some degree among qPCR assays. 222 Clear inhibition of qPCR, where at least some mean CN estimates were significantly lower than those 223 of control reactions, was observed at 50 ng/µl with humic acid, at 500 ng/µl with phytic acid, and 250 224 ng/µl with tannic acid (Figs. 1-3). With humic acid (Fig. 1), qPCR sensitivity appeared to be unaffected 225 over several dilution classes and then rapidly dropped for some assays at a concentration of 50 ng/µl, 226 with absolute inhibition of all qPCR assays observed at 125 ng/µl. With phytic acid (Fig. 2), there was 227 an immediate increase in inhibition (decline in estimated CN) at 0.5 ng/µl, followed by a leveling off of 228 inhibitor effect until absolute inhibition of all assays at between 500-1000 ng/µl. Tannic acid inhibition 229 (Fig. 3) exhibited the most extreme pattern, with most assays having very rapid declines into absolute 230 inhibition between 125-500 ng/µl tannic acid. 231

The decline in sensitivity with increasing concentrations of humic acid (Fig. 1) significantly fit an exponential decay curve ($R^2 = 0.867$, p < 0.001) with no significant variance among assays, but a significant interaction between inhibitor concentration and assay ACTM3 (p = 0.015). The decline in sensitivity with increasing concentrations of phytic acid (Fig. 2) also significantly fit an exponential decay curve ($R^2 = 0.779$, p < 0.001) with no significant variance among assays, nor any significant inhibitor concentrations. The decline in sensitivity with increasing concentrations of tannic acid (Fig. 3) significantly fit an exponential decay curve ($R^2 = 0.799$, p < 0.001) with a

significant variance among assays (p = 0.050), but no significant inhibitor concentration-assay

240 interactions. Of note, SCTM4 was absolutely inhibited at lower concentrations than all other assays for

all three inhibitors, with the exception of BHTM1 and tannic acid, for which both assays were

absolutely inhibited at a lower concentration than other assays.

243

244 Red maple and black willow extracts

245 Extracts from the leaves of both red maple and black willow inhibited qPCR, but at different levels of dilution (Figs. 4-5). Inhibition with the maple leaf extract appeared to cause a more rapid decline 246 from no observable effect to complete or nearly complete inhibition for all assays than was observed 247 for the willow leaf extract. The decline in sensitivity with increasing concentrations of red maple leaf 248 extract fit an exponential decay curve ($R^2 = 0.691$, p < 0.001) with no significant variance among 249 assays (p = 0.411), and no significant inhibitor concentration-assay interactions. The decline in 250 sensitivity with increasing concentrations of willow leaf extract significantly fit an exponential decay 251 curve ($R^2 = 0.659$, p < 0.001) with a significant variance among assays (p < 0.001), particularly 252 SCTM4 (p < 0.001), but no significant inhibitor concentration-assay interactions. With the black 253 willow trial, a significant variance was found between the two reporter dyes (p = 0.010), with 254 6FAMTM-labeled assays appearing to be more inhibited than VIC[®]-labeled assays (Fig. S1). There were 255 no significant inhibitor concentration-dye interactions observed in this assay. 256

257

258 Golden shiner, salmon, and E. coli DNA

All three DNA types inhibited qPCR, with inhibition and absolute inhibition with golden shiner and

salmon DNA (Figs. 6-7) occurring at much lower concentrations (largely by $600-1,000 \text{ ng/}\mu\text{l}$) than with

E. coli DNA (at approximately $3,250 \text{ ng/}\mu\text{l}$ or greater; Fig. 8). The decline in sensitivity with increasing

concentrations of golden shiner gDNA significantly fit an exponential decay curve ($R^2 = 0.736$, p < 0.736) 262 0.001) with a significant variance among assays (p = 0.025), and significant interactions effects for 263 inhibitor concentration and assays (p < 0.001), including with BHTM2 (p < 0.001) and SCTM5 (p < 0.001) 264 0.001). The decline in sensitivity with increasing concentrations of salmon sperm gDNA significantly 265 fit an exponential decay curve ($R^2 = 0.596$, p < 0.001) with no significant variance among assays (p =266 0.115), nor any significant interactions effects for inhibitor concentration and assays (p = 0.998). The 267 decline in sensitivity with increasing concentrations of *E. coli* gDNA significantly fit an exponential 268 decay curve ($R^2 = 0.707$, p < 0.001) with a significant variance among assays (p < 0.001), including 269 BHTM1 (p < 0.001), BHTM2 (p < 0.001), and SCTM5 (p < 0.042). Significant interactions effects for 270 inhibitor concentration and assays occurred with BHTM1 (p < 0.001), BHTM2 (p < 0.001), and 271 SCTM5 (p < 0.001). In all three trials there were no significant dye effects (p = 0.217-0.814), but for 272 each there were significant inhibitor concentration-dye interactions (p < 0.001). In all three cases, 273 6FAMTM-labeled assays were more inhibited at higher exogenous DNA concentrations than VIC[®]-274 labeled assays (Fig. S2). 275

276

277 **Discussion**

In our study we found, not unexpectedly, that qPCR assays had a strong tendency for exponential decay in sensitivity to target DNA with increasing concentrations of different inhibitors. Interestingly, we found that the two different reporter dyes used for our assays seemed to influence apparent levels of inhibition when leaf extract from black willow was applied to qPCR. This finding contrasts, to some degree, with Sidstedt et al. (2015), who found that, though humic substances inhibited qPCRs, the tested humics did not actually impact the fluorescent properties of the hydrolysis-probe dye. It may be that the willow leaf extract we prepared contained particular compounds that act to quench 6-FAM to a greater degree than VIC. The lack of any significant effect of red maple extract on reporter dyes, along
with similar lack of effect from the other organic carbon inhibitors (humic, tannic, and phytic acid),
reinforces the observation that inhibitor effects on qPCR-based eDNA assays are complex and difficult
to anticipate.

High concentrations of exogenous DNA proved to be inhibitory to qPCR assays, though at 289 concentrations that may be rarely, if ever, encountered with eDNA samples. This inhibition is likely a 290 291 function of PCR constituent saturation similar to what occurs with PCR amplification plateaus (Kainz 2000). The observation that qPCR assays were inhibited to different degrees at different concentrations 292 of fish and E. coli DNA indicates that sequence similarity (Table S1) can play a role in exogenous 293 DNA inhibition of PCR. Given this observation, it is clear that the development of assays that not only 294 do not fluoresce, but also do not amplify DNA from related taxa that co-occur with target taxa, 295 especially if these nontarget taxa are more abundant, should be a critical concern. We further note that, 296 contrary to our expectation, we did not observe a clear difference between inhibition levels resulting 297 from high concentrations of Golden Shiner and salmon DNA (Figs. 6-7). It may be that sequence 298 mismatch levels between Golden Shiner and salmon DNA at loci homologous to bigheaded carp eDNA 299 assays (i.e., primers and probes) were small enough that high concentrations of gDNA from the two 300 species had similar inhibitory effects on those bigheaded carp assays. In any case, the degree of 301 302 phylogenetic relatedness that must be taken into account when designing new assays may have to reach beyond the level of genera and families. We observed such a phenomenon during the design and testing 303 of conventional, endpoint PCR assays for bighead carp, silver carp, and the black carp 304 (Mylopharyngodon piceus; R. Lance and X. Guan, personal observation). Also of note, based on the 305 inhibition of our assays by *E. coli* gDNA, even temporary binding interactions between qPCR 306 oligonucleotides (i.e., primers and probes) and exogenous DNA template with which PCR primers have 307 poor complementarity can reduce the rate at which assay oligonucleotides bind to target loci (Kainz 308

2000). We also found that inhibition resulting from high exogenous DNA concentrations was influenced by DNA concentration-reporter dye interactions. We do not know why this effect emerged in our trials, but, again, such observations reflect the challenge of conducting molecular-level analyses on complex environmental samples. In any case, considering the very high levels of exogenous DNA required, as a single factor, to bring about detectable qPCR inhibition in our trials, it may be that exogenous DNA either only functions as co-factor in eDNA assay inhibition or is almost never an issue.

In terms of eDNA surveys, at a basic level, inhibitors can cause false negative results in eDNA 316 surveys and result in misinformed management decisions. Models of the relationships between target 317 organism presence (or abundance or biomass) and the probability of eDNA detection or eDNA copy 318 number (e.g., Schmidt et al. 2013; Schultz and Lance 2015; Chambert et al. 2018) would also be 319 confounded by inhibition. Currently, one way the problem of PCR inhibitors in eDNA samples is dealt 320 with is by attempting to select DNA extraction/isolation kits and protocols that are comparatively more 321 effective in removing inhibitory substances (Goldberg et al. 2015; McKee et al. 2015*a*; Eichmiller et al. 322 2016). However, the typically low and highly variable DNA yields obtained from environmental 323 samples using most kits and protocols make it very difficult to surmise whether PCR suffers from 324 inhibition or from little or no template DNA. Other approaches include either diluting DNA elutions in 325 hopes that dilution of inhibitory substances will minimize their impacts on PCR or incorporating a 326 second DNA clean-up (e.g. spin column purification) in an attempt to further minimize the 327 concentration of inhibitors in the final DNA elution (Pilliod et al. 2013; McKee et al. 2015a; Eichmiller 328 et al. 2016). However, as target eDNA tends to be a very scarce constituent of eDNA samples (there 329 are often fewer than 5 copies of target DNA per μ l of elution; R. Lance, personal observation), 330 additional dilutions could result in qPCR aliquots lacking target DNA template and, thus, false negative 331 results (McKee et al. 2015*a*). Additional cleaning steps also can lead to false negatives, as such steps 332

nearly always fail to recover all the DNA — and often result in the loss of a majority of DNA — in the 333 original elution (Sørensen et al. 2003; McKee et al. 2015*a*; R. Lance, personal observation). In some 334 cases, inhibitor effects may be ameliorated using measures such as optimizing PCR component 335 336 concentrations (e.g., MgCl₂), amending PCR with compounds that can reduce inhibitor effects (e.g. bovine serum albumin (BSA); Dejean et al. 2012; Schmidt et al. 2013; Deiner et al. 2015), utilizing 337 DNA polymerases that are more robust in the presence of inhibitors, etc. These PCR modifications, 338 however, often do not completely ameliorate inhibitor effects, and their utility varies with the type of 339 inhibitors present (Albers et al. 2013). Furthermore, water chemistries of systems from which eDNA 340 samples are obtained typically are not characterized and samples may contain a mixture of different 341 inhibitor classes. 342

A recurring pattern in our study was considerable variance in how different qPCR assays respond to 343 the same inhibitor. Though not widely discussed in the eDNA literature, other studies have observed 344 that different inhibitors impact eDNA assays in different ways (McKee et al. 2015a). Given the 345 complexities in inhibitor constituencies that are likely to be found in environmental matrices (e.g., 346 water, sediment), the variance in how different qPCR assays are affected is likely to be quite large. 347 Under such conditions, if multiple assays targeting the same species are used in an eDNA survey it may 348 be that one or more assays are considerably inhibited, while other assay(s) may not be — providing 349 detections that might otherwise be missed. This strategy, currently employed in eDNA monitoring for 350 Hypophthalmichthys carp in the U.S. (USFWS 2018), reduces the risk of false negative results due to 351 qPCR inhibition. Multiple assays could be run either in a multiplex qPCR or in separate qPCRs. 352 In order to save on sample processing time and costs, and the losses of DNA, associated with 353 additional sample treatments or sample dilution, researchers often incorporate IPCs into qPCR in order 354 to identify those samples that are likely inhibited. Typically, a single qPCR IPC assay, either purchased 355

from a commercial provider or designed in-house, is employed for some or all assays. A significant 356 implication arising from the observation that assays may vary considerably in how they are affected by 357 inhibitors is that, in some cases, the expectation that qPCR eDNA assays and associated IPCs are 358 responding to inhibitory factors in the same fashion, and that IPCs are providing useful information 359 about assay inhibition, may be unfounded. For example, considering the apparently rapid shift from no 360 significant inhibition to notable levels of inhibition for some inhibitor-assay combinations (Fig. 1-8), 361 there will be situations where an IPC may show no notable inhibition, while the assay of interest is 362 strongly inhibited and vice versa. This problematic issue had been noted previously in fields outside of 363 eDNA. For example, Huggett et al. (2008) conducted a series of experiments demonstrating that 364 different assays responded differently to inhibitors associated with human urine samples. These 365 researchers noted, "If two different PCR reactions are to be compared, or one is to be used as a 366 reference reaction for the other ... it is important that the two reactions are affected by potential 367 inhibitors to the same extent . . ." An IPC would be a clear instance of one assay being used as a 368 "reference reaction for the other." Huggett et al. (2008) further coined a term, "inhibition 369 compatibility", as a way of describing how well matched two or more assays are in their responses to 370 the same inhibitor (and to the same concentrations of the same inhibitors). Until now, however, and to 371 the best of our knowledge, the issue of inhibition compatibility, and its implications for IPCs, have not 372 been explicitly addressed in the field of eDNA surveys and monitoring. 373

Huggett et al. (2008) also state "Recognition of the importance of assessing inhibition compatibility should contribute to reducing error and increasing accuracy in both gene expression studies and PCRbased molecular diagnostics." Likewise, it seems important then that more effective IPC strategies be developed among eDNA practitioners. Methods for designing IPC assays that maximize inhibition compatibility between eDNA assays and IPCs will need to be devised. Our study results indicate that

simply taking into account amplicon lengths, GC contents, and/or primer, probe, and amplicon melting 379 temperatures will be insufficient. These findings reflect those Huggett et al. (2008), who also found no 380 correlations between inhibition and several other primer and amplicon features (e.g., secondary 381 structure). Likewise, common expectations for how IPCs are vetted and tested for use in eDNA surveys 382 will need to emerge. If nothing else, it is advisable that eDNA studies utilizing IPCs compare the 383 performances of IPCs and planned assays in the presence of some classes of inhibitors expected to be 384 present in eDNA samples. In summary, the field of aquatic macrobial eDNA monitoring is rapidly 385 evolving and becoming an increasingly important tool for ecologists. New approaches for minimizing 386 assay inhibition and for accounting for its effects on survey results will significantly advance the role of 387 eDNA surveys in conservation and natural resources management. 388

389

390 Acknowledgments

We are grateful for a review of an earlier version of this manuscript provided by C. Jung, and for M. 391 Jung and N. Beane for help with leaf collection. We express thanks to the Government eDNA Working 392 Group and our ERDC Environmental Processes Branch colleagues that provided much needed input on 393 study design and data interpretation. Three anonymous reviewers provided comments and 394 recommendations that were tremendously helpful. We are indebted to K. Baerwaldt for continued 395 leadership and assistance with funding for eDNA research. This project was funded by the Great Lakes 396 Restoration Initiative (GLRI), as administered by the United States Environmental Protection Agency, 397 and overseen by the GLRI Asian Carp Coordinating Committee. Views expressed in this manuscript 398 are those of the authors' and do not reflect the official policy or position of the Department of the 399 Army, Department of Defense, or the U.S. Government. Use of trade, product, or firm names in this 400 study is for descriptive purposes only and does not imply endorsement by the U.S. Government. 401

403 **References**

- 404 Albers, C.N., Jensen, A., Bælum, J., and Jacobsen, C.S. 2013. Inhibition of DNA polymerases used in
- 405 Q-PCR by structurally different soil-derived humic substances. Geomicrobiol. J. 30:675-681.
- 406 Benson, D. A., M. Cavanaugh, K. Clark, I. Karsch-Mizrachi, D. J. Lipman, J. Ostell, and E. W. Sayers.
- 407 2013. GenBank. Nucleic Acids Res. 41:D36-42.
- Chambert, T., Pilliod, D.S., Goldberg, C.S., Doi, H., and Takahara, T. 2018. An analytical framework
 for estimating aquatic species density from environmental DNA. Ecol. Evol. 8:3468-3477.
- 410 Deiner, K., Walser, J., Mächler, E., and Altermatt, F. 2015. Choice of capture and extraction methods
- affect detection of freshwater biodiversity from environmental DNA. Biol. Conserv. 183:53-63.
- 412 Dejean, T., Valentini, A., Miquel, C., Taberlet, P., Bellemain, E., and Miaud, C. 2012. Improved
- detection of an alien invasive species through environmental DNA barcoding: The example of the
 American bullfrog *Lithobates catesbeianus*. J. Appl. Ecol. 49:953-959.
- 415 Djurhuus, A., Port, J., Closek, C.J., Yamahara, K.M., Romero-Maraccini, O., Walz, K.R., Goldsmith,
- D.B., Michisaki, R., Breitbart, M., and Boehm, A.B. 2017. Evaluation of filtration and DNA
- 417 extraction methods for environmental DNA biodiversity assessments across multiple trophic levels.
- 418 Front. Mar. Sci. 4:314.
- 419 Doi, H., Takahara, T., Minamoto, T., Matsuhashi, S., Uchii, K., and Yamanaka, H. 2015. Droplet
- digital polymerase chain reaction (PCR) outperforms real-time PCR in the detection of
- 421 environmental DNA from an invasive fish species. Environ. Sci. Technol. 49:5601-5608.
- 422 Eichmiller, J.J., Miller, L.M., and Sorensen, P.W. 2016. Optimizing techniques to capture and extract
- 423 environmental DNA for detection and quantification of fish. Mol. Ecol. Resour. 16:56-68.
- 424 Farrington, H.L., Edwards, C.E., Bartron, M., and Lance, R.F. 2017. Phylogeography and population
- genetics of introduced silver carp (*Hypophthalmichthys molitrix*) and bighead carp (*H. nobilis*) in
- 426 North America. Biol. Invasions 19:2789-2811.

- 427 Farrington, H.L., Edwards, C.E., Guan, X., Carr, M.R., Baerwaldt, K., and Lance, R.F. 2015.
- 428 Mitochondrial genome sequencing and development of genetic assays for the detection of DNA of
- 429 invasive bighead and silver carp (*Hypophthalmichthys nobilis* and *H. molitrix*) in environmental
- 430 water samples from the united states. PloS One 10:e0117803.
- 431 Fujiwara, A., S. Matsuhashi, H. Doi, S. Yamamoto, and T. Minamoto. 2016. Use of environmental
- 432 DNA to survey the distribution of an invasive submerged plant in ponds. Freshw. Sci. 35:748-754.
- 433 Goldberg, C.S., Pilliod, D.S., Arkle, R.S., and Waits, L.P. 2011. Molecular detection of vertebrates in
- 434 stream water: A demonstration using rocky mountain tailed frogs and idaho giant salamanders. PloS
- 435 One 6:e22746.
- 436 Goldberg, C.S., Strickler, K.M., and Pilliod, D.S. 2014. Moving environmental DNA methods from
- 437 concept to practice for monitoring aquatic macroorganisms. Biol. Conserv. 183:1-3.
- 438 Goldberg, C.S., Turner, C.R., Deiner, K., Klymus, K.E., Thomsen, P.F., Murphy, M.A., Spear, S.F.,
- 439 McKee, A., Oyler-McCance, S.J., Cornman, R.S., Laramie, M.B., Mahon, A.R., Lance, R.F.,
- 440 Pilliod, D.S., Strickler, K.M., Waits, L.P., Fremier, A.K., Takahara, T., Herder, J.E., and Taberlet, P.
- 441 2016. Critical considerations for the application of environmental DNA methods to detect aquatic
- 442 species. Methods Ecol. Evol. 7:1299-1307.
- 443 Goring, C., and Bartholomew, W.V. 1952. Adsorption of mononucleotides, nucleic acids, and
- nucleoproteins by clays. Soil Sci. 74:149-164.
- Huggett, J.F., Novak, T., Garson, J.A., Green, C., Morris-Jones, S.D., Miller, R.F., and Zumla, A.
- 446 2008. Differential susceptibility of PCR reactions to inhibitors: An important and unrecognised
- 447 phenomenon. BMC Res. Notes 1:70.
- Jane, S.F., Wilcox, T.M., McKelvey, K.S., Young, M.K., Schwartz, M.K., Lowe, W.H., Letcher, B.H.,
- and Whiteley, A.R. 2015. Distance, flow and PCR inhibition: eDNA dynamics in two headwater
- 450 streams. Mol. Ecol. Resour. 15: 216-227.

- 451 Kainz, P. 2000. The PCR plateau phase towards an understanding of its limitations. Biochim.
- 452 Biophys. Acta 1494:23-27.
- 453 Klymus KE, Richter CA, Chapman DC, Paukert C. 2015. Quantification of eDNA shedding rates from
- 454 invasive bighead carp *Hypophthalmichthys nobilis* and silver carp *Hypophthalmichthys molitrix*.
- 455 Biol. Conserv. 183:77-84
- 456 McKee, A.M., Spear, S.F., and Pierson, T.W. 2015a. The effect of dilution and the use of a post-
- 457 extraction nucleic acid purification column on the accuracy, precision, and inhibition of
- 458 environmental DNA samples. Biol. Conserv. 183:70-76.
- 459 McKee, A.M., Calhoun, D.L., Barichivich, W.J., Spear, S.F., Goldberg, C.S., and Glenn, T.C. 2015b.
- Assessment of environmental DNA for detecting presence of imperiled aquatic amphibian species in
- isolated wetlands. J. Fish Wildl. Manag. 6:498-510.
- 462 Mostofa, K.M., Liu, C., Mottaleb, M.A., Wan, G., Ogawa, H., Vione, D., Yoshioka, T., and Wu, F.
- 463 2013. Dissolved organic matter in natural waters. *In* Photobiogeochemistry of organic matter. *Edited*
- 464 by K. M. G. Mostofa, T. Yoshioka, A. Mottaleb, and D. Vione. Springer-Verlag, Berlin, DE. pp. 1-
- 465 137.
- Opel, K.L., Chung, D., and McCord, B.R. 2010. A study of PCR inhibition mechanisms using real
 time PCR. J. Forensic Sci. 55: 25-33.
- Osburn, C.L., Wigdahl, C.R., Fritz, S.C., and Saros, J.E. 2011. Dissolved organic matter composition
 and photoreactivity in prairie lakes of the U.S. great plains. Limnol. Oceanogr. 56:2371-2390.
- 470 Pilliod, D.S., Goldberg, C.S., Arkle, R.S., and Waits, L.P. 2013. Estimating occupancy and abundance
- 471 of stream amphibians using environmental DNA from filtered water samples. Can. J. Fish. Aquat.
- 472 Sci. 70:1123–1130.
- 473 R Development Core Team. 2014. R: a language and environment for statistical computing. R
- 474 Foundation for Statistical Computing, Vienna, Austria.

- 475 Schmidt, B.R., Kery, M., Ursenbacher, S., Hyman, O.J., and Collins, J.P. 2013. Site occupancy models
- in the analysis of environmental DNA presence/absence surveys: A case study of an emerging
- amphibian pathogen. Methods Ecol. Evol. 4: 646-653.
- 478 Schrader, C., Schielke, A., Ellerbroek, L., and Johne, R. 2012. PCR inhibitors-occurrence, properties
- and removal. J. Appl. Microbiol. 113:1014-1026.
- 480 Shahraki, A.H., Chaganti, S.R., and Heath, D. In press. Assessing high-throughput environmental
- 481 DNA extraction methods for meta-barcode characterization of aquatic microbial communities. J.
- 482 Water Health
- 483 Sigsgaard, E.E., Carl, H., Møller, P.R., and Thomsen, P.F. 2015. Monitoring the near-extinct european
- 484 weather loach in Denmark based on environmental DNA from water samples. Biol. Conserv. 183:
 485 46-52.
- 486 Sobek, S., Tranvik, L.J., Prairie, Y.T., Kortelainen, P., and Cole, J.J. 2007. Patterns and regulation of
- dissolved organic carbon: An analysis of 7,500 widely distributed lakes. Limnol. Oceanogr.
 52:1208-1219.
- - 489 Sørensen, E., Hansen, S.H., Eriksen, B., and Morling, N. 2003. Applications of thiopropyl sepharose
 - 6B for removal of PCR inhibitors from DNA extracts from different sources. Int. Congr. Ser.
 1239:821-823.
 - 492 Stoeckle, B.C., Beggel, S., Cerwenka, A.F., Motivans, E., Kuehn, R., and Geist, J. 2017. A systematic
 - approach to evaluate the influence of environmental conditions on eDNA detection success in
 - aquatic ecosystems. PloS One 12:e0189119.
 - 495 Tamariz, J., Voynarovska, K., Prinz, M., and Caragine, T. 2006. The application of ultraviolet
 - 496 irradiation to exogenous sources of DNA in plasticware and water for the amplification of low copy
 - 497 number DNA. J. Forensic Sci. 51:790-794.

- 498 Turner, C.R., Miller, D.J., Coyne, K.J., and Corush, J. 2014. Improved methods for capture, extraction,
- 499 and quantitative assay of environmental DNA from Asian bigheaded carp (*Hypophthalmichthys*
- spp.). PloS One 9: e114329.
- 501 Turner, C.R., Uy, K.L., and Everhart, R.C. 2015. Fish environmental DNA is more concentrated in
- aquatic sediments than surface water. Biol. Conserv. 183: 93-102.
- 503 USFWS. 2018. Quality assurance project plan eDNA monitoring of bighead and silver carps.
- 504 United States Fish & Wildlife Service, Bloomington, MN.
- 505 Verhaegen, B., De Reu, K., De Zutter, L., Verstraete, K., Heyndrickx, M., and Van Coillie, E. 2016.
- 506 Comparison of droplet digital PCR and qPCR for the quantification of shiga toxin-producing
- *Escherichia coli* in bovine feces. Toxins 8:157.
- 508 Williams, K.E., Huyvaert, K.P., and Piaggio, A.J. 2017. Clearing muddled waters: Capture of
- environmental DNA from turbid waters. PloS One. 12:e0179282.
- 510 Wilson, C., Wright, E., Bronnenhuber, J., MacDonald, F., Belore, M., Locke, B., MacNeil, C., and
- 511 Campbell, M. 2014. Tracking ghosts: Combined electrofishing and environmental DNA surveillance
- efforts for Asian carps in Ontario waters of Lake Erie. Manag. Biol. Invasion 5:225-231.
- Wilson, I.G. 1997. Inhibition and facilitation of nucleic acid amplification. Appl. Environ. Microbiol.
 63: 3741-3751.

- 1 **Table 1.** PCR inhibitors used in tests of six *H. molitrix* and *H. nobilis* eDNA
- 2 markers, with details on tested concentrations (within 20 μ l qPCR solutions). X =
- 3 N-fold dilution of a mixed, unquantified inhibitor stock.

Inhibitor	Dilution series			
	units	concentrations		
Humic acid	ng/µl	250, 125, 50, 25, 5, 2.5, 0.5		
Phytic acid	ng/µl	2500, 1000, 500, 250, 50, 12.5, 2.5, 0.5		
Tannic acid	ng/µl	750, 500, 250, 125, 50, 12.5, 2.5, 0.5		
Red maple	Х	2.5, 1.25, 0.25, 0.05, 0.01, 0.002, 0.0004		
Black willow	Х	2.5, 1.25, 0.25, 0.05, 0.01, 0.002, 0.0004		
Golden shiner DNA	ng/µl	900, 600, 300, 100, 47.5, 31, 16.8, 8.4, 4.2		
Salmon DNA	ng/µl	1500, 1000, 500, 250, 125, 50, 25, 5		
E. coli DNA	ng/µl	3250, 1625, 975, 530, 265, 115, 50, 25		

- Table 2. The qPCR markers (i.e., primers and probes) for bighead carp (BHC) and 4
- silver carp (SC) that were used to characterize the effects of different PCR inhibitors 5
- on qPCR. 6

	Target	Primer & Probe		Amp		Amp
Marker	species	GC%	Tm	BP	Amp GC%	Tm
SC-TM4	SC	50.0	56.1	169	46.2	79.8
SC-TM5	SC	47.6	58.3	99	43.4	75.9
BH-TM1	BHC	46.6	56.4	145	44.1	78.4
BH-TM2	BHC	50.0	60.2	97	45.0	77.0
AC-TM1	Both	46.7	60.2	146	47.3	79.7
AC-TM3	Both	46.8	56.6	134	45.5	78.5

Note: Primer & Probe GC% details the mean percent GC nucleotide content of the forward primer, reverse primer, and hydrolysis probe. Tm and Amp TM detail the 8 melting temperatures (°C) of the primer-probe set (mean value) and resulting 9 amplicon, respectively. Amp BP and Amp GC% detail the length (in DNA base pairs) 10 and percent GC nucleotide content of the resulting amplicons. Amplicon statistics 11 based on National Center for Biotechnology Information GenBank (Benson et al. 12 2013) DNA sequence accessions KR756343.1 and KJ729076.1 for BHC and SC 13 respectively. 14

1	Figure 1. The effects of increasing concentrations of humic acid on qPCR estimates of original DNA
2	template copy number (CN; calculated at 1000 copies across all treatments). Lines represent trends in
3	mean estimated CN for each marker.
4	
5	Figure 2. The effects of increasing concentrations of phytic acid on qPCR estimates of original DNA
6	template copy number (calculated at 1000 copies across all treatments). Lines represent trends in mean
7	estimated CN for each marker.
8	
9	Figure 3. The effects of increasing concentrations of tannic acid on qPCR estimates of original DNA
10	template copy number (calculated at 1000 copies across all treatments).
11	
12	Figure 4. The effects of increasing concentrations of maple leaf extract on qPCR estimates of original
13	DNA template copy number (calculated at 1000 copies across all treatments). Lines represent trends in
14	mean estimated CN for each marker.
15	
16	
17	Figure 5. The effects of increasing concentrations of willow leaf extract on qPCR estimates of original
18	DNA template copy number (calculated at 1000 copies across all treatments). Lines represent trends in
19	mean estimated CN for each marker.
20	
21	Figure 6. The effects of increasing concentrations of golden shiner DNA on qPCR estimates of original
22	DNA template copy number (calculated at 1000 copies across all treatments). Lines represent trends in
23	mean estimated CN for each marker.
24	

25	Figure 7. The effects of increasing concentrations of salmon DNA on qPCR estimates of original DNA
26	template copy number (calculated at 1000 copies across all treatments). Lines represent trends in mean
27	estimated CN for each marker.

- Figure 8. The effects of increasing concentrations of E. coli DNA on qPCR estimates of original DNA
- template copy number (calculated at 1000 copies across all treatments). Lines represent trends in mean
- 31 estimated CN for each marker.

32 Fig 1.



3

35





38





40

7







10