



Variations in Mass of the PM₁₀, PM_{2.5} and PM₁ during the Monsoon and the Winter at New Delhi

S. Tiwari, D.M. Chate^{*}, P. Pragma, Kaushar Ali, Deewan Singh Bisht

Indian Institute of Tropical Meteorology, Pune - 411 008, India

ABSTRACT

PM₁₀, PM_{2.5} and PM₁ mass concentrations have been measured at Delhi (28°35'N; 77°12'E) during the August to December 2007. The running mean of PM₁₀, PM_{2.5} and PM₁ data shows large variations. The PM₁₀, PM_{2.5} and PM₁ were ranged from 20 to 180 µg/m³ during the monsoon and from 100 to 500 µg/m³ during the winter (up to 1200 µg/m³ in November due to Deepavali fireworks). For the same running mean cycles, higher mass concentrations in the PM₁₀, PM_{2.5} and PM₁ were corresponded with peaks in the relative humidity and lower levels linked to peaks in the ambient temperature. The evolutions of PM₁₀, PM_{2.5} and PM₁ concentrations after the elapsed times are simulated with mean mass scavenging coefficients. These evolution patterns clearly show the difference in washout of PM₁₀ with impaction scavenging relative to those for PM_{2.5} and PM₁ particles over different rainfall durations. Air-mass pathways traced with HYSPLIT model over the study area illustrates the nature of PM₁₀, PM_{2.5} and PM₁ levels with monsoon and winter air-mass circulations over Delhi.

Keywords: Particulate matters; Visibility; Washout pattern; Time-series variations; Mean mass scavenging coefficient.

INTRODUCTION

Atmospheric particulate matter (PM) can be classified as PM₁₀, PM_{2.5} and PM₁ by size with mass median aerodynamic diameter less than 10 µm, 2.5 µm and 1 µm respectively. PM plays pivotal role in the climate change, cloud dynamics, health impact, fog formation and visibility through a variety of atmospheric processes (Pillai *et al.*, 2002; Pope *et al.*, 2002; Das *et al.*, 2009). High concentrations in the PM₁₀, PM_{2.5} and PM₁ can cause human health problems, related to both short-term and long-term exposure to these particles (Schwartz *et al.*, 1996; Massey *et al.*, 2009; Chate, 2010). Based on an epidemiological data, WHO states that an increase in total PM by 10 µg/m³ per year results in a 6% increase in mortality and a short-term increase in levels of the PM₁₀, PM_{2.5} and PM₁ for several days causes more coughing, respiratory problems, bronchodilator use and even mortality (WHO, 2000, 2003). Since health impacts of PM_{2.5} and PM₁ in air are more adverse than larger particles (PM₁₀), measuring them together with PM₁₀ is highly important (Wilson and Suh, 1997). During a dust storm, Choi and Choi, (2008) reported high concentrations in the variation of PM₁₀, PM_{2.5} and PM₁ owing to circulation patterns and

boundary layer heights in the Korean mountainous coast. Li *et al.* (2009) have presented PM_{2.5} particles at an urban, industrial and coastal site in Tianjin, China. Also, recently Sabbagh-Kupelwieser *et al.* (2010) have presented PM₁₀, PM_{2.5} and PM₁ at Vienna, Austria.

In Delhi, India, measurements of aerosol number size distributions have already been performed for air-quality monitoring and aerosol formation purposes. For instance, high concentrations of ultrafine particles (diameter < 0.1 µm) were frequently recorded in the urban atmosphere of Delhi (Monkkonen *et al.*, 2004). With the rapid urbanization and corresponding increase in the traffic and energy consumption, there has been growing evidence that ambient concentration levels of PM_{2.5} and PM₁ are also high in Delhi (Gupta *et al.* 2007). The major source of PM₁₀, PM_{2.5} and PM₁ are referred as windblown dust, secondary aerosol, coal combustion, traffic exhausts and biomass burning, etc. (Tiwari *et al.* 2009). Furthermore, PM_{2.5} and PM₁ remain air-borne through nonlinear processes for days-to-weeks during monsoon months as washout processes are least efficient for cleansing particles in these size bins. Since background number concentrations of PM₁₀, PM_{2.5} and PM₁ particles are very high in megacities (e.g. Delhi) their formation and removal processes by rainfall are not clearly understood.

Aerosol distributions presented by taking seasonal or annual simple averages of data can suppress the peaks owing to local effects and also by variations with rain scavenging over very short durations (Chate *et al.*, 2005). Therefore, time series distributions of PM₁₀, PM_{2.5} and

^{*} Corresponding author. Tel: +91-20-25904257;
Fax: +91-20-25893825
E-mail address: chate@tropmet.res.in

PM₁ can be presented by performing running mean on raw data in order to address environmental and rain scavenging processes in those size regimes. The National Ambient Air Quality Standards (http://cpcb.nic.in/National_Ambient_Air_Quality_Standards.php) for PM₁₀ and PM_{2.5} are set at averages over 24 hours (1440 minutes). The main advantage in time series presentation of PM₁₀, PM_{2.5} and PM₁ mass concentrations is that a mean can be performed around every point along a continuous raw data. Furthermore, data points with very high peaks for random spikes due to electronic glitches are to be removed from the raw data.

The mass concentration of PM₁₀, PM_{2.5} and PM₁ have been measured continuously at Delhi (28°35'N; 77°12'E) for days-to-weeks period in a month during August–December, 2007. We present 1440 minutes running mean variations in the PM₁₀, PM_{2.5} and PM₁ concentration during the monsoon (August–September) and the winter period (October–December). Differential behavior of PM₁₀ as compared to PM_{2.5} or PM₁ for washout patterns is discussed to understand removal mechanisms in these size regimes. We discuss relative humidity and temperature variations for the available data of these parameters for very short period during August. Also, HYSPLIT (Draxler and Rolph, 2003, <http://www.arl.noaa.gov/ready/hysplit4.html>) backward trajectories aid us in tracing air-mass pathways for likely pollutant source regions. Variations in PM₁₀, PM_{2.5} and PM₁ concentration with 1440 minutes running mean cycles and influences of rain, relative humidity, temperature etc. on these variations are useful in formulating control measures of particulate matter air-quality in Delhi. Also, PM₁₀, PM_{2.5} and PM₁ concentration changes with short and long term atmospheric processes are significant from the environmental point of view due to differential exposure of ambient aerosols to local population.

Observational Site

The observational site, IITM, Delhi (28°35'N; 77°12'E) is located in urbanized central part of Delhi at about 218 m above mean sea level. The transport and dispersion of pollutants, particularly those in the lower levels of the atmosphere, are believed to be affected by the circulation pattern associated with the surrounded area. The site represents a typical residential area free from major pollutant sources, which is suitable for background air pollution studies. The long-term atmospheric measurements of cloud microphysics and air-quality parameters are available in the literature. Delhi experiences a severe cold and foggy weather during winter. The prevailing winds throughout the year are easterly, northerly and northwesterly. The entire northern part of India, especially the Indo-Gangetic Plain, experiences a thick foggy weather during winter with low boundary layer height. The low winds and reduced boundary layer heights results in very low ventilation factors which determine the dispersion of pollutants. During such conditions, pollutants could not be dispersed or mixed with free troposphere (Srivastava *et al.*, 2005). The impact of such conditions is poor visibility and high levels of pollutants in this region. The daily minimum temperature falls rapidly by the end of October and continental air masses, rich in

pollutants of continental origin, pass over the experimental site during winter.

METHOD

The sampling of aerosols for this study was carried out at about 15 m above the ground level, on the rooftop of an IITM Building (Delhi) situated in the urbanized central part of Delhi. The area is primarily a residential area, and no large pollutant source exists nearby which could have influenced the sampling site directly.

The GRIMM Model 1.108 (OPC, GRIMM Inc.) is a portable particle analyzer and is specifically designed for PM₁₀, PM_{2.5} and PM₁ ambient air analysis using dual technology consisting of both optical and gravimetric analysis. This technology enables the Model 1.108 to make precise cut off diameters for all three PM sizes. This system allows collecting all three PM fractions simultaneously without changing sampling heads. Coarse particles (PM₁₀) and fine particulates (PM_{2.5} and PM₁) have been monitored with the GRIMM particles sampler. The GRIMM particle counter was operated continuously during August 2007 to December 2007. A constant flow rate ~1.2 L/min is maintained throughout the measurements. The GRIMM particles measuring system is equipped with GRIMM 1174 Software for data acquisition. It was set to collect data at 1 minute intervals and store them in memory to be downloaded to a PC and analyzed further for time-series during August to December 2007. The data of relative humidity and ambient temperature during August, 2007 were collected through an automatic weather station. The variations in relative humidity (RH) with the temperature over a very short period at an experimental site during 29–31 August, 2007 are presented in Fig. 1(a). During monsoon month of August, RH varied between 50% and 90% with the temperature variations. As seen in this Figure, at lower temperature ~27°C, the RH found to be at its maxima (90%) and for higher temperature ~35°C, the RH drops below 50% and vice versa.

RESULTS AND DISCUSSION

We have performed a mean around every point along a continuous raw data of PM₁₀, PM_{2.5} and PM₁ mass concentrations obtained during August–December 2007. The running mean is performed with a time constant of 1440 minutes, or 1440 data points calculating the mean of the first 1440 points, then subtracting the value of first point, and adding the value of 1441st point for the next mean value. Eq. (1), computes running means in time-series data of PM₁₀, PM_{2.5} and PM₁ mass concentrations over several days.

$$y_{j+1} = y_j + \frac{(x_{j+1} - y_j)}{j+1} \quad (1)$$

where x – data point in original time series, y – data point in running mean time series and j – position of the data point. The time constant of 1440 minutes found to be suitable for the running 1440 minutes PM₁₀, PM_{2.5} and PM₁ mass

concentrations means, as the original 1 minute time-series data are of very high fluctuations owing to random spikes by electronic/electrical glitches. The National Ambient Air Quality Standards for PM_{10} and $PM_{2.5}$ are set at $100 \mu\text{g}/\text{m}^3$ and $60 \mu\text{g}/\text{m}^3$ respectively over 1440 minutes average.

Accordingly, the 1440 minutes running mean variation in PM_{10} , $PM_{2.5}$ and PM_1 mass concentrations for the rainy months (August and September) are shown in Figs. 1(b) and (c) respectively. The winter months (October, November and December) time-series for PM_{10} , $PM_{2.5}$ and PM_1 mass

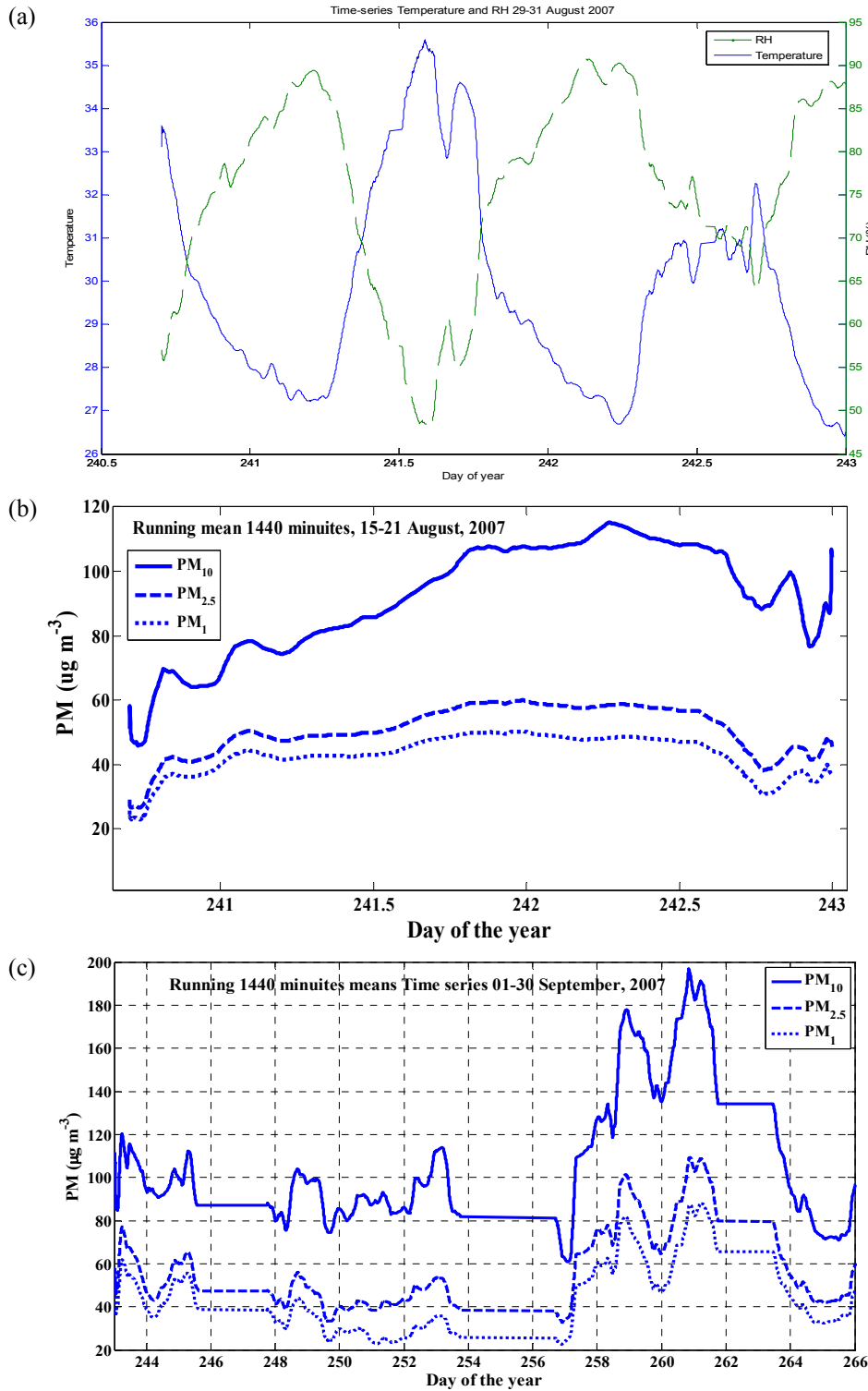


Fig. 1. (a) Relative humidity with ambient temperature during 29–31 August, 2007, (b) Time-series for PM_{10} , $PM_{2.5}$ and PM_1 mass concentrations during 15–21 August, 2007 at 1440 minutes running mean, (c) Same as (b) during 01–30 September, 2007.

concentrations are shown in Figs. 2(a–c) respectively. Higher levels of PM₁₀, PM_{2.5} and PM₁ mass concentrations (1200–1400 µg/m³) were observed on Deepawali days (8th–9th November 2007) owing to large emissions by fireworks displays (Fig. 2(b)). In the month of December, the ABL height remains shallow (50 m) and almost no winds till 0900 hour results in too low ventilation factors over

continental areas as reported elsewhere (Murugavel and Chate, 2011). Lower boundary layer heights should confine ambient aerosols at the Earth surface and manifest as an increase in their levels. Low PM₁₀, PM_{2.5} and PM₁ mass concentrations were observed during 15–21 August (20–110 µg/m³) and 01–30 September (20–120/180 µg/m³) as against those observed during December due to the washout of these particles by

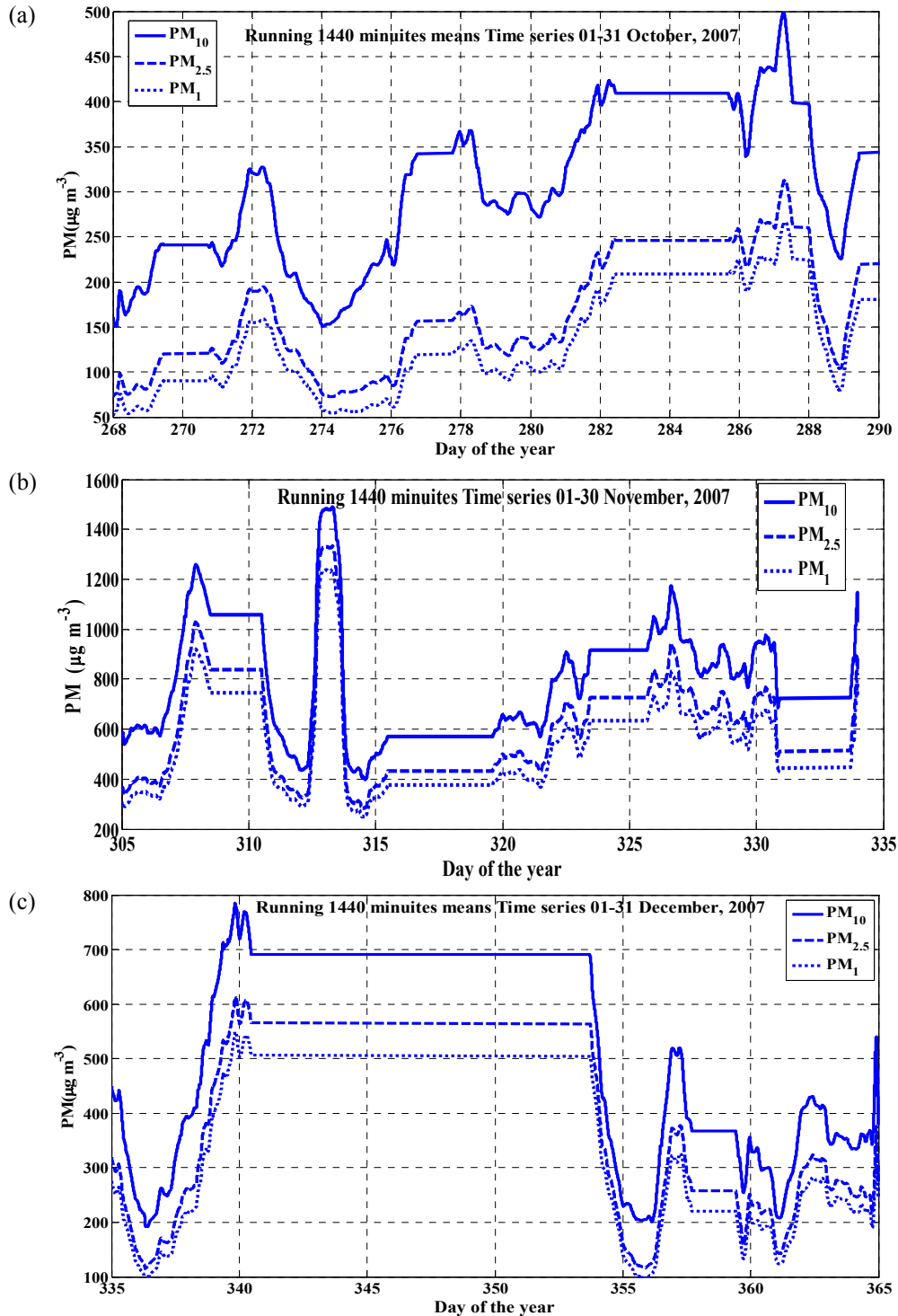


Fig. 2. (a) Time-series for PM₁₀, PM_{2.5} and PM₁ mass concentrations during 01–24 October, 2007, (b) Same as (a) during 01–30 November, 2007, (c) Same as (a) and (b) during 01–31 December, 2007.

rainfall (Figs. 1(b) and (c)). The mass concentration is reduced considerably in August, due to washout effect during precipitation (Fig. 1 (b)). It can also be seen from Fig. 2(a–c) that during the months of continental air-mass (winter), levels of PM₁₀, PM_{2.5} and PM₁ lie close by, whereas during the monsoon months, these curves are well resolved, with PM₁₀ much higher than that of PM_{2.5} and PM₁ (Figs. 1(b) and (c)). It is worth to note that mass concentration trends in the monsoon months resulted in an order of difference in mass concentrations between PM₁₀ and PM₁ or PM_{2.5} due to the extensive monsoon rain, replenishment of PM₁₀ aerosols by stronger monsoonal winds and superimpositions of aerosols by splashing of raindrops on the Earth surface (Chate and Kamra, 1993). Variations in PM₁₀, PM_{2.5} and PM₁ display distinct winter and monsoon months' peaks. Winter months' peaks are attributed to enhancement in continental aerosols and monsoon months' peaks upon impositions of aerosols with splashing of raindrops on the Earth surface.

Role of Rain Scavenging in Distributions of PM₁, PM_{2.5} and PM₁₀ over Delhi

PM_{2.5} and PM₁ particles falls in the accumulation size bins and thus they are air-borne for days-to-weeks even during monsoon months as washout processes are least efficient for cleansing these particles. Also, background number concentrations of PM₁₀, PM_{2.5} and PM₁ particles are very high in Delhi in turn their washout patterns by rainfall are not clearly understood. Therefore to understand washout processes, a typical rain scavenging patterns for PM₁, PM_{2.5} and PM₁₀ are simulated for August for assumed rainfall rates of 5, 25 and 50 mm/h for PM₁₀ and 25 and 50 mm/h for PM₁ and PM_{2.5}. Terms related to Brownian diffusion, directional interception, inertial impaction, thermo-diffusion-phoresis and electro-scavenging with dimensionless parameters used for simulations are published elsewhere with appropriate references (Chate and Devara, 2005). Overall collision efficiency $E(D, D_p)$ computations are performed for PM₁₀, PM_{2.5} and PM₁ particles ($D_p < 10 \mu\text{m}$, $D_p < 2.5 \mu\text{m}$, $D_p < 1 \mu\text{m}$) and raindrops of size (D) for electro-scavenging, Brownian diffusion, inertial impaction, interception, and thermo-diffusion-phoresis.

For a spatially homogeneous system, assuming rain scavenging is the only sink for PM₁₀, PM_{2.5} and PM₁, changes in mass or number concentrations for non-interacting and non-growing particles $N(D_p)$ with diameters between D_p and $D_p + dD_p$ at time t after scavenging is related to the concentration $N_0(D_p)$ at $t = 0$, as

$$N(D_p) = N_0(D_p) \exp[-A(D_p)t] \quad (2)$$

where $A(D_p)$ is the scavenging coefficient (Seinfeld and Pandis, 1998) and expressed as,

$$A(D_p) = \int_0^{\infty} (\pi/4) D^2 E(D_p, D) V_t(D) N(D) dD \quad (3)$$

where, $N(D)$ is raindrop size distribution adopted from classical Marshall–Palmer (1948) drop size distribution. $V_t(D)$

terminal velocities of raindrop computed using Beard's formulae (1976). Mean mass scavenging coefficient A_m is calculated for all the three particle size regimes (PM₁₀, PM_{2.5} and PM₁) at rainfall rates of 5, 25 and 50 mm/h with equation given below,

$$A_m = \frac{\int_0^{\infty} A(D_p) D_p^3 N(D_p) dD_p}{\int_0^{\infty} N(D_p) D_p^3 dD_p} \quad (4)$$

Figs. 3(a) and (b) show washout pattern at rainfall intensity of 25 and 50 mm/h after elapsed time of 5 hours rain for PM₁ particles and 3 hours rain for PM_{2.5} particles. Also, Fig. 3(c) shows the washout pattern for PM₁₀ particles at rainfall intensity of 5, 25 and 50 mm/h after elapsed time of 1 hour (rainfall duration). Since large mass fractions of PM₁ and PM_{2.5} particles resides in the accumulation mode, they are too large to have sufficient Brownian diffusivity and too small to get collected effectively by falling raindrops due to inertial impaction mechanism (see Seinfeld and Pandis, 1998 for wet deposition). The evolutions of PM₁ and PM_{2.5} particles in Figs. 3(a) and (b) show very small fractions of initial mass concentrations of these particles' depletion by rain intensity of 25 and 50 mm/h even after elapsed time of rain over 5 hours and 3 hours respectively. As significant fractions of mass concentrations of PM₁₀ particles are in the size range of coarser mode, they are effectively washed out by rain over 1 hour of elapsed time at intensity of 5, 25 and 50 mm/h due to their inertia (Fig. 3(c)) higher than that of PM₁ or PM_{2.5} particles. The evolution of PM₁₀ particles in Fig. 3(c) clearly shows the dominance of inertial impaction in depleting significant amount of initial mass of these particles by rain over 1 hour at intensity of 5, 25 and 50 mm/h respectively. However, the evolutions of PM₁ and PM_{2.5} particles indicate that, impaction scavenging or Brownian diffusivity is insignificant collection mechanism of raindrops for depleting initial mass of these particles by heavy rainfalls of intensity 25 and 50 mm/h over longer durations (5 and 3 hours for PM₁ and PM_{2.5} particles).

Air-mass pathways are traced to assess possible sources during the monsoon (August–September) and during the winter (December). Air-mass backward trajectories that reached Delhi at 50, 100 and 500m above ground level were traced with Hybrid Single-Particle Lagrangian Trajectory (HYSPLIT, NOAA/ARL) (Draxler, 2004). The global meteorological data from the National Centers for Environmental Prediction's Global Data Assimilation System (GDAS) were used for the trajectory calculation. Back trajectories are traced using a 168-hour (01–07 December) ending at 1900 UTC [(0030 hrs local time (LT))] and 144 hour (01–06 December) ending at 1400 UTC (1930 LT) backward trajectory with every six hour interval. The air-mass pathways for high PM days (6–7 December, 2007) and also relatively low events (3–4 December, 2007) during December are analyzed as shown in Figs. 4(a–f). During winter, air-mass pathways arriving at Delhi are affected by

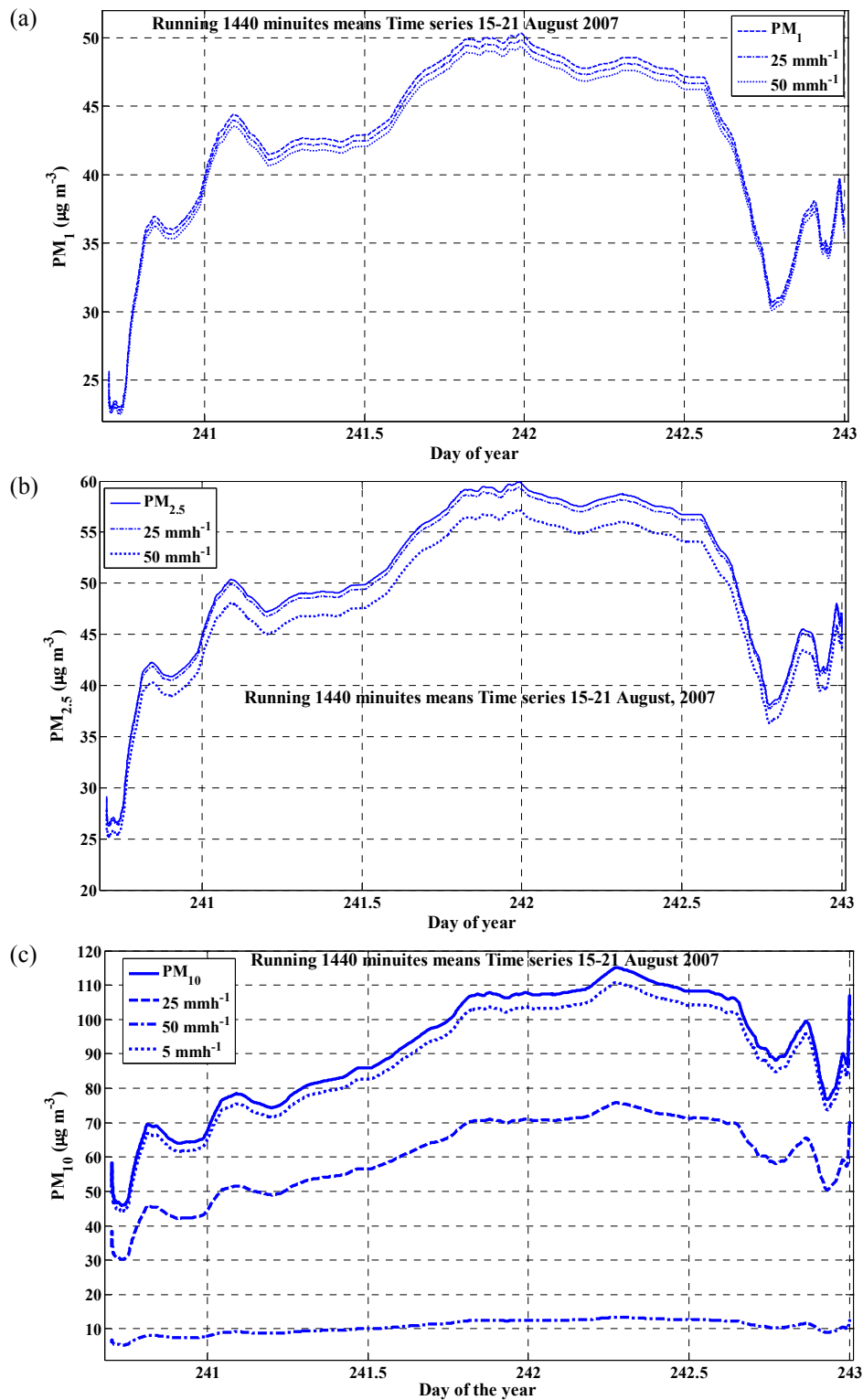


Fig. 3. (a) Washout pattern for PM₁ particles at rain intensity of 25 and 50 mm/h over 5 hours of rainfall, (b) Same as (a) for PM_{2.5} particles for 3 hours of rain, (c) The evolution of PM₁₀ particles by rain over duration 1 hour at intensity, 5, 25 and 50 mm/h.

anthropogenic continental and local aerosols during 01-07 December 2007 as seen in Fig. 4(a-b). The anthropogenic aerosols from continental air-mass favor the increase in fine particle levels (PM₁ and PM_{2.5}). However, during monsoon

months, air-mass loaded with natural marine aerosols with salt contents were passed over Delhi during 29–31 August and during 25–30 September, which enhanced coarser particles (PM₁₀) (Figs. 4(c)–(f)). In addition to these synoptic

air-mass patterns, atmospheric conditions over the Delhi site were affected by episodic events involving continuous heavy precipitation during monsoon (29–31 August) and dense foggy condition during winter (December).

Running mean in PM_{10} , $PM_{2.5}$ and PM_1 mass concentrations during the monsoon and the winter shows much wide variations at Delhi and could have significant adverse impacts on visibility and health owing to short-term local effects. In

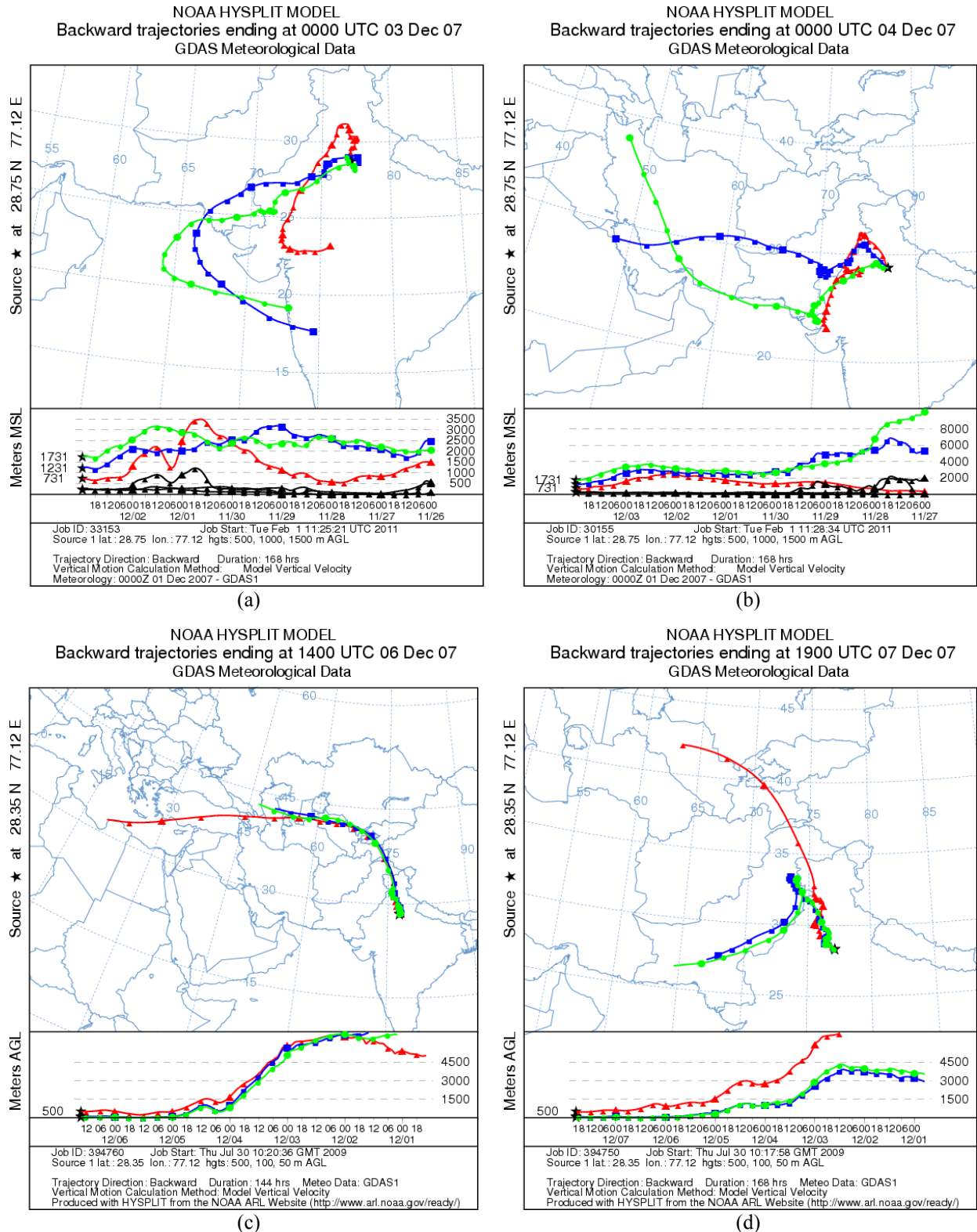


Fig. 4. (a–h) Back-trajectories for air-mass pathways for 03–04 (Fig. 4(a) and (b)), 07–06 (Fig. 4(c) and (d)), December, 2007, 30–28 (Fig. 4(e) and (f)), September, and 29–31 August (Fig. 4(g) and (h)).

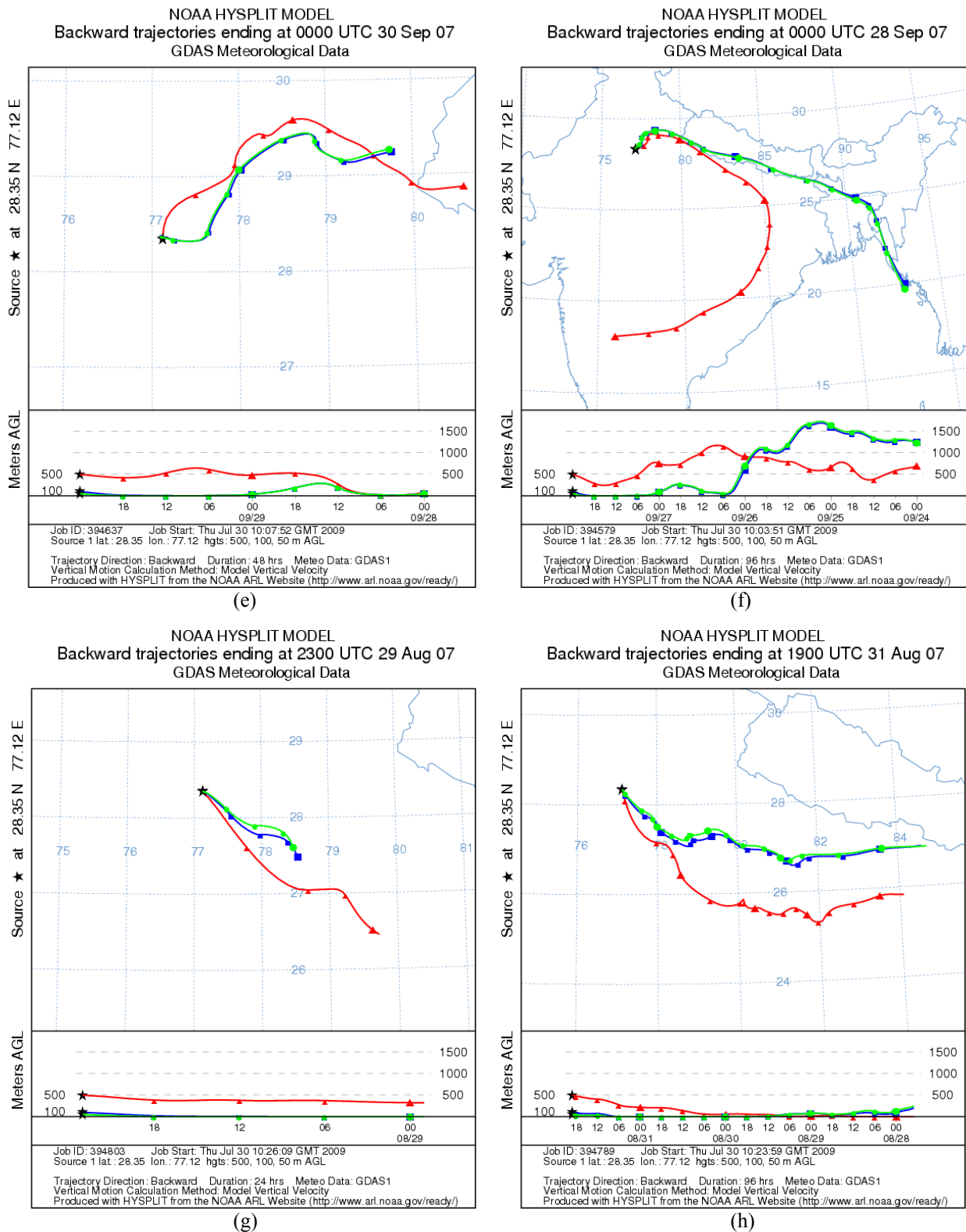


Fig. 4. (continued).

comparison, annual mean PM₁₀ and PM_{2.5} concentration (\pm standard deviation) reported as 219 (\pm 84) and 97 (\pm 56) $\mu\text{g}/\text{m}^3$ respectively during January to December, 2007, are about twice to prescribed Indian National Ambient Air Quality Standards (Tiwari et al., 2009). TSP, PM₁₀ and

PM_{2.5} at Erzurum (Turkey) urban atmosphere were measured on an average of 129, 31 and 13 $\mu\text{g}/\text{m}^3$, respectively for February 2005 to February 2006 (Hanefi et al., 2008). At an urban, industrial and coastal site in Tianjin, China, Li et al. (2009) observed daily average concentrations of PM_{2.5} ranged

from 61.5–566.8 $\mu\text{g}/\text{m}^3$ with an average of 223.0 $\mu\text{g}/\text{m}^3$. However, reducing $\text{PM}_{2.5}$ and PM_1 emissions and so to decreasing their negative effects on health, visibility etc., is a difficult task especially where they are mainly accumulated through nonlinear atmospheric processes.

CONCLUSIONS

In the nature of variation in PM_{10} , $\text{PM}_{2.5}$ and PM_1 , low concentration occurred during the monsoon (August–September) due to washout of these particles, whereas, higher levels observed during the winter (October–December), when lower boundary heights prevails over this site. During the monsoon, since wet removal processes are most effective, PM_{10} , $\text{PM}_{2.5}$ and PM_1 concentrations reaches at their annual minimum. After the monsoon, scenario reverses as air-mass shifts to continental followed by reduction in rainfall and cloudiness and thus PM_{10} , $\text{PM}_{2.5}$ and PM_1 concentration gradually increases. The winter peak is attributed to aerosols in the PM_{10} , $\text{PM}_{2.5}$ and PM_1 levels due to mixing of continental aerosols with background local particles. On the other hand, monsoon peaks are due to the superposition of transported aerosols from marine environment on local continental aerosols and on aerosols due to splashing of raindrops on the Earth surface during rain. The information presented on running mean variations in PM_{10} , $\text{PM}_{2.5}$ and PM_1 mass concentrations are not enough to take steps on controlling fine particulate pollution because of lack of long-term and continuous observations for the PM_{10} , $\text{PM}_{2.5}$ and PM_1 in such a rapidly developing region.

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