



Weak mixing in Lake Kivu: New insights indicate increasing risk of uncontrolled gas eruption

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[1] The deep waters of the East African Rift Lake Kivu contain large amounts of dissolved carbon dioxide and methane. The release of a fraction of these gases, which could be triggered by a magma eruption within the lake, would have catastrophic consequences for the two million people living on its shore. Up to now the safety assessment of the lake was based on the assumption that the gas concentrations in the deep waters are in a steady state with a residence time of 400 years. Turbulent transport was regarded as the main pathway of vertical exchange. Recent measurements and the analysis of the vertical transport processes in the lake radically change this evaluation. The vertical turbulent exchange is negligible, as documented by a spectacular set of several hundred double-diffusive layers. Gases are mainly transported out of the deep zones by a slow upwelling with a residence time of 800–1000 years. Our results indicate that the methane production within the sediment has recently increased, leading to a gas accumulation in the deep waters and consequently decreasing the heat input needed to trigger a devastating gas release. With the estimated current CH₄ production, the gas concentrations could approach saturation within this century.

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1. Introduction

[2] Lake Kivu is an East African rift lake between Rwanda and the Democratic Republic of Congo (Figure 1) with maximum depth of 485 m and a volume of 550 km³. Seasonal mixing reaches only a depth of 50–80 m. Below, about 300 km³ STP (gas volume at 0°C and 1 atm) carbon dioxide (CO₂) and 60 km³ STP methane (CH₄) are dissolved in the permanently stratified

deep waters. CH₄ concentrations are approximately 21% of CO₂, but because of the low solubility of CH₄, its contribution to the total gas pressure exceeds that of CO₂ [Tietze, 1978; Schmid *et al.*, 2004b]. Consequently, a gas release from Lake Kivu would be triggered by supersaturation of CH₄, while CO₂ would be stripped into the gas bubbles and contribute most to the released gas volume. Although the lake is currently in a relatively safe state, a gas release

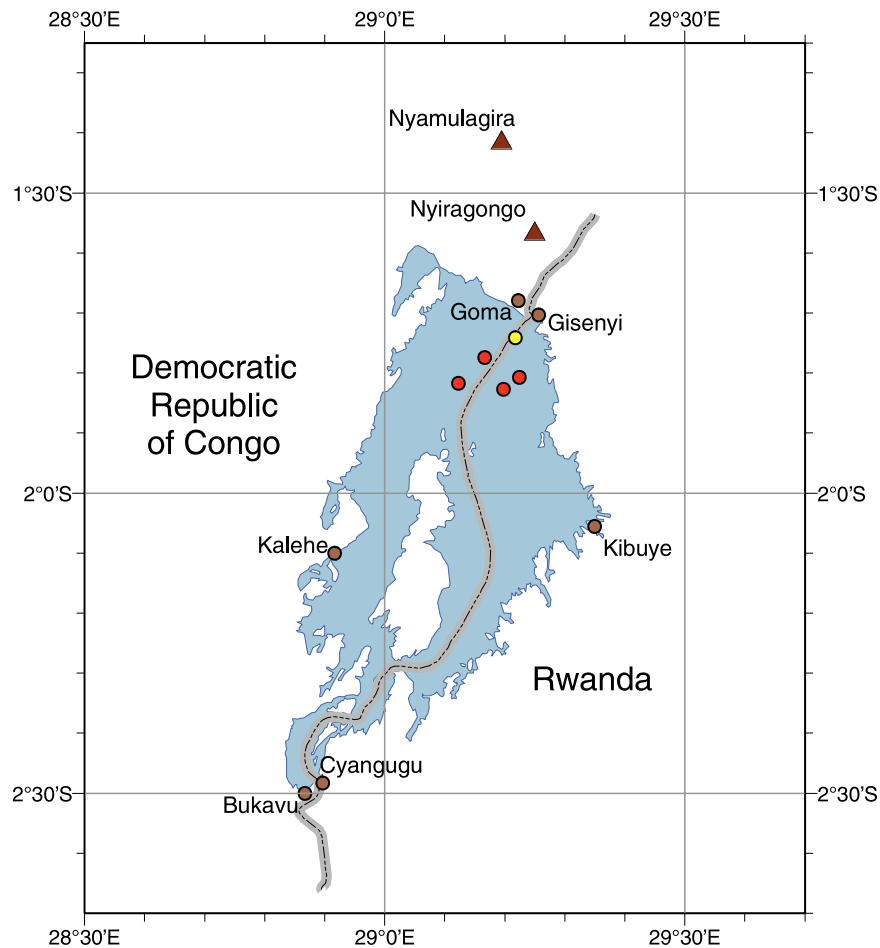


Figure 1. Map of Lake Kivu. The red circles show the locations of the four CTD profiles, and the yellow circle marks the location where water samples for chemical analysis were taken.

could be initiated by a sufficiently large volcanic eruption within the deep water [Schmid *et al.*, 2004b]. On the basis of the sediment stratigraphy it has been suggested that a gas release had indeed happened before [Haberyan and Hecky, 1987]. Two active volcanoes are situated near the northern shore, Nyiragongo which erupted in January 1977 [Durieux, 2004] and January 2002 [Komorowski *et al.*, 2004] and Nyamulagira which erupts every few years [Tedesco, 2004]. New fractures toward the lake opened during the last eruption of Nyiragongo, and the tectonic activity has been unusually high during the past few years [Komorowski *et al.*, 2004]. Due to its large size and the high population density, a gas release from Lake Kivu could cause an unimaginable disaster, far beyond the events at the “Killer Lakes” Monoun in 1984 [Sigurdsson *et al.*, 1987] and Nyos in 1986 [Kling *et al.*, 1987; Sigvaldason, 1989] which claimed 37 and more

than 1700 lives, respectively. These lakes are now artificially degassed [Halbwachs *et al.*, 2004; Krajick, 2003].

[3] Bacteria produce CH_4 in the lake sediment both by decomposing settled organic material and using magmatic CO_2 and hydrogen [Deuser *et al.*, 1973; Schoell *et al.*, 1988]. In the surface layer CH_4 is oxidized using oxygen and sulfate as electron acceptors. Two transport mechanisms bring the CH_4 to the surface: (1) a deep water input leads to a slow upwelling over the whole lake area which transports CH_4 upward, and (2) turbulent mixing causes a down-gradient transport of CH_4 . It has previously been assumed that the gases are mainly transported to the surface by turbulent diffusion with a residence time of 400 years in the deep water [Tietze, 1978]. In the following we will revise this picture of the vertical transport processes by first quantifying

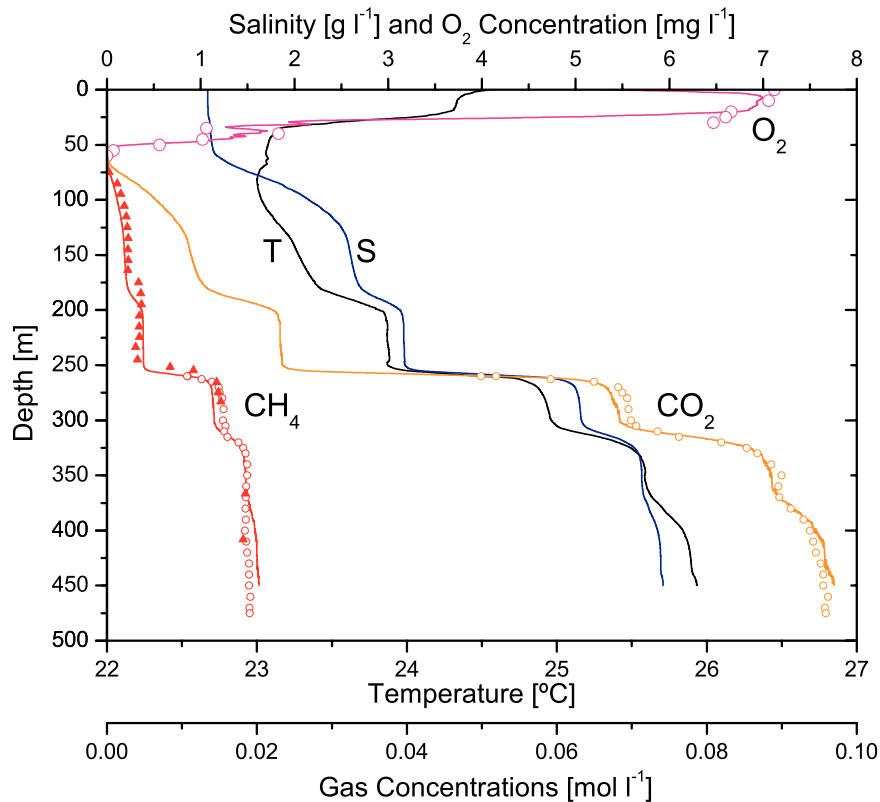


Figure 2. Vertical profiles of temperature (T), salinity (S), and dissolved gas concentrations in Lake Kivu in February 2004. The orange line is the CO_2 concentration calculated from pH and alkalinity, the red line is the CH_4 concentration assuming a constant CH_4/CO_2 ratio, the open circles show the gas concentrations measured in November 2003, and the triangles show the CH_4 concentrations measured with the Capsum Mets sensor. The oxygen concentrations measured with the CTD probe and using the Winkler method are shown by the magenta line and the magenta circles, respectively.

the turbulent exchange and then the upwelling transport of CH_4 .

2. Measurement Methods

2.1. CTD Profiles

[4] Thirteen vertical profiles of temperature and conductivity (CTD) were measured in the northern half of Lake Kivu with a Sea-Bird SBE-19 which was equipped with a Sea-Bird SBE-22B combined pH and oxygen sensor. With the exception of a signature at 250 m depth (see below), temperature and salinity in the deep water were horizontally practically homogeneous. The four deepest reaching profiles were averaged to produce the temperature and salinity profiles in Figure 2.

2.2. Temperature Microstructure Profiles

[5] Ten vertical profiles of temperature microstructure were measured with a free-falling Sea-Bird

SBE 11 equipped with two Thermometrics FP07 microstructure thermistors with nominal response times of 7 ms and a pressure sensor to determine the sinking rate. The profiler sinking rate of $\sim 0.1 \text{ m s}^{-1}$ and the sampling frequency of 96 Hz resulted in a vertical resolution of $\sim 1 \text{ mm}$. The temperature resolution (a few $10^{-5} \text{ }^\circ\text{C}$) of the microstructure probe was sufficient to resolve the temperature variance within the double-diffusive mixed layers (Figure 3).

2.3. Water Chemistry

[6] Water samples (Table 2) were analyzed with standard methods for alkalinity, nutrient contents and main ionic composition. Salinity was calculated from conductivity and ionic composition [Wüest *et al.*, 1996] and agreed within 2% with the values calculated from measured ion concentrations. Historic conductivity and salinity values are inconsistent: The levels measured by Tietze [1978] in the deep water were about 10% higher than our values,

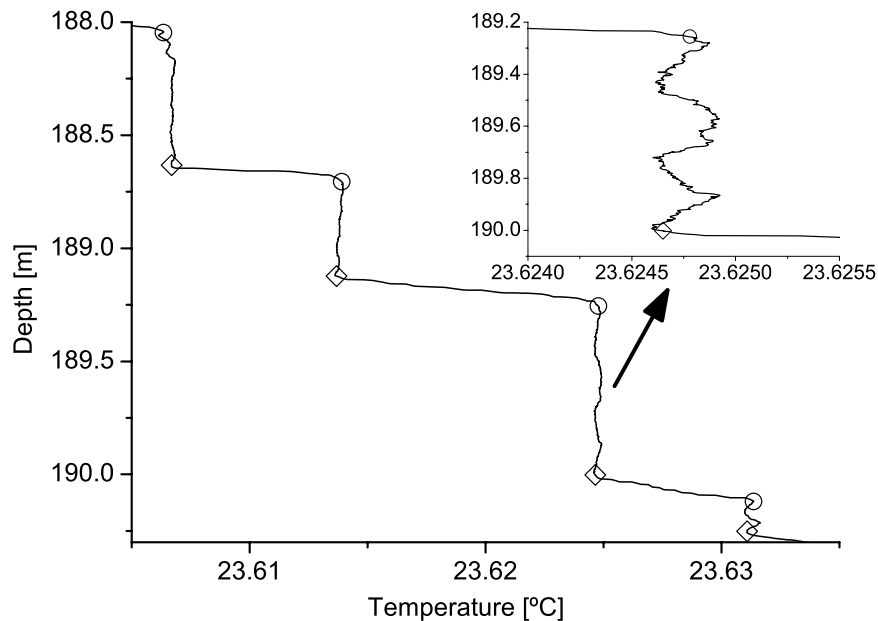


Figure 3. An example of double-diffusive layering observed in Lake Kivu at 188 to 190 m depth. The microstructure profiles were split into well-mixed layers and interfaces as depicted by the open circles (upper boundaries of mixed layers) and diamonds (lower boundaries of mixed layers). The sharpness of the transition between layers evidences the current activity of the double-diffusion. The inset shows the observed temperature variability within one of the mixed layers.

whereas those of *Degens et al.* [1973] were lower by almost a factor of two. However, recalculating conductivity from the main ionic composition given by *Degens et al.* [1973] yields values comparable to ours. CO_2 was calculated from pH and alkalinity, including the effects of temperature and salinity on the first dissociation constant of H_2CO_3 [Cai and Wang, 1998]. The uncertainty in this calculation is mainly due to the pH, since an error in pH of ± 0.05 produces an error in the CO_2 concentrations of approximately $\pm 10\%$. CH_4 was measured with a Capsum Mets methane sensor, which was calibrated for the high concentrations in Lake Kivu. The standard deviation observed by the manufacturer in the calibration was 0.4 mmol L^{-1} . In the deep water of Lake Kivu this corresponds to an error of 2–5%. The sensor recorded the CH_4 concentration every 0.5 second. In the field, it exhibited an unexpectedly slow response time of 46 min. The measured concentrations were extrapolated by fitting time series of 200 to 600 consecutive samples to an exponential curve. The error due to the fitting procedure was 0.5% below the chemocline at 270 m depth, 2–3% within the chemocline and up to 10% above 230 m depth. It was larger at lower concentrations due to the higher relative noise of the instrument at low voltages.

[7] In November 2003, CH_4 and CO_2 concentrations had been measured by transferring water

through a polyethylene tube to the lake surface, separating the gas and water phases, measuring their flow rates, and determining the CH_4 and CO_2 concentrations in the gas phase with a portable GA2000 infrared gas analyzer. The gas analyzer was tested by measuring gas samples from the lake with another identical instrument and before and after the expedition with a standard of 40% CH_4 and 60% CO_2 . Errors were in the range of $\pm 0.2\%$ volume, i.e., less than 2% of CH_4 concentrations observed in the lake. The measurement error for the Schlumberger gas flowmeter is about $\pm 2\%$, while the error of the water flow can be neglected. Consequently, a total error of $\pm 4\%$ is assumed. Measured ratios of CH_4 to CO_2 concentrations ranged between 0.196 and 0.224. For the analysis of the double-diffusive transport, where the effect of CH_4 is negligible compared to salinity and CO_2 , CH_4 was assumed to be a constant fraction of CO_2 (Figure 2).

2.4. Temperature Time Series

[8] Four Vemco minilog temperature loggers which recorded temperature every 15 min with a resolution of 0.01°C were installed in November 2003 at depths of 15, 185, 260 and 315 m and retrieved in February 2004. From this time series and the background temperature gradient, vertical displacements due to internal waves and conse-

quently their energy content was determined with the same method as for Lake Nyos [Schmid *et al.*, 2004a].

3. Results and Discussion of Measurements

[9] Average vertical profiles of temperature and salinity, as well as CO₂ and CH₄ concentrations in February 2004 are shown in Figure 2. Compared to previous measurements in the 1970s [Tietze, 1978], the measured CH₄ concentrations in the deep water have increased by 15–20% and CO₂ concentrations by about 10%. The increase in the CH₄ concentrations is clearly significant if we exclude a systematic error of much more than 5% in the measurements of Tietze, while the observed 10% change in CO₂ indicates an augmentation but is hardly sufficient to refute the null hypothesis of unaltered concentrations.

[10] Below the zone reached by seasonal mixing, temperature, salinity, CO₂ and CH₄ all increase with depth. Salinity and CO₂ are stabilizing the water column, whereas temperature and CH₄ are destabilizing (dissolved CH₄ decreases the water density). Since the molecular thermal diffusivity is about 2 orders of magnitude larger than that of dissolved substances, the destabilizing effect of temperature can lead to the formation of mixed layers by double-diffusive convection [Turner, 1973]. In fact, our temperature microstructure profiles show 250–350 well-mixed layers between 120 and 460 m depth. Most of these layers are between 0.2 and 1.0 m thick, with interfaces as thin as 0.05–0.4 m and temperature steps between 0.001 and 0.005°C. A typical example of four successive layers is shown in Figure 3. Generally the interfaces were free of temperature inversions, indicating complete absence of any vertical turbulence. Consequently, below 120 m depth, double-diffusive convection must be dominating over the background turbulent diffusivity and we can neglect the latter in our analysis.

[11] The intensity of double-diffusive convection is characterized by the density ratio $R_\rho = \beta(\partial S/\partial z)/\alpha(\partial T/\partial z)$, the ratio between the stabilizing density effect of the salinity (S) gradient and the destabilizing effect of the temperature (T) gradient [Turner, 1973]. Here, α is the thermal expansivity and β the coefficient of haline contraction. In the case of Lake Kivu, gas concentration gradients multiplied by their contraction coefficients have

to be added to the numerator [Schmid *et al.*, 2004a]. Below 150 m depth, R_ρ ranges between 2 and 4. The ratio of the apparent diffusivity of dissolved substances to that of heat decreases with increasing R_ρ [Kelley, 1990; Turner, 1965].

[12] The heat fluxes through the double-diffusive steps were calculated with three different methods: (1) from the interface temperature gradients between convective layers assuming molecular diffusion with a diffusion coefficient of $1.4 \cdot 10^{-7} \text{ m}^2 \text{ s}^{-1}$, (2) using the semiempirical flux law for double-diffusion by Kelley [1990], which successfully predicted heat fluxes in the comparable case of Lake Nyos [Schmid *et al.*, 2004a], and (3) from observed temperature fluctuations within convectively mixed layers by applying the convective scaling relation and assuming that dissipation within a layer is equal to the buoyancy flux [Matzner, 2001].

[13] Figure 4 shows the results of methods 1 and 2. Method 3 yielded similar results as method 2 but with a larger scatter. Method 1 was expected to underestimate the heat fluxes since there is some additional heat transport through the interfaces besides molecular diffusion due to intrusions of rising and sinking double-diffusive convective plumes. Except for the regions near the high gradient zones, and for the zone between 180 and 200 m depth, the heat fluxes determined by method 2 are only on the order of 0.01–0.04 W m⁻², which is below typical heat fluxes observed in East Africa [Pollack *et al.*, 1993] but within the large range of 0.016–0.18 W m⁻² measured in the sediment of Lake Kivu [Degens *et al.*, 1973]. The corresponding CH₄ flux through the chemocline at 260 m depth, calculated with the flux laws of Kelley [1990] as described by Schmid *et al.* [2004a], is only on the order of 1 g CH₄-C m⁻² yr⁻¹, indicating a residence time of CH₄ in the deep water of ~20,000 years. It will be shown below that the true residence time is much smaller due to the upwelling transport. The double-diffusive fluxes are enhanced near the gradient zones (Figure 4) because the high gradients are sustained by lateral water inputs (see below) while double-diffusive convection tends to smooth out the large-scale gradients.

[14] A previous analysis of double-diffusive steps in Lake Kivu [Newman, 1976] yielded much higher heat fluxes of 0.7–1.6 W m⁻². These calculations were biased by an underestimation of the salinity gradient of at least a factor of 2, overestimating the

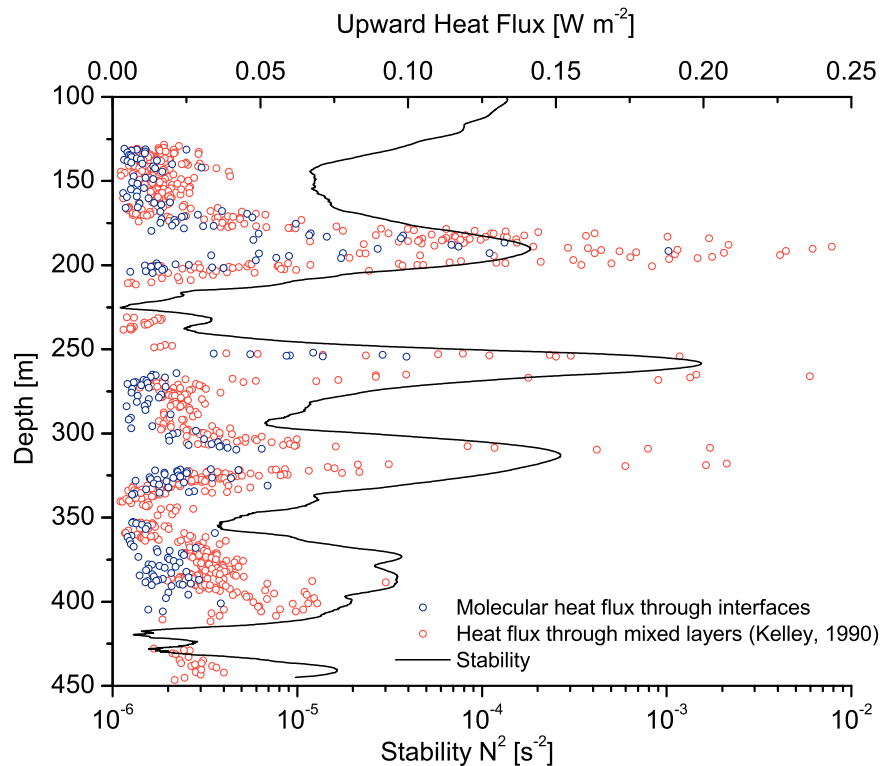


Figure 4. Vertical upward heat fluxes calculated from the structure of the mixed layers using the semiempirical flux law of *Kelley* [1990] (red dots) and by assuming molecular diffusion within the interfaces (blue circles), and vertical profile of the stability N^2 .

heat fluxes by a factor of 4. Nevertheless, the mixed layers observed by *Newman* [1976] were larger with an average thickness of 1.0 m and 0.02°C temperature steps, values that were reached by only 20% (layer thickness) and 3% (temperature steps) of our observations. This indicates that double-diffusive convection was more intense in the 1970s. Since then, temperature has increased in the top 200 m of the lake by about $0.3\text{--}0.5^\circ\text{C}$, most probably due to global warming [*Lorke et al.*, 2004], weakening the average temperature gradient by about 15%. CO_2 concentrations seem to have increased by about 10% and CH_4 concentrations by about 15–20% in the deep waters. In total, R_p has increased by about 20–30% within 30 years, weakening the upward transport of salts and gases.

4. Model Description

[15] A simple one-dimensional diffusive-advective model for salinity and CH_4 was set up with the lake module of the software AQUASIM 2.1 [*Reichert*, 1994]. Simulations were performed with a 1-m vertical grid and with variable time

stepping chosen by the integration algorithm. Shortly, the following knowledge and assumptions about the system were implemented:

[16] There must be at least one source in the deep waters of the lake which compensates for the salt losses by the Ruzizi outflow [*Degens et al.*, 1973]. The water input by this source displaces the ambient water causing a slow upwelling transport. The strong temperature gradient at 260 m depth with an increase of 0.7°C within 6 m, which was observed by all measurements since *Damas* [1937], needs to be sustained by an active process. Otherwise it would be destroyed by molecular heat diffusion within a few years. A distinct negative temperature signal (and a weaker signal in conductivity) was observed at 250 m depth which was strongest in the profiles measured in the North-Eastern part of the lake and can only be explained by an input of cooler and less saline water at this depth. This input is most probably the process sustaining the strong temperature gradient below. Furthermore, the observed decrease in CH_4 must be due to dilution with water containing little CH_4 since oxidation can be excluded in the

anoxic and sulphate-depleted deep water. No clear signals from water inputs were observed above the other strong gradients, but since there is no other explanation for these gradients, we assume that they are sustained by similar water inputs free of dissolved CH₄. The flow rates, temperatures and salt concentrations of these inputs as well as the geothermal heat flux were fitted to reproduce the observed salinity and temperature profiles.

[17] Below 120 m, turbulent vertical diffusivity was set to $0.15 \cdot \epsilon/N^2$ [Osborn, 1980] for temperature and, based on our analysis of the fluxes through the double-diffusive layers, a factor 10 lower for salt and gases. Here, N^2 is the stability of the density stratification. The energy dissipation ϵ below 120 m depth is on the order of $1.0 \cdot 10^{-10} \text{ W kg}^{-1}$, which was confirmed by three independent methods: by analyzing Batchelor spectra of temperature microstructure measurements as described by Kocsis *et al.* [1999], by applying scaling relations to the observed double-diffusive staircase, and by estimating the energy content of vertical displacements due to internal waves observed with temperature loggers. No double-diffusive layers were observed above 120 m depth, and the temperature minimum at 85 m indicates occasional convective mixing down to this depth. The top 50 m of the lake are mixed annually. Consequently, diffusivity was steadily increased from 120 m depth to $10^{-5} \text{ m}^2 \text{ s}^{-1}$ at 50 m depth and kept constant above.

[18] A fast process removing CH₄ by oxidation above 60 m depth was introduced into the model. CH₄ production was assumed to be homogeneous at the sediment surface; i.e., the CH₄ input is proportional to the sediment surface area at each depth. It was assumed that the historic value of CH₄ production had been constant for a sufficiently long time to approach equilibrium (in our model we arbitrarily used 930 years, but the exact value has no significant influence on the results), that it started to increase linearly with time with the growing population 30 years before the 1974 measurements and increased even more between 1974 and 2004. The three values of CH₄ production (historic, 1974 and 2004) were determined by fitting to the CH₄ profiles observed in 1974 [Tietze, 1978] and 2004. Of course the chosen temporal development is one of several reasonable possibilities, but using another time frame would not alter the average CH₄ production needed to

produce the observed increase in concentrations between 1974 and 2004.

5. Results and Discussion of Simulations

[19] Figure 5 compares simulated and observed temperatures, salt and CH₄ concentrations. All observations were reproduced with high accuracy except for surface temperatures, which are influenced by seasonal surface heat fluxes not included in the model. Table 1 shows the water sources and their properties as used in the model to optimally reproduce the observations. The observed large-scale temperature profile could be best predicted with a geothermal heat flux of 0.02 W m^{-2} which agrees well with the independent estimates of small-scale double-diffusive heat fluxes (Figure 3).

[20] The total simulated water input below the surface layer is $1.3 \text{ km}^3 \text{ yr}^{-1}$, which is 40% of the Ruzizi outflow [Degens *et al.*, 1973]. The absolute flow rates of the sources given in Table 1 cannot be exactly determined, but to produce the observed steps, it is absolutely necessary that the inputs at 180 and 250 m depth are about one order of magnitude larger than those below. The fact that both deuterium and ¹⁸O isotopes lie on straight mixing lines with conductivity [Tietze, 1978], indicates that the water within the lake is a blend of only one groundwater source and lake surface water, implying that also the mid-depth source water consists of the same two end-members. This could be caused by either lake surface water infiltrating into the groundwater system, or by groundwater entering the lake above 100 m depth and being diluted with lake water while sinking down to its intrusion depth just above a high density gradient. In the second case, the sources should already contain some CH₄ from the lake water which would require slight changes in the properties of the mid-depth sources.

[21] The calculated residence time for the water below 260 m depth is ~ 870 years, with an upward CH₄ transport on the order of $\sim 30 \text{ g C m}^{-2} \text{ yr}^{-1}$. From this analysis we conclude that the water body below the chemocline at 260 m depth is more strongly decoupled from the upper part of the lake than previous budget calculations assumed. This means that there is more time available for CH₄ to accumulate in the deep water.

[22] The CH₄ production needed to fit the observed profiles was a constant historic production of 32 g

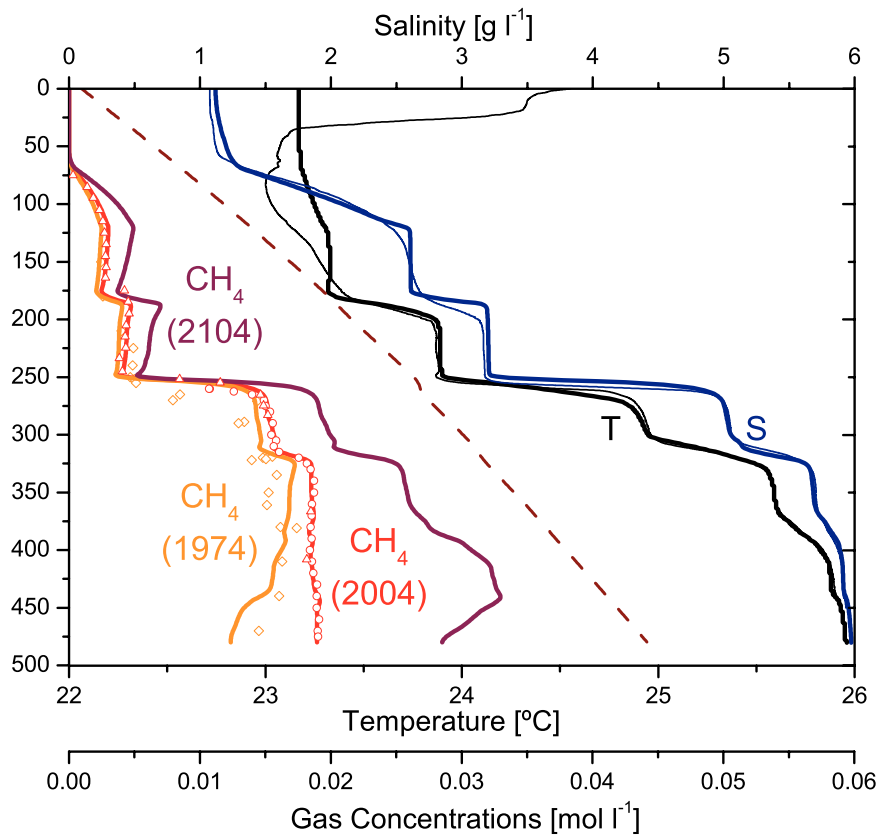


Figure 5. Simulated (thick lines) CH_4 concentrations, temperatures (T), and salinities (S) compared to measurements (thin lines and symbols), and simulated CH_4 concentration in 100 years compared to 80% CH_4 saturation (dashed line).

$\text{CH}_4\text{-C m}^{-2} \text{ yr}^{-1}$, rising to values of 70 in 1974 and 120 in 2004. The development of CH_4 production could have been different, but an average of about $100 \text{ g CH}_4\text{-C m}^{-2} \text{ yr}^{-1}$ in the past 30 years is needed to reproduce the observed CH_4 concentrations. Even if we conservatively assume a growth in CH_4 concentrations of only 10%, the recent production would still be at least twice the historic value. The historic production agrees with the oxidation rate of $30 \text{ g CH}_4\text{-C m}^{-2} \text{ yr}^{-1}$ at the oxic/anoxic interface at 35 to 50 m depth observed in 1971 [Jannasch, 1975]. Because of the high residence time of CH_4 , the recent production would

not yet have had a large influence on this oxidation rate. Most probably the present structure took longer to evolve than the 990 years simulated, as weaker stratification would have facilitated turbulent upward transport in earlier times. Altogether the simulated time frame is consistent with the theory based on observations in the lake sediment stratigraphy that the present stable stratification was set up about 2000 years ago [Stoffers and Hecky, 1978; Haberyan and Hecky, 1987].

[23] Primary production within Lake Kivu was estimated at $375 \text{ g C m}^{-2} \text{ yr}^{-1}$ in 1971 [Degens

Table 1. Input Sources Used in the One-Dimensional Diffusive-Advective Model of Lake Kivu

Depth, m	Depth Range, m	Flow, $\text{m}^3 \text{ s}^{-1}$ ($\text{km}^3 \text{ yr}^{-1}$)	Salinity, g l^{-1}	Temperature, $^{\circ}\text{C}$
180	10	22 (0.69)	2.1	22.7
250	5	15 (0.47)	2.7	23.3
315	10	1 (0.03)	3.4	25.2
365	50	1.5 (0.05)	5.5	24.5
425	60	0.8 (0.03)	5.8	25.3
465	30	1.25 (0.04)	6.0	26.0

Table 2. Alkalinity, pH, and Concentrations of Calcium, Sodium, Chloride, Silica, Phosphate, Ammonium, Hydrogen Sulfide, and Sulfate Measured in Water Samples From Lake Kivu in February 2004^a

Depth, m	Alk, mM	pH	Ca ²⁺ , mM	Na ⁺ , mM	Cl ⁻ , mM	Si, μM	PO ₄ ³⁻ , μM	NH ₄ ⁺ , μM	H ₂ S, μM	SO ₄ ²⁻ , μM
0	12.9	9.11	0.10	4.30	0.51		0.37	0	0	120
[0]			[1.19]	[5.28]		[231]	[0.8]	[18]		
10	12.8	9.11	0.10	4.27	0.50		0.39	0	0	128
20	12.8	9.11	0.10	4.33	0.49		0.36	0	0	130
30	12.8	8.97	0.10	4.30	0.50	133	0.39	0	0	128
40	12.9	8.92	0.11	4.36	0.50	134	0.34	0	0	123
50	13.0	8.86	0.13	4.41	0.54	138	0.37	0	0	127
60	13.4	8.68	0.17	4.48	0.54	166	0.37	4	4	125
70	15.5	7.74	0.37	4.86	0.66	229	4.74	126	101	83
80	19.1	7.19	0.91	6.10	0.72	342	13.28	281	233	46
90	23.1	7.00	1.27	7.19	0.82	481	23.97		275	26
100	25.9	6.90	1.57	7.56	0.90	564	29.45	578	292	23
[100]			[1.60]	[8.35]		[424]	[18.8]	[487]		
120	29.0	6.79	1.94	8.50	0.98	590	36.76	708	305	22
140	30.4	6.74	1.98	8.46	1.02	630	41.26	810	356	18
160	31.0	6.71	2.07	8.75	1.02	645	42.10	866		
180	31.5	6.67	1.99	8.64	1.06	657	42.52	1101		
200	37.2	6.49	2.27	9.93	1.33	810	60.80	1336		
[200]			[2.07]	[10.64]		[428]	[32.7]	[1314]		
220	37.6	6.47	2.47	10.84	1.36	829	65.72	1373		
240	37.6	6.47	2.48	10.77	1.40	825	65.29	1367		
260	54.6	6.24	2.99	14.33	1.91	1196	128.44	2722		
280	60.5	6.16	3.35	17.00	1.96	1283	156.56	3361		
300	61.0	6.16	3.31	16.78	2.05	1286	158.67	3514		
320	66.9	6.10	3.46	18.89	2.07	1371	177.64	3696		
340	69.2	6.08	3.47	19.34	2.11	1413	181.86	4001		
[350]			[2.77]	[20.24]		[1056]	[53.2]	[5460]		
380	69.5	6.07	3.36	18.91	2.24	1426	182.91	3909		
420	71.5	6.06	3.64	20.60	2.38	1613	183.91	3955		
[450]			[2.81]	[21.20]		[1226]	[54.8]	[7105]		

^a Alk, alkalinity. The values from *Degens et al.* [1973] are given in brackets for comparison.

et al., 1973]. The riparian population has at least doubled since then, and nutrient input to the lake and primary production should have risen significantly. A comparison of our chemical data with those of *Degens et al.* [1973] confirm the increased biological activity: Concentrations of silica (by 30–35%), calcium (by 25–30%) and phosphate (by more than a factor 3) all strongly increased in the deep waters and decreased in the surface waters (Table 2), indicating a higher export of organic material from the surface layer to the deep waters and a corresponding precipitation of calcite in the surface layer and its redissolution in the deep waters. It should be noted that the deep water ammonium (NH₄⁺) concentrations given by *Degens et al.* [1973] disagree with this interpretation, as they were almost a factor of two larger than ours. However, their NH₄⁺ data did not match the mixing lines shown by all other constituents, whereas we observed a perfect correlation with phosphate with a N:P ratio of 21.5.

[24] If we conservatively estimate the average primary production from 1974 to 2004 at 400 g C m⁻² yr⁻¹, 25% of the produced C would have to be converted to CH₄ at the sediment surface, assuming that all CH₄ is biogenic [*Schoell et al.*, 1988]. Another more speculative hypothesis for the increased export of organic matter to the deep waters is the introduction of the sardine *Limnothrissa miodon* to Lake Kivu in 1959. In Lake Kariba, the introduction of this species led to a drastic reduction of the zooplankton biomass and an increase in phytoplankton [*Marshall*, 1991]. Similar effects were observed in Lake Kivu [*Dumont*, 1986], even though in the meantime a new equilibrium seems to have been established [*Isumbisho et al.*, 2004].

[25] The conclusion that diffusive transport is very weak and upward transport is dominated by upwelling with a deep water residence time of 800–1000 years has several important consequences.

We do not know whether the deep water input will be sustained constantly in the future. If this water input weakens, the accumulation of CH₄ in the deep water will intensify. The residence time of dissolved gases in the deep water is longer than previously assumed, which implies that a larger part of an increasing CH₄ production accumulates in the deep water and dangerous concentrations could be reached more quickly. Since CH₄ is accumulating in the water during the upwelling, in a steady state condition the maximal concentrations would develop not in the deepest zones but just below the chemocline at 260 m depth where a lower concentration is needed to reach saturation.

[26] At the estimated current CH₄ production, the gas concentrations in the deep waters of the lake would approach saturation within this century. Presently, CH₄ causes ~80% of the gas pressure. The 80% saturation line is shown for comparison in Figure 5. The probability of a gas release triggered by a magmatic eruption within the lake increases much faster than the CH₄ concentration due to the nonlinear relation between gas concentrations and the heat input needed to produce a rising plume that reaches saturation depth [Schmid *et al.*, 2004b]. Once the total gas pressure approaches the hydrostatic pressure, a magmatic eruption is no longer required, and a gas release might even be triggered by a strong internal wave. Considering the two million people at risk and the strong recent tectonic activity [Komorowski *et al.*, 2004], we suggest that the gas concentrations in the lake should be critically monitored and possibly reduced by gas stripping techniques in the next decades.

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